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THE EFFECTS OF DIESEL EXHAUST POLLUTION ON FLORAL VOLATILES AND THE CONSEQUENCES FOR HONEY BEE OLFACTION

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Abstract – There is growing evidence of a substantial decline in pollinators within Europe 1 and North America, most likely caused by multiple factors such as diseases, poor nutrition, 2 habitat loss, insecticides and environmental pollution. Diesel exhaust could be a contributing 3 4 factor to this decline, since we found that diesel exhaust rapidly degrades floral volatiles, which honey bees require for flower recognition. In this study we exposed eight of the most 5 common floral volatiles to diesel exhaust in order to investigate, whether it can affect volatile 6 7 mediated plant-pollinator interaction. Exposure to diesel exhaust altered the blend of common 8 flower volatiles significantly: myrcene was considerably reduced, β -ocimene became undetectable, and β -caryophyllene was transformed into its *cis*-isomer isocaryophyllene. 9 Proboscis extension response (PER) assays showed that the alterations of the blend reduced 10 the ability of honey bees to recognize it. The chemically reactive nitrogen oxides fraction of 11 diesel exhaust gas was identified as capable of causing degradation of floral volatiles. 12 13

Key Words - Floral scent compounds, diesel exhaust, nitrogen oxides, scent degradation,
scent recognition, proboscis extension response.

INTRODUCTION

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Since the winter and spring of 2006-7 there have been reports of elevated colony 18 losses of managed European honey bees (Apis mellifera) in the U.S.A., Europe, the Middle 19 East, and Japan (Abrol 2012; Oldroyd 2007). Furthermore, there is growing evidence of 20 substantial losses of many wild pollinator species worldwide, mainly within Europe and North 21 22 America (Biesmeijer et al. 2006; Potts et al. 2010). Declines of managed honey bees and wild pollinators may have serious implications, particularly because global food security is 23 considered to be dependent on animal pollination (Abrol 2012). Even though most staple food 24 25 crops do not require insect pollination (Ghazoul 2005), 35% of the world crop production for human food depends on pollinators (Klein et al. 2007). Those pollinated crops are particularly 26 important for our food diversity and add nutritional value to our diet (Steffan-Dewenter et al. 27 28 2005). Some crops, such as oilseed rape (Brassica napus), are considered self-fertile (Free 1993), but insect pollination can strongly contribute to increased yield and market value 29 (Bommarco et al. 2012). About 73% of cultivated crop varieties are pollinated by some type 30 of bee, with the European honey bee dominating crop pollination worldwide (Abrol 2012). 31 The drivers behind the global pollinator decline are likely to be multifactorial and include 32 33 fragmentation and loss of habitat, increased pesticide use, decreased resource diversity, alien species, spread of pathogens, and climate change (Epstein et al. 2013; Kerr et al. 2015; 34 Oldroyd 2007; Potts et al. 2010, Vanbergen et al. 2013). However, it is unlikely that this is a 35 36 definitive list, and it is, therefore, important to investigate additional potential stressors that could results in negative effects on bee fitness. 37

Bees use mixtures of olfactory and visual stimuli to find suitable host plants (Dötterl and Vereecken 2010). They rely mainly on olfactory cues during their initial foraging bouts, and visual cues become more important in host-plant location as bees gain more experience (Dobson 1994). However, floral scent remains an important stimulus for experienced bees,

since it allows them to discriminate between rewarding and non-rewarding flowers (Dobson
2006; Wright and Schiestel 2009).

Ozone or reactive species, such as hydroxyl and nitroxyl or nitrate radicals, readily
react with volatile organic compounds (VOC) (Atkinson and Arey 2003; Calogirou et al.
1999), which contribute to floral scent and could, therefore, impact upon VOC-mediated
plant-insect interactions. McFrederick et al. (2008) modelled the dispersion of three common
floral scent compounds (linalool, myrcene, and β-ocimene) under different air pollution
scenarios and concluded that increasing pollution levels may impair the recognition of floral
scents by pollinators.

