

Future global mortality from changes in air pollution attributable to climate change

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1 FUTURE GLOBAL MORTALITY FROM CHANGES IN AIR POLLUTION 2 ATTRIBUTABLE TO CLIMATE CHANGE

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40 Ground-level ozone and fine particulate matter (PM_{2.5}) are associated with premature human mortality¹⁻⁴; their future concentrations depend on changes in emissions, which dominate the 41 near-term⁵, and on climate change^{6,7}. Previous global studies of the air quality-related health 42 effects of future climate change^{8,9} used single atmospheric models. However, in related studies, 43 mortality results differ among models¹⁰⁻¹². Here we use an ensemble of global chemistry-climate 44 models¹³ to show that premature mortality from changes in air pollution attributable to climate 45 change, under the high greenhouse gas scenario RCP8.5¹⁴, is likely positive. We estimate 3,340 46 47 (-30,300 to 47,100) ozone-related deaths in 2030, relative to 2000 climate, and 43,600 (-195,000 48 to 237,000) in 2100 (14% of the increase in global ozone-related mortality). For PM_{2.5}, we 49 estimate 55,600 (-34,300 to 164,000) deaths in 2030 and 215,000 (-76,100 to 595,000) in 2100 50 (countering by 16% the global decrease in PM_{2.5}-related mortality). Premature mortality 51 attributable to climate change is estimated to be positive in all regions except Africa, and is 52 greatest in India and East Asia. Most individual models yield increased mortality from climate 53 change, but some yield decreases, suggesting caution in interpreting results from a single model. 54 Climate change mitigation will likely reduce air pollution-related mortality.

55 Climate change can affect air quality through several pathways, including changes in the 56 ventilation and dilution of air pollutants, photochemical reaction rates, removal processes, 57 stratosphere-troposphere exchange of ozone, wildfires, and natural biogenic and lightning emissions^{6,7}. Overall, changes in these processes are expected to increase ozone in polluted 58 59 regions during the warm season, especially in urban areas and during pollution episodes, but 60 decrease ozone in remote regions due to greater water vapour concentrations leading to greater 61 ozone destruction. These effects are exacerbated by the greater decomposition of reservoir species such as PAN⁷. PM_{2.5} will also be affected by climate change, but impacts vary in sign 62

among models and show regional variation related to differences in precipitation, wildfires,
biogenic emissions, PM_{2.5} composition, and other factors.

65 Previous studies have examined the impact of future climate change on human health via air quality globally^{8-9,15} in the US^{10, 16-20}, and in Europe²¹. However, only two studies have 66 previously used an ensemble of models to assess air pollution-related mortality attributable to 67 climate change: one for the US¹⁰, and our previous global work with the same ensemble used 68 here, but evaluating the effects of historical climate change prior to 2000¹¹. Both studies found a 69 large spread of mortality outcomes depending on the atmospheric model used. Silva et al.¹¹ 70 71 found that the multi-model average suggested a small detrimental effect of climate change on 72 global present-day air pollution-related mortality, but individual models yielded estimates of opposing sign. 73

74 The Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP) ensemble (Supplementary Table 1) simulated air quality in 2000, and in 2030, 2050 and 2100 for the four 75 global Representative Concentration Pathway scenarios (RCPs)²². We previously estimated 76 77 future air pollution premature mortality under all four RCP scenarios, estimating the net effect of both emissions changes and climate change¹². Under RCP8.5, ozone concentrations increase in 78 79 most locations in 2100 relative to 2000, due to increases in methane emissions and the effect of climate change^{7,23}, but PM_{2.5} decreases in 2100 due to a projected decrease in particulate and 80 precursor emissions²⁴. These changes in pollutant concentrations lead to 316,000 (95% C.I.: -81 187,000 to 1.38 million) ozone-related excess deaths yr⁻¹ and -1.31 (-2.04 to -0.17) million 82 PM_{2.5}-related (avoided) deaths yr⁻¹ in 2100¹². Here we present results from additional ACCMIP 83 84 simulations that were designed to isolate the influences of future climate change under RCP8.5. 85 by simulating the projected climates of 2030 and 2100 (imposed by prescribing sea-surface

temperatures, sea ice cover, and greenhouse gas concentrations for radiation) together with air
pollutant emissions from 2000. The effects of climate change are then isolated by a difference
with historical 2000 simulations. Premature mortality attributable to RCP8.5 climate change is
estimated following the methods of Silva et al.¹², including projected population and baseline
mortality rates (see Methods), such that mortality estimates here can be compared directly with
overall changes in air pollution-related mortality in RCP8.5.

92 We estimate that global ozone mortality attributable to RCP8.5 climate change will be 3,340 (-30,300 to 47,100) deaths yr⁻¹ in 2030 and 43,600 (-195,000 to 237,000) deaths yr⁻¹ in 2100 93 94 (Figures 1a and 2a). In 2100, ozone mortality increases in most regions, especially in highly 95 populated and highly polluted areas, with marked spatial differences within regions that include 96 both positive and negative mortality changes (Figure 3a, Supplementary Table 2, Supplementary 97 Figures 1 and 2a). The effect on ozone mortality in 2100 is greatest in East Asia (45,600 deaths yr⁻¹, 41 deaths yr⁻¹ per million people), India (16,000 deaths yr⁻¹, 8 deaths yr⁻¹ per million people) 98 and North America (9,830 deaths yr⁻¹, 13 deaths yr⁻¹ per million people), but some areas within 99 100 these and other regions show decreases in mortality. East Asia has high mortality effects per 101 person in part because of its higher projected mortality rate from respiratory diseases. Climate 102 change contributes 14% of the overall increase in ozone mortality estimated for RCP8.5 in 2100 relative to 2000^{12} . However, three of 8 models in 2030 and three of 9 in 2100 show global 103 104 decreases in ozone mortality due to climate change. For each model, the uncertainty range does 105 not include zero; only the spread of models causes the overall uncertainty to span zero. 106 Uncertainty in modeled ozone concentrations contributes over 97% to the overall uncertainty in 107 both 2030 and 2100, with the remainder from uncertainties in relative risk (RR). Results from a 108 sensitivity analysis using present-day population and baseline mortality rates (Table 1) show

109 32% and 67% lower mortality estimates in 2030 and 2100, respectively, largely because the 110 projected baseline mortality rates of chronic respiratory diseases increase through 2100. The 111 models agree that ozone will increase due to climate change in some polluted regions, notably the northeast US as found in other studies⁶ and decrease in the tropics over the oceans 112 113 (Supplementary Figures 3 and 4a). These changes are consistent with those analysed by Schnell et al.²⁵ for 2100, using four of these same models, and were attributed to a greater efficiency of 114 115 precursor emissions to generate surface ozone in polluted regions, along with reductions in the 116 export of precursors to downwind regions.

117 The impact of climate change on PM_{2.5} mortality is estimated to result in 55,600 (-34,300 to 164,000) deaths yr⁻¹ in 2030 and 215,000 (-76,100 to 595,000) deaths yr⁻¹ in 2100 (Figures 1b 118 119 and 2b). Mean estimates of PM_{2.5} mortality increase in 2100 in all regions except Africa (-25,200 deaths yr⁻¹) (Figure 3b, Supplementary Table 3, Supplementary Figure 2b). The greatest 120 increases in mortality in 2100 occur in India (80,200 deaths yr⁻¹, 40 deaths yr⁻¹ per million 121 people), Middle East (50,400 deaths yr⁻¹, 45 deaths yr⁻¹ per million people) and East Asia 122 (47,200 deaths yr⁻¹, 43 deaths yr⁻¹ per million people), although the Former Soviet Union shows 123 greater mortality per million people in 2100 (11,800 deaths yr⁻¹, 57 deaths yr⁻¹ per million 124 125 people). Similar to ozone mortality, there are substantial spatial differences within each region, including both increases and decreases in mortality. For PM_{2.5}, a large decrease in mortality is 126 projected in RCP8.5 relative to 2000 (when accounting for changes in both emissions and 127 climate)¹², but climate change alone increases mortality, partially counteracting the decrease 128 129 associated with declining emissions in RCP8.5. Without climate change, the decrease in PM_{2.5}-130 related mortality would be roughly 16% greater in 2100 relative to 2000. Propagating 131 uncertainty in RR to the mortality estimates leads to coefficients of variation (CVs) of 8-31%