51 In an earlier study we were able to show that diesel exhaust alters the VOC composition of a synthetic floral odour blend (Girling et al. 2013) designed to mimic the 52 oilseed rape (OSR) cultivar Brassica napus, cv. Topas. It consisted of eight compounds, 53 54 which are behaviourally active in honey bees, mixed in their naturally occurring ratios (Blight et al. 1997). The exposure of the OSR blend to diesel exhaust rendered two compounds 55 undetectable, which significantly reduced the ability of honey bees to recognize the altered 56 floral scent (Girling et al. 2013). Whether such changes deleteriously affect honey bee 57 foraging or foraging by other pollinators that utilise these compounds is currently unknown. 58 59 Three of the volatile compounds in OSR are floral scent compounds that occur in more than 50% of seed plants (reviewed by Knudsen et al. 2006). 60

In order to investigate further the effects of diesel exhaust emissions on floral VOCs and to study whether such effects could be a more widespread phenomenon, we exposed eight of the 12 most common floral scent compounds (Knudsen et al. 2006) to diesel exhaust, none of which are found in OSR, mixed in equal amounts. To investigate if bees that were conditioned to our floral volatile blend were able to still recognize it after they underwent the alterations caused by diesel exhaust, we used the classical conditioning of the proboscis extension reflex (PER; Bitterman et al. 1983; Kuwabara 1957). Additionally, we exposed all

68	eight common floral scent compounds individually to diesel exhaust with the intention of
69	determining their possible fate in a diesel exhaust polluted environment. In order to confirm
70	the general mechanism behind the alteration of floral scent components in such an
71	environment, we also exposed all compounds to various concentrations of NO_X , since that
72	was the observed mechanism behind the floral scent degradation in our earlier study (Girling
73	et al 2013).
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76	METHODS AND MATERIALS
77	
78	Floral Volatile Blend. Based on a review by Knudsen et al. (2006) on the diversity and
79	distribution of floral scent, we selected eight common floral compounds that occur in more
80	than half of all the families of seed plants: β -Pinene, myrcene, limonene, β -ocimene,
81	benzaldehyde, β -caryophyllene, methyl salicylate, and benzyl alcohol (Table 1). We mixed all
82	eight compounds in equal amounts to produce a common floral volatile (CFV) blend.
83	
84	Floral Volatile Exposures. For the exposures the general procedure was to place 1μ L of CFV
85	blend on a filter paper (2.1cm circle, grade 3MM; Whatman plc, Maidstone, UK) and a 4.5cm
86	stir bar in a 1000mL glass bottle (VWR International Ltd., West Sussex, UK). To assess the
87	influence of UV light the tests were replicated in both clear and amber glass bottles. The CFV
88	blend was exposed to either ambient air or diesel exhaust at room temperature. The ambient
89	air treatment consisted of a bottle of air collected from the laboratory. Diesel exhaust was
90	collected into the bottle from the exhaust pipe of a Suntom SDE 6500 E diesel generator
91	(Fuzhou Suntom Power Machinery Co., Ltd. Fuzhou, China) for 3min at a flow rate of
92	1L/min. Bottles were sealed with two layers of Parafilm M [®] (Bemis Flexible Packaging,
93	Oshkosh, WI, USA) and a lid. The lid contained a 1mm bore hole in its centre to allow

insertion of the solid-phase microextraction (SPME) fibre holder for the collection of 94 95 volatiles. The bottle content was mixed with a stir bar at 300rpm. The content of the bottle was sampled after 1min, 30min, 1h and 2h. At each sample time point the SPME fibre was 96 extended into the bottle and exposed to the volatiles for 5min. The SPME fibre was then 97 inserted into the injector of the gas chromatograph linked to the mass spectrometer (GC/MS) 98 and allowed to desorb for 10min. This 10min desorption period was sufficient to clean the 99 100 SPME fibre so that the same fibre could be used to collect VOCs from the same bottle and filter paper at a later time point. After the collection of VOC at all four time points the bottle 101 was cleaned with hexane, dried under nitrogen gas, and left in the fume hood overnight to 102 103 remove the hexane residue. Each of the air treatment and bottle type combinations were repeated five times at room temperature for each of the two floral blends. 104

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106 Floral Blend GC/MS Analyses. After each VOC adsorption period the SPME fibre assembly (65µm PDMS/DVB, fused silica, 24Ga, manual holder; Supelco, Bellefonte, PA, USA) was 107 108 injected into a Hewlett Packard HP6890/5972A GC/MS at an injection port temperature of 109 250°C and a split ratio of 10:1. The GC/MS was equipped with a SPME injection sleeve (I.D. 0.75mm, Supelco) and HP-Innowax column (I.D. 0.25mm, length 30m; film thickness 110 0.25µm; Agilent Technologies, Santa Clara, CA, USA). The helium carrier gas flow was set 111 at 1.0mL/min, and the oven temperature was held for 2min at 50°C, increased to 70°C by 5°C 112 per min and then ramped up to 240°C by 10°C/min. 113

Peaks were identified using the same standards used to produce the floral volatile
blends (Table 1). Isocaryophyllene the rearrangement product of β-caryophyllene, was
identified based on catalytic conversion of β-caryophyllene in clove oil following US patent
3621070 (Rachlin 1971).

118

Preparation of Bees. Honey bees were kept on the University of Southampton campus in an 119 apiary (50° 56' 10"N, 1° 23' 39"W). During the summers of 2012 and 2013, on days when 120 bees showed high activity, 30-35 returning forager bees were caught at their hive entrance 121 122 between 14:00-16:00 GMT. Bees were caught in Sterilin 30mL universal containers (Sterilin Limited, Cambridge, UK) and kept individually. They were immobilized on ice and 123 transferred into blue 1mL pipette tips that were cut to resemble the PER tubes used by 124 Bitterman et al. (1983) and harnessed to it with two strips of cloth tape (tesa[®] extra Power 125 Perfect). Harnessed bees could freely move their antennae, mouth parts and forelegs (Fig. 1). 126 The PER tubes containing bees were kept in Eppendorf centrifuge tube racks placed inside 127 128 plastic boxes with perforated lids. The bottom of the box was lined with wet tissue to provide humidity and prevent the bees from desiccating. The bees were fed with a 30% sucrose 129 solution between 16:00-18:00 and kept in an environmentally controlled room at 20°C 130 131 overnight. PER assays were conducted the next morning for which bees were equally divided into groups of 7-10 bees (depending on overnight survival). 132

133

Proboscis Extension Response Assay. One harnessed bee at a time was placed in the 134 experimental arena (W×D×H= 60cm×45cm×55cm) 3cm in front of an odour delivery system. 135 136 Behind the bee, an extraction fan removed the odour from the arena in order to avoid a buildup of any stimuli. The odour delivery system was custom-built. It allowed for a constant 137 stream of fresh air in the arena through Teflon tubing which flowed into a glass tube. The 138 conditioning stimuli were delivered though a three channel system that ended in the same 139 glass tube as the fresh air, before it discharged into the arena. All three channels consisted of 140 electronic valves and Teflon tubing connected to 10mL glass tubes, in which the stimuli were 141 placed on a piece of filter paper. One channel served as the control (clean air), which was 142 always open when no volatiles were delivered. In order to deliver scented air, the control 143 channel was switched. The airflow passing through the odour delivery part of the system was 144

set at 300mL/min and the fresh air at 200mL/min air. Hence, the bees were constantly 145 146 exposed to an airflow of 500mL/min. The following timeline applied to the PER assay: A bee was allowed to adapt to the experimental arena for 10s, then was presented with the 147 conditioning stimulus (CS) for 10s. Five seconds after the onset of CS, the antennae of the bee 148 was touched with a cocktail stick that was dipped in 30% sucrose solution (unconditioning 149 stimulus (US)). Upon extension of her proboscis, the bee was allowed to feed on the sucrose 150 151 solution for the remaining part of the 10s US period. The bee was then removed from the arena. Thirty seconds was allowed between bees to give the extraction fan additional time to 152 clear the experimental area from any remaining odours. 153