132 (2030) and 11-46% (2100) for the different models, but the spread of model results increases 133 overall CVs to 123% in 2030 and 106% in 2100. In both years, one model (GISS-E2-R) yields a 134 decrease in global mortality from climate change while the other three (2030) or four (2100) 135 show an increase. Uncertainty in modeled PM2.5 concentrations in 2000 makes a similar 136 contribution to the overall uncertainty (50% in 2030 and 52% in 2100) compared with 137 uncertainty in modeled PM_{2.5} concentrations in future years (50% in 2030, 48% in 2100). 138 Uncertainty in RR makes a negligible contribution in both periods (<1%), as the multi-model 139 mean is small and different models disagree on the sign of the influence. Considering present-140 day population and baseline mortality rates (Table 1), we estimate 23% and 33% lower mortality 141 in 2030 and 2100, respectively, mostly associated with the increase in projected baseline 142 mortality rates through 2100.

143 PM_{2.5}-related mortality was estimated above for the sum of PM_{2.5} species reported by five 144 models, using a common formula (see Methods), to increase the number of models considered 145 and to increase consistency among PM2.5 estimates. Additionally, we present a sensitivity 146 analysis considering the PM2.5 concentrations reported by four models using their own PM2.5 147 formulas, for which multi-model average mortality results are modestly higher: 15% greater in 148 2030 and 12% in 2100 (Supplementary Figure 5). The degree of agreement between the two 149 estimates varies among the four models, and for one model (GISS-E2-R) the two sources of 150 PM_{2.5} estimates yield impacts of different sign in 2030.

151 There is considerable agreement among models regarding the increase in PM_{2.5} concentrations in

152 many locations in 2100, including most polluted regions, due to RCP8.5 climate change

153 (Supplementary Figure 4b). Allen et al.²⁶ analysed four of these same models in 2100 and found

154 that global average surface PM_{2.5} concentrations increased due to climate change, reflecting

155 increases in nearly all relevant species for each model. They attributed this increase in PM_{2.5} 156 mainly to a decrease in wet deposition associated with less large-scale precipitation over land. Our multi-model mean estimates of global population-weighted changes for PM25 and individual 157 species (Supplementary Table 4; Supplementary Figure 6) are similar to those of Allen et al.²⁶. 158 Unlike Allen et al.²⁶, however, GISS-E2-R shows a net decrease in global population-weighted 159 160 concentrations of total PM2.5 and of each PM2.5 species except sea salt, in 2100, likely due to 161 projected concentration decreases over densely-populated eastern China. Models also differ 162 strongly in the sign and magnitude of changes in dust, particularly over North Africa and the 163 Middle East; HadGEM2 projects increases in PM2.5 for all species except dust, but a strong 164 decrease in dust over the Middle East and South Asia. In Africa, the decrease in PM25 near the equator is likely caused by increased precipitation, whereas PM2.5 increases are associated with 165 precipitation decreases in Southern Africa²⁶. Differences in PM_{2.5} (and ozone) responses to 166 167 climate change among models likely result from differences in large-scale meteorological 168 changes, and different treatments of atmospheric chemistry and feedback processes among the 169 models (such as the response of dust to climate change).

In the US, our multi-model mean mortality estimates for the impact of RCP8.5 climate change 170 for ozone (1,130 deaths yr⁻¹ in 2030; 8,810 deaths yr⁻¹ in 2100) compare well with those of Fann 171 et al.²⁰, who report 420 to 1900 ozone-related deaths yr⁻¹ for RCP8.5 climate change in 2030, 172 173 despite differences in concentration-response functions and population and baseline mortality projections. These results for ozone and those for PM_{2.5} (6,900 deaths yr⁻¹ in 2030; 19,400 deaths 174 yr⁻¹in 2100) are also consistent with the increases in mortality and spatial heterogeneity 175 attributed to climate change in 2050 by Bell et al.¹⁶ for ozone and Tagaris et al.¹⁷ for ozone and 176 177 PM_{2.5}, although these studies used different climate change scenarios besides other

methodological differences. Across models, our estimates for ozone mortality in the US vary
between -435 and 4,750 deaths yr⁻¹ in 2030 and between -1,820 and 27,012 deaths yr⁻¹ in 2100.
This spread of model results, with a few models suggesting avoided mortality due to climate
change, is similar to that of Post et al.¹⁰ (-600 to 2,500 deaths yr⁻¹ in 2050) using SRES scenarios
of GHG emissions. Similarly, results show spatial heterogeneity within several regions (Figure
that is similar to Post et al.¹⁰ for the US and Orru et al.²¹ for Europe.