154 The bees were conditioned six times to the CFV blend. During these conditioning trials most bees learned to associate the sugar reward with the odour stimuli. Bees which had 155 learned to extend their proboscis in response to the odour stimuli by the sixth conditioning 156 157 trial were used in the subsequent recognition trials. The recognition trial resembled the conditioning trial but without the use of the US. Its purpose was to test the proboscis response 158 of the bees to the original flower blends and to three different test stimuli for each blend 159 (Table 2). The test stimuli were based on the original blend but some chemical compounds 160 were omitted based on our results from the floral volatile exposure experiment. Extension of 161 162 the proboscis in response to the test stimulus was rated as a positive recognition. The PER assay was repeated until $n \ge 25$ was achieved for each test group of bees. The results show the 163 percentage of bees that had successfully learned to recognize the floral odour blends and then 164 responded to either of the test stimuli. 165

166

Fate of Individual Floral Volatiles. The floral volatile exposure method described above was
altered slightly when investigating the fate of all individual compounds as these analyses were
conducted on a replacement Agilent Technologies 7890B/5977A GC/MS. Because of its
higher sensitivity only 0.2µL of each compound was placed on a filter paper and transferred

into an amber bottle with either ambient air or diesel exhaust. The content of the bottle was 171 172 sampled after 30min, and the SPME fibre was exposed to the volatile for 1min. The SPME fibre was then inserted into the injection port of the GC/MS at 250°C and a split ratio of 10:1. 173 The GC/MS was equipped with a SPME injection sleeve, a HP-Innowax column (I.D. 174 0.25mm, length 30m; film thickness 0.25µm) and a retention gap (I.D. 0.18mm, length 4.1m, 175 film thickness 0µm; both Agilent Technologies, Santa Clara, CA, USA). The helium carrier 176 gas flow and oven temperature programme remained the same as for the floral blend analyses 177 described above. 178

179

Diesel Exhaust Measurements. In order to determine the concentration of toxic and reactive
gases in our diesel exhaust we used a Wolfsense TG501 probe equipped with O₂, NO, NO₂,
SO₂, and CO photochemical sensors (GrayWolf Sensing Solutions, Shelton, CT, USA).
Diesel exhaust was pumped through the probe at a flow rate of 1L/min by attaching a
calibration hood. Measured values were logged every 30s. Five 3min measurements were
taken to mimic the diesel exhaust collection for floral volatile exposures.

186

Influence of Nitrogen Oxides. For the exposures to nitric oxides (NO_X) a filter paper with 1µL 187 188 of the CFV blend and a stir bar were added to an amber 1000mL bottle. Nitrogen oxides were produced by reducing nitric acid with elemental copper, which resulted in a 1:1 ratio of NO 189 and NO₂, which was confirmed by the Wolfsense TG501 probe. Concentrations of 20ppm, 190 2ppm, 0.2ppm per bottle were achieved by using gas tight syringes and volumetric 191 calculations. The bottle was closed with two layers of Parafilm M[®] and a lid. The content was 192 allowed to stir for 30min at 300rpm before the SPME fibre was extended into the bottle. 193 SPME exposure and the GC/MS method were the same as for floral blend volatile exposures. 194 195

Statistical Analysis. As we found that UV light had no statistically significant effect on the 196 composition of volatiles in our experiments, we pooled the data gained from exposures in 197 amber and clear glass bottles. Due to the large number of samples for the floral volatile 198 exposures, it was necessary to use more than one SPME fibre, which did not allow 199 comparative statistical analyses based on peak area according to the SPME guidelines of the 200 journal. However, the SPME fibres used were from the same lot, and the peak area data of the 201 202 floral volatile exposures had less variance than that of the NO_X exposure data that was collected using a single SPME fibre. Therefore, we plotted the peak areas as bar graphs with 203 95% confidence interval error bars, which provides a graphical display of the significance of 204 205 differences of the CFV blend when exposed to either ambient air or diesel exhaust (Fig. 6, online supplementary material). 206

The PER assay data was analysed by comparing the recognition of each test stimuli to the response of bees to the complete CFV blend. Pairwise comparisons were made using χ^2 tests. The criterion for significance was corrected after Bonferroni: $\alpha' = \alpha/k$, where k is the number of comparisons.