184 The spread of results among models highlights the uncertainty in the effect of climate change on 185 air quality. Further improvements in chemistry climate models are needed to better model the 186 interaction and feedbacks between climate and air quality, including the sensitivity of biogenic 187 emissions to climate change, the effects of meteorological changes on air quality (e.g., aerosol-188 cloud interactions, secondary aerosol formation, wet deposition, and gas-aerosol partitioning), 189 and the impact of climate change on wildfires. Stratosphere-troposphere exchange of ozone is 190 also important, as is the impact of land use changes on regional climate and air pollution. Our 191 results are specific to climate change as projected under RCP8.5 and would differ for other 192 scenarios. We estimate the effect of climate change as the difference between simulations with 193 future climate and year 2000 climate, both with year 2000 emissions, although global emissions of PM_{2.5} and its main precursors decrease under RCP8.5. Had we instead modelled future 194 195 emissions with present vs. future climate, we would likely have attributed smaller changes in air 196 pollution and mortality to climate change, given the projected emission reductions. Whereas the 197 net effect of missing and uncertain processes does not clearly indicate an under- or overestimate 198 for the effect of climate change on air quality, we likely underestimate the magnitude of the 199 health impact by omitting mortality for people under 25, and morbidity effects. We also neglect 200 possible synergistic effects of a warmer climate to modify air pollution-mortality relationships.

Although a few studies have suggested stronger relationships between $ozone^{27}$ and $PM_{2.5}^{28}$ and health at higher temperatures, there is insufficient evidence to include those effects here.

203 Despite these uncertainties, this study is the first to use a multi-model ensemble to show that 204 global air pollution-related mortality attributable to climate change is likely positive. The spread 205 of results among models within the ensemble, including differences in the sign of global and 206 regional mortality estimates, suggests that results from studies using a single model and a small 207 number of model years should be interpreted cautiously. Actions to mitigate climate change, 208 such as reductions in long-lived GHG emissions, will likely benefit human health by reducing 209 the effect of climate change on air quality in many locations. These health benefits are likely to be smaller than those from reducing co-emitted air pollutants²⁹, but both types of health benefits 210 211 via changes in air quality would add to reductions in many other influences of climate change on human health³⁰. 212

213

214 Additional information

215 Supplementary information is available in the online version of the paper.

216 **References**

- Jerrett, M. *et al.* Long-Term Ozone Exposure and Mortality. *N. Engl. J. Med.* 360, 1085–
 1095 (2009).
- Krewski, D. *et al.* Extended Follow-Up and Spatial Analysis of the American Cancer
 Society Study Linking Particulate Air Pollution and Mortality. *Respir. Rep. Heal. Eff. Inst.* 140, 5–114 (2009).
- Lepeule, J., Laden, F., Dockery, D. & Schwartz, J. Chronic Exposure to Fine Particles and Mortality: An Extended Follow-up of the Harvard Six Cities Study from 1974 to 2009. *Environ. Health Perspect.* **120**, 965–970 (2012).

225 4. Burnett, R. T. et al. An Integrated Risk Function for Estimating the Global Burden of 226 Disease Attributable to Ambient Fine Particulate Matter Exposure. Environ. Health 227 Perspect. 122, 397-403 (2014). 228 5. Kirtman, B. et al. Near-term Climate Change: Projections and Predictability. In: Climate 229 Change 2013: The Physical Science Basis (Chapter 11). Contribution of Working Group 230 I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change 231 [Stocker, T.F.et al. (eds.)]. Cambridge University Press, Cambridge, United Kingdom 232 and New York, NY, USA (2013). 233 Fiore, A. M., Naik, V. & Leibensperger, E. M. Air Quality and Climate Connections. J 6. 234 Air Waste Manage. Assoc. 65, 645–685 (2015). 235 von Schneidemesser, E. et al. Chemistry and the Linkages between Air Quality and 7. 236 Climate Change. Chem. Rev. 150430065937004 (2015). 237 doi:10.1021/acs.chemrev.5b00089 238 West, J. J., Szopa, S. & Hauglustaine, D. A. Human mortality effects of future 8. 239 concentrations of tropospheric ozone. C. R. Geosci. 339, 775-83, 2007. 240 9. Selin, N. E. et al. Global health and economic impacts of future ozone pollution. Environ. 241 Res. Lett. 4, 044014 (2009). 242 Post, E. S. et al. Variation in Estimated Ozone-Related Health Impacts of Climate 10. 243 Change due to Modeling Choices and Assumptions. Environ. Health Perspect. 120, 244 1559-1564 (2012). 245 Silva, R. A. et al. Global premature mortality due to anthropogenic outdoor air pollution 11. 246 and the contribution of past climate change. Environ. Res. Lett. 8, 034005 (2013). 247 12. Silva, R. A. et al. The effect of future ambient air pollution on human premature 248 mortality to 2100 using output from the ACCMIP model ensemble. Atmos. Chem. Phys. 249 16, 9847-9862 (2016). Lamarque, J. F. et al. The Atmospheric Chemistry and Climate Model Intercomparison 250 13. 251 Project (ACCMIP): Overview and description of models, simulations and climate diagnostics. Geosci. Model Dev. 6, 179–206 (2013). 252 253 14. Stevenson, D. S. et al. Tropospheric ozone changes, radiative forcing and attribution to 254 emissions in the Atmospheric Chemistry and Climate Model Intercomparison Project 255 (ACCMIP). Atmos. Chem. Phys. 13, 3063–3085 (2013). 256 Fang, Y., Mauzerall, D. L., Liu, J., Fiore, A. M. & Horowitz, L. W. Impacts of 21st 15. century climate change on global air pollution-related premature mortality. Clim. Change 257 258 121(2), 239–253 (2013) 259 16. Bell, M. L. et al. Climate change, ambient ozone, and health in 50 US cities. Clim. 260 Change 82, 61–76 (2007). Tagaris, E. et al. Potential Impact of Climate Change on Air Pollution-Related Human 261 17. 262 Health Effects. Environ. Sci. Technol. 43, 4979–4988 (2009). Chang, H. H., Zhou, J. & Fuentes, M. Impact of Climate Change on Ambient Ozone 263 18. 264 Level and Mortality in Southeastern United States. Int. J. Environ. Res. Public Health 7, 265 2866-2880 (2010). 266 Sheffield, P. E., Knowlton, K., Carr, J. L. & Kinney, P. L. Modeling of Regional Climate 19. Change Effects on Ground-Level Ozone and Childhood Asthma. Am. J. Prev. Med. 41, 267 251-257 (2011). 268

Fann, N. et al. The geographic distribution and economic value of climate change-related 269 20. 270 ozone health impacts in the United States in 2030. J. Air Waste Manage. Assoc. 65, 570-271 580 (2015). 272 21. Orru, H. et al. Impact of climate change on ozone-related mortality and morbidity in 273 Europe. Eur. Respir. J. 41, 285–294 (2013). 274 22. van Vuuren, D. P. et al. The representative concentration pathways: an overview. Clim. 275 *Change* **109**, 5–31 (2011). 276 23. Young, P. J. et al. Pre-industrial to end 21st century projections of tropospheric ozone 277 from the Atmospheric Chemistry and Climate Model Intercomparison Project 278 (ACCMIP). Atmos. Chem. Phys. 13, 2063–2090 (2013). 279 24. Shindell, D. T. et al. Radiative forcing in the ACCMIP historical and future climate 280 simulations. Atmos. Chem. Phys. 13, 2939–2974 (2013). 281 25. Schnell, J. L. et al. Effect of climate change on surface ozone over North America, 282 Europe, and East Asia. Geophys. Res. Lett., 43, 3509-3518 (2016). 283 Allen, R.J., Landuyt, W. & Rumbold, T. An increase in aerosol burden and radiative 26. 284 effects in a warmer world. Nat. Clim. Change 6, 269-274 (2016). 285 Wilson, A., Rappold, A. G., Neas, L. M. & Reich, B. J. Modeling the effect of 27. 286 temperature on ozone-related mortality. Ann. Appl. Stat. 8, 1728–1749 (2014). 287 Ren, C., Williams, G. M. & Tong, S. Does Particulate Matter Modify the Association 28. between Temperature and Cardiorespiratory Diseases?. Environ. Health Perspect. 114, 288 289 1690-1696 (2006). 290 29. West, J. J. et al. Co-benefits of Global Greenhouse Gas Mitigation for Future Air Quality 291 and Human Health. Nat. Clim. Change 3, 885-889 (2013). 292 Smith, K.R. et al. Human health: impacts, adaptation, and co-benefits. In: Climate 30. 293 Change 2014: Impacts, Adaptation, and Vulnerability. Part A: Global and Sectoral 294 Aspects. Contribution of Working Group II to the Fifth Assessment Report of the 295 Intergovernmental Panel on Climate Change [Field, C.B. et al. (eds.)], pp. 709-754 296 (2014). 297 298

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317	Author contributions: JJW, JFL, DTS and RAS conceived the study. All other co-authors				
318	conducted the model simulations. RAS processed model output and estimated human mortality.				
319	RAS and JJW analyzed results. RAS and JJW prepared the manuscript and all co-authors				
320	commented on it.				