211 The influence of NO_X on the CFV blend was measured with the same SPME fibre, so we were able to directly compare the effect of the different NO_X concentrations on the 212 213 abundance of compounds. The peak area data was analysed with a MANOVA followed by univariate ANOVAs for each compound and Tukey post-hoc tests for comparing the 214 influence of different NO_X concentrations on the single compounds. All statistical analyses 215 were conducted with SPSS 20.0 for Windows (IBM Corporation, Armonk, New York, USA). 216 217 218 RESULTS 219

220

Floral Volatile Exposures. At the one minute time point of diesel exhaust exposure of the CFV blend the compounds myrcene, *Z*- and *E*-β-ocimene were drastically reduced, and β -caryophyllene started to be transformed into its *cis*-isomer, isocaryophyllene (Fig. 2). After 30min, when all compounds in the mixture had reached equilibrium in the glass bottle, myrcene remained strongly reduced, whereas *Z*- and *E*-β-ocimene (assignment of geometrical isomers followed Babushok et al. 2011) became undetectable, and β-caryophyllene was completely transformed into isocaryophyllene.

228

Proboscis Extension Response Assays. The recognition assay for the CFV blend revealed that 229 the PER response of the honey bees was significantly influenced by the test stimuli 230 composition. The original CFV blend was recognized by 93% of all bees in the absence of a 231 sugar reward (Fig. 3). The CFV blend without β -ocimene (CFV-1) was still recognized by 232 76% of bees compared to the original blend, which is non-significant ($\chi^2_{(1)} = 2.928$, P = 233 0.092). Omission of myrcene (CFV-2) had a significant effect on the bees' PER response, and 234 only 37% still recognized the test stimuli ($\chi^2_{(1)} = 18.941$, P < 0.001). The test stimuli which 235 lacked β -ocimene and myrcene (CFV-3) was recognized by 39% of the bees, which is 236 significantly lower than the recognition of the full CFV blend ($\chi^2_{(1)} = 17.923$, P < 0.001). 237

238

Fate of Individual Floral Volatiles. Investigation on the effect of diesel exhaust on each single floral volatile showed that β -caryophyllene was transformed into its *cis*-isomer, confirming the results that we had seen when we exposed the complete CFV blend to diesel exhaust. These further investigations revealed that the β -caryophyllene standard sample also contained α -copaene and α -humulene as impurities (verified by standard injection). The amount of both of them was also drastically reduced after being exposed to diesel exhaust (see supplementary material, Fig. 6), however, we were unable to identify any degradation products. Similarly, we

246	could not trace the fate of myrcene and β -ocimene, the two acyclic monoterpenes in the CFV		
247	blend. Benzyl alcohol was partly oxidized to benzaldehyde.		
248			
249	Diesel Exhaust Measurement. The averages of all measurements of all toxic and reactive		
250	gases \pm S.E. acquired with the Wolfsense TG501 probe are given in Table 3. Carbon		
251	monoxide represented the biggest fraction of the emission gases. Within the NO _X fraction, NO		
252	and NO ₂ were produced approximately at a ratio of 1:1. No sulphur dioxide was detected		
253	which is most likely due to the fact that we used low sulphur diesel.		
254			
255	Influence of Nitrogen Oxides. The abundances of myrcene, Z- β -ocimene, E- β -ocimene, β -		
256	caryophyllene, and benzyl alcohol were altered significantly by NO_X (Table 4). In particular,		
257	the abundances of myrcene, Z- β -ocimene, E- β -ocimene, β -caryophyllene, and benzyl alcohol		
258	were significantly reduced when exposed to 20ppm NO _X (Fig. 4). The conversion of β -		
259	caryophyllene into isocaryophyllene began at a concentration of 0.2 ppm NO _X .		
260			
261			
262	DISCUSSION		
263			
264	Exposure to diesel exhaust led to a reduction of myrcene in the CFV blend, β -ocimene		
265	disappeared, and β -caryophyllene was converted into its <i>cis</i> -isomer, isocaryophyllene.		
266	Investigations on nitration mechanisms have shown that the reaction of unsaturated		
267	compounds with nitrogen dioxide, which is abundant in diesel exhaust, can lead to cis/trans-		
268	isomerization (Augusto et al. 2002; Titov 1963). In the blend used for the behavioural		
269	recognition assays we omitted myrcene and β -ocimene. The lack of β -ocimene did not		
270	influence the recognition of the test stimulus significantly. When myrcene was absent,		
271	recognition decreased significantly to only 37%. Similarly, when both compounds were		

missing 39% of honey bees recognized the test stimuli, suggesting that the absence of 272 273 myrcene predominantly contributes to the lack of recognition. Similarly, Reinhard et al. (2010) identified myrcene as a key odorant and ocimene as a non-key odorant in the 274 275 recognition of complex scent mixtures by honey bees. Isocaryophyllene was not available for the behavioural assays, however, we hypothesise that the change from β -caryophyllene to 276 isocaryophyllene could also have changed the insects' odour perception of the CFV blend. 277 278 Geometric isomers may differ in their quality of odour (Roderick 1966), e.g. to humans, the cyclic sesquiterpene (E)- γ -bisabolene smells soapy and spicy, whereas the odour of (Z)- γ -279 bisabolene is described as fatty and woody (Kjeldsen et al. 2003). The discrimination of 280 281 geometric isomers by insects has been mainly studied in the context of straight chain sexpheromones of moths (e.g. Klun et al. 1973; Naka et al. 2013; Silverstein and Young 1976). 282

Investigations into the fates of individual flower volatiles could not detect substantial amounts (if any) of degradation products of the acyclic terpenes myrcene and β -ocimene. On their oxidative way to carbon dioxide, most biogenetic volatiles pass acetone and formaldehyde (Atkinson and Arey 2003), which, due to their low boiling points, would have been difficult to find by our analytical method. However, primary oxidation products of myrcene such as 4-methylene-5-hexenal and of β -ocimene such as 4-methyl-3,5-hexadienal (Lee et al. 2006, Reissell et al. 2002) should have been reliably registered.

Each of the floral scent compounds used in our study that degraded when exposed to diesel exhaust also occur in numerous other plant families: β -caryophyllene in 52% (46 families), β -ocimene in 71% (*E*- β -ocimene in 64 families and *Z*- β -ocimene in 44 families), and myrcene in 70% (63 families) of all investigated seed plant families (Knudsen et al. 2006). Half of the highly ubiquitous flower scent compounds that we tested experienced significant reduction as a result of exposure to diesel exhaust, indicating that it has the potential to disrupt and modify plant floral volatile signalling at a broad scale.