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325 Figure Legends:

326 Figure 1 – Impact of RCP8.5 climate change on global mortality for individual models and the

multi-model average. Estimates are for 2030 and 2100 for (a) ozone respiratory mortality (9

328 models) and (b) PM2.5 IHD+STROKE+COPD+LC mortality (5 models). PM2.5 is calculated as

a sum of species. Uncertainty for each model is the 95% CI taking into account uncertainty in

RR. Uncertainty for the multi-model average is the 95% CI including uncertainty in RR andacross models.

331 ac 332

Figure 2 – Geographical impact of climate change on mortality. Estimates are for 2030 and 2100

for (a) ozone respiratory mortality and (b) PM2.5 IHD+STROKE+COPD+LC mortality,

showing the multi-model average in each 0.5°x0.5° grid cell. PM2.5 is calculated as a sum of

336 species.

- 338 Figure 3 Projected mortality for ten world regions. Estimates are for 2030 and 2100 for (a)
- 339 ozone respiratory mortality and (b) PM2.5 IHD+STROKE+COPD+LC mortality, showing the
- 340 multi-model regional average. PM2.5 is calculated as a sum of species. Uncertainty for the multi-
- 341 model regional average is the 95% CI including uncertainty in RR and across models. World
- regions are shown in Supplementary Figure 1.

Table 1 – Sensitivity analysis for changes in global air pollution-related mortality attributable to climate change. Estimates are for multi-model averages (deaths yr^{-1}) for the deterministic

results.

	PM _{2.5} -related mortality		Ozone-related mortality	
	2030	2100	2030	2100
Base results	56,300	218,000	10,700	128,000
$PM_{2.5}$ using Krewski et al. ²	66,200	318,000		
Present-day (2011) population	35,500	93,800	2,970	59,400
Present-day (2010) baseline	69,600	510,000	2,790	13,300
mortality rates				
Present-day population and	43,300	144,000	2,300	14,500
baseline mortality rates				

353 Methods

The Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP)¹³ included 354 355 contributions from 14 modelling groups, of which 9 completed simulations that are used here 356 (Supplementary Table 1). ACCMIP models incorporate chemistry-climate interactions, including 357 mechanisms by which climate change affects ozone and PM_{2.5}, although models do not all include the same interactions, and do not always agree on their net effects⁷. Of these nine, three 358 359 models are not truly coupled chemistry-climate models: MOCAGE is a chemical transport model 360 driven by external meteorology, and UM-CAM and STOC-HadAM3 do not model the feedback of chemistry on climate¹³. As a result, these models do not fully capture the effects of changes in 361 362 air pollutant concentrations on processes that affect meteorology, such as through radiative 363 transfer and clouds. Prescribed anthropogenic and biomass burning emissions were very similar 364 for the different models, but they used different natural emissions (e.g. biogenic volatile organic compounds, ocean emissions, soil and lightning NO_x)^{14, 23}. Modelled 2000 concentrations show 365 good agreement with observations for ozone²³ and $PM_{2.5}^{24}$, although models tend to overestimate 366 367 ozone in the Northern Hemisphere and underestimate it in the Southern Hemisphere, and to 368 underestimate PM_{2.5}, particularly in East Asia.

We isolate the effect of climate change on air quality as the difference in concentrations between ACCMIP simulations using year 2000 emissions together with future year climate, imposed by prescribing RCP8.5³¹ sea surface temperatures, sea ice cover, and GHGs (for radiation) for 2030 and 2100 (referred to as "Em2000Cl2030" and "Em2000Cl2100"), and simulations with 2000 emissions and climate ("acchist2000")¹³. We analyse results from the nine models reporting ozone from the Em2000Cl2030/2100 simulations, and the five reporting PM_{2.5} (Supplementary Table 1). Ozone and PM_{2.5} species surface concentrations from each model are calculated in each 376 grid cell, after regridding output from the native horizontal resolutions of each model $(1.9^{\circ}x1.2^{\circ}$ 377 to $5^{\circ}x5^{\circ}$) to a common $0.5^{\circ}x0.5^{\circ}$ resolution. To be consistent with the epidemiological studies 378 considered^{1,4}, we use the seasonal average of daily 1-hr maximum ozone concentrations for the 379 six consecutive months with highest concentrations in each grid cell, and annual average PM_{2.5} 380 concentration.