Diesel exhaust emissions have received considerable attention due to their health risks 297 298 and environmental concerns, the main focus has been on particulate matter and NO_X (Setiabudi et al. 2004), which are considered to be the main air pollutants of diesel exhaust 299 together with carbon monoxide and hydrocarbons. The NO_X concentrations of our diesel 300 exhaust were approximately 20ppm for both NO and NO₂. When we exposed the floral 301 volatiles to 20ppm NO_X we observed that it had a similar effect as diesel exhaust: β -302 303 caryophyllene was turned into isocaryophyllene, the amount of myrcene was reduced, β ocimene was reduced as well, but still detectable, whereas it had completely disappeared 304 when exposed to diesel exhaust. In contrast to oxygenation involving oxygen species, radical 305 306 reactions induced by NO_X appear to be responsible for the majority of the changes observed in our CFV blend when exposed to diesel exhaust. In addition, the particulate matter in diesel 307 exhaust may have yet unknown adsorptive and catalytic properties. Because of their isoprene 308 309 (butadiene) substructure, myrcene and β -ocimene are particularly labile, and radical reactions may easily cause polymerisation, leading to products that would escape our analytical system. 310 Ambient air quality standards in the EU (ec.europa.eu) and the US (www.epa.gov) 311 suggest a maximum hourly average NO₂ concentration of 0.1ppm, a concentration at which β -312 caryophyllene started turning into isocaryophyllene in this study, and α -terpinene and 313 314 phenylacetaldehyde were already significantly degraded, in our previous study. Furthermore, in 2013 the UK Supreme Court declared that the nitrogen dioxide limits are regularly 315 exceeded in 16 zones across the UK (European Commission, press release). 316 Ambient NO_X levels can exhibit strong diurnal patterns, with maximum values in the 317 early morning and another increase in the late afternoon (Fuentes et al. 2007), following 318 319 periods of high traffic. Fluctuating NO_X levels will change the ratio of compounds in floral scent throughout the course of the day (McFrederick et al. 2008). Additionally, high NO_X 320 levels lead to increased ozone levels (Fehsenfeld et al. 1992), which may further contribute to 321 the degradation of floral scent. Changing floral scent of a flower can cause floral-constant 322

324 impact on pollination efficiency. To date, there is only little research on how air pollution 325 directly impedes on plant-insect interactions. Fuentes et al. (2013) found that ozone pollution 326 obstructs host finding in the cucumber beetle (<i>Acalymma vittatum</i>). Pinto et al. (2007a; 2007) 327 2008) investigated the impacts of ozone on tritrophic interactions under both laboratory and 328 field conditions. They showed that parasitoids are able to use herbivore-induced plant VOCs 329 to find their hosts in the presence of ozone, but preferred the intact signal over the ozone- 329 degraded signal. There is also a possibility that the reaction products of floral scent 320 compounds with ozone and other air pollutants might form new compounds that may serve a 321 novel signals (McFrederick et al. 2009). 322 Our study demonstrates that NOx and/or other components of diesel exhaust are 323 capable of degrading floral signals and have the potential to alter floral recognition. How this 323 impacts upon the fitness of pollinators requires further investigations, however it has been 326 proposed that pollinators in polluted areas may need to spend more time searching for 326 adequate host plants, which if true could lead to decreased fitness and decreased pollination 327 adequate host plants, wh	323	pollinators, like honey bees, to reject the flower (Wright and Schiestl 2009), which may			
325 directly impedes on plant-insect interactions. Fuentes et al. (2013) found that ozone pollution 326 obstructs host finding in the cucumber beetle (<i>Acalymma vittatum</i>). Pinto et al. (2007a; 2007) 327 2008) investigated the impacts of ozone on tritrophic interactions under both laboratory and 328 field conditions. They showed that parasitoids are able to use herbivore-induced plant VOCs 329 to find their hosts in the presence of ozone, but preferred the intact signal over the ozone- 329 degraded signal. There is also a possibility that the reaction products of floral scent 320 compounds with ozone and other air pollutants might form new compounds that may serve a 321 novel signals (McFrederick et al. 2009). 322 Our study demonstrates that NO _X and/or other components of diesel exhaust are 323 capable of degrading floral signals and have the potential to alter floral recognition. How this 325 impacts upon the fitness of pollinators requires further investigations, however it has been 326 proposed that pollinators in polluted areas may need to spend more time searching for 336 adequate host plants, which if true could lead to decreased fitness and decreased pollination 337 rates (McFrederick et al. 2008). 338 reacer has been provided through the Leverhulme Trust Research Project Grant RPG-08	324	impact on pollination efficiency. To date, there is only little research on how air pollution			
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Table 1 Information about compound purity, vapour pressure, supplier information and their occurrence in seedplant families as reported in Knudsen et al. 2006

Common flower volatile blend					
vapour pressure					
compound	purity (%)	(mm Hg ⁻¹ at 25°C)	supplier	occurrence	
β-Pinene	98	2.9300	Sigma Aldrich ^a	59%	
Myrcene	90	2.0900	Sigma Aldrich ^a	70%	
Limonene	97	1.9800	Sigma Aldrich ^a	71%	
β-Ocimene	\geq 90	1.5590	Sigma Aldrich ^a	71%	
Benzaldehyde	99	0.1270	Sigma Aldrich ^a	64%	
β-Caryophyllene	≥ 80	0.0130	Sigma Aldrich ^a	52%	
Methyl salicylate	\geq 99	0.0343	Fluka ^b	57%	
Benzyl alcohol	99	0.0940	Sigma Aldrich ^a	56%	

^a Sigma Aldrich (St. Louis, MO, USA), ^b Fluka (Sigma-Aldrich, Buchs, Switzerland) 473 Table 2 Compound composition of the common floral volatile and the oilseed rape blend and their according test
 474 stimuli for the proboscis extension response assays

Common flower volatile blend				
compound	CFV	CFV-1	CFV-2	CFV-3
β-Pinene	1 µL	1 µL	1 µL	1 µL
Myrcene ^a	1 µL	1 µL	-	-
Limonene	1 µL	1 µL	1 µL	1 µL
β-Ocimene ^b	1 µL	-	1 µL	-
Benzaldehyde	1 µL	1 µL	1 µL	1 µL
β-Caryophyllene ^c	1 µL	1 µL	1 µL	1 µL
Methyl salicylate	1 µL	1 µL	1 µL	1 µL
Benzyl alcohol	1 µL	1 µL	1 µL	1 µL
dose	8 µL	7 μL	7 μL	6 µL

475

^a compound reduced by diesel exhaust, ^b compounds erased by diesel exhaust, ³ compound turns into its *cis*-isomer

476 Table 3 Diesel exhaust gas composition

gas (unit)	mean		S.E.
Oxygen (%)	19.78	±	0.10
Nitric Oxide (ppm)	19.84	±	2.27
Nitrogen Dioxide (ppm)	17.54	±	1.25
Sulphur Dioxide (ppm)	0.00	±	0.00
Carbon Monoxide (ppm)	226.84	±	20.47

478 Table 4 ANOVA results for single compounds which abundances were significantly decreased by NO_X exposure
 479 (also see Fig. 4)

compound	<i>F</i> -value	P-value
Myrcene	$F_{(3,12)} = 11.803$	0.001
Z-β-Ocimene	$F_{(3,12)} = 89.226$	< 0.001
<i>E</i> -β-Ocimene	$F_{(3,12)} = 96.951$	< 0.001
β-Caryophyllene	$F_{(3,12)} = 568.459$	< 0.001
Benzyl alcohol	$F_{(3,12)} = 9.184$	0.002

481 **Fig. 1** Harnessed honey bee ready for PER assay

482483 Fig. 2 Total ion current (TIC) chromatogram of the common flower volatile under 30min

- ambient air and diesel exhaust exposure. Consecutive numbers represent the following
- 485 chemical compounds: β -Pinene (1), myrcene (2), limonene (3), Z- β -ocimene (4a),
- 486 E- β -ocimene (4b), benzaldehyde (5), β -caryophyllene (6a), isocaryophyllene (6b) methyl
- 487 salicylate (7), benzyl alcohol (8).
- 488
- **Fig. 3** Proboscis extension response assay results representing the ability of honey bees to
- 490 recognize the common floral volatile (CFV) blend, CFV blend without β-ocimene (CFV-1), 491 CFV blend without myrcene (CFV-2), and CFV blend without both (CFV-3). The response to
- the test stimuli was compared to the response to the full blend. A Bonferroni correction was
- 492 applied, and all significant differences are indicated with an asterisk at a 0.016 level of
 494 significance
- 495
- 496 Fig. 4 The effect of 30min exposures of different nitrogen oxides ($NO_X = NO$ and NO_2 at a
- 497 1:1 ratio) concentrations on the abundance of all floral compounds used in this study. Bars
- 498 represent the average peak area (\pm S.D). Different lowercase letters indicate a statistically
- 499 significant difference.
- 500