- 381 Seven of the nine models with Em2000Cl2030/2100 simulations reported both hourly and
- 382 monthly ozone concentrations, while two reported only monthly values. We calculate the ratio
- 383 of the 6-month average of daily 1-hr maximum concentrations to the annual average
- 384 concentrations, for each grid cell and each year, for those models that reported both hourly and
- monthly concentrations; then, we apply that ratio to the annual average ozone concentrations for the other two models, following Silva *et al.*^{11,12}.
- 387 We calculate $PM_{2.5}$ concentration using the sum of $PM_{2.5}$ species mass mixing ratios reported by 388 five models and a common formula:

where BC – Black Carbon, OA – (Primary) Organic Aerosol corrected to include species other

PM2.5 = BC + OA + SO4 + SOA + NH4 + 0.25*SS + 0.1*Dust,

390

than carbon, NH4 – NH₄ in ammonium sulfate, SOA – Secondary Organic Aerosol, and SS – Sea Salt, as had been done previously by Fiore et al.³³ and Silva *et al.*^{11,12}. The factors 0.25 and 0.1 are intended to approximate the fractions of sea salt and dust that are in the PM_{2.5} size range. Nitrate was reported by three models, but we chose to omit nitrate from our PM_{2.5} formula to avoid imposing changes inconsistent with the effect of climate change for other models, following Silva *et al.*¹¹, although nitrate was included in estimates of total PM_{2.5} by Silva *et al.*¹². Four of these models also reported their own estimate of PM_{2.5} (Supplementary Table 1). The impact of climate change on global population-weighted differences (Em2000Cl2030/2100 minus acchist2000) in PM_{2.5} and ozone concentrations for the different models are shown in Supplementary Tables 4 and 5, respectively, while regional multi-model average differences are shown in Supplementary Figures 7 and 8.

402 We estimate premature mortality by calculating the fraction of cause-specific mortality 403 attributable to long-term changes in pollutant concentrations, using methods that are identical to those of Silva et al.¹², so that mortality attributable to climate change can be compared simply 404 405 with changes in mortality under the RCP scenarios. We use relative risks (RRs) from Jerrett et al.¹ for ozone and respiratory diseases and Burnett et al.⁴ for PM_{2.5} and cardiopulmonary diseases 406 407 and lung cancer. Then, we apply that attributable fraction in each grid cell to future adult 408 population (age 25 and older) and baseline mortality rates based on projections from the International Futures (IFs) integrated modelling system³². Using country-level projections per 409 410 age group, we mapped and gridded to the 0.5°x0.5° grid assuming that the present-day spatial 411 distribution of total population within each country is unchanged in the future, as well as the 412 present-day ratio of baseline mortality for the specific causes included in the epidemiological 413 studies and for three disease groups projected in IFs (chronic respiratory diseases, cardiovascular 414 diseases and malignant neoplasms). We select population projections from IFs instead of those 415 underlying RCP8.5 to ensure consistency between projections of population and baseline mortality, since the latter are not available for RCP8.5, and for consistency with Silva et al.¹². IFs 416 417 projections of future total population are lower than those of RCP8.5 (-5% in 2030 and -27% in 418 2100) (Supplementary Figure 9). Had we used projections of population underlying RCP8.5, we 419 would have likely estimated greater changes in premature mortality relative to 2000. IFs 420 projections of baseline mortality rates reflect an aging population and regional demographic

421 changes, showing a steep rise in chronic respiratory diseases (roughly tripling globally by 2100),
422 particularly in East Asia and India, some regional increases in cardiovascular diseases (e.g.
423 Middle East, Africa), and global decreases in lung cancer.

424 Overall uncertainty in mortality estimates includes uncertainty from the RRs and from air 425 pollutant concentrations. First, we conduct 1000 Monte Carlo (MC) simulations separately for 426 each model-year to propagate uncertainty from the RRs to mortality estimates. For ozone, we use the 95% Confidence Intervals (CIs) for RR reported by Jerrett *et al.*¹ and assume a normal 427 distribution, while for PM_{2.5} we use the parameter values of Burnett *et al.*⁴ for 1000 MC 428 429 simulations. Then, we calculate the average and 95% CI for the pooled results of the 1000 MC 430 simulations for each model to quantify the spread of model results. We do not include 431 uncertainties associated with population and baseline mortality rates, since these are not reported. 432 As ACCMIP models used the same anthropogenic and biomass burning emissions, we do not 433 consider uncertainty in emissions inventories, however we acknowledge that this is an important source of uncertainty, especially in particular regions³⁴⁻³⁷. Our mortality estimates are affected by 434 435 our choices of and underlying assumptions regarding concentration-response functions, 436 population, and baseline mortality rates. Although a number of factors, such as vulnerability of 437 the exposed population and $PM_{2.5}$ composition, vary spatially and possibly temporally, we 438 assume that the RRs estimated for the present day apply on a global scale and in future time 439 periods. Also, our assumption that the spatial distribution of population within each country is 440 constant in the future likely understates the effects of rural-to-urban migration, which is currently 441 underway and expected to continue. However, the effects of climate change on air pollutant 442 concentrations may be somewhat spatially uniform (as opposed to changes in emissions), and the

443 coarse grid resolution of global models would not resolve air pollutant concentrations well in444 urban areas.

445

446 Data Availability

- 447 Data used in this project are archived here:
- 448 <u>Air pollutant concentrations</u>: Atmospheric Chemistry & Climate Model Intercomparison Project
- 449 (ACCMIP) datasets http://catalogue.ceda.ac.uk/uuid/b46c58786d3e5a3f985043166aeb862d.
- 450 Data retrieved from 08/2012 to 12/2013.
- 451 Present-day population: Oak Ridge National Laboratory (ONRL) LandScan 2011 Global
- 452 Population Dataset, <u>http://spruce.lib.unc.edu.libproxy.lib.unc.edu/content/gis/LandScan/</u>. Data
- 453 retrieved on 12/05/2012.
- 454 <u>Present-day baseline mortality</u>: Institute for Health Metrics and Evaluation (IHME): Global
- 455 Burden of Disease Study 2010 (GBD 2010) Results by Cause 1990-2010 Country Level,
- 456 Seattle, United States, 2013.
- 457 <u>https://cloud.ihme.washington.edu/index.php/s/d559026958b38c3f4d12029b36d783da?path=%2</u>
- 458 <u>F2010</u>. Data retrieved from 12/2013 to 03/2014.
- 459 <u>Future population and baseline mortality</u>: Web-Based IFs The International Futures (IFs)
- 460 modeling system, version 6.54., <u>www.ifs.du.edu</u>. Data retrieved on 07/2012.
- 461 <u>IER model</u>: Global Burden of Disease Study 2010. Global Burden of Disease Study 2010 (GBD
- 462 2010) Ambient Air Pollution Risk Model 1990 2010. Seattle, United States: Institute for

- 463 Health Metrics and Evaluation (IHME), 2013. <u>http://ghdx.healthdata.org/record/global-burden-</u>
- 464 <u>disease-study-2010-gbd-2010-ambient-air-pollution-risk-model-1990-2010</u>. Data retrieved on
- 465 11/08/2013.
- 466 **References**
- 467 31. Riahi, K. *et al.* RCP 8.5 A scenario of comparatively high greenhouse gas emissions.
 468 *Clim. Change* 109, 33–57 (2011).
- 469 32. Hughes, B. B. *et al.* Projections of global health outcomes from 2005 to 2060 using the
 470 International Futures integrated forecasting model. *Bull. World Health Organ.* 89, 478–486
 471 (2011).
- 472 33. Fiore, A. M. et al. Global air quality and climate. Chem. Soc. Rev. 41, 6663–6683 (2012).
- 473 34. Bond, T. C. *et al.* Historical emissions of black and organic carbon aerosol from energy-474 related combustion, 18502000. *Glob. Biogeochem. Cycles* **21**, GB2018 (2007).
- Schopp,W., Klimont, Z., Suutari, R. & Cofala, J. Uncertainty analysis of emissions
 estimates in the RAINS integrated assessment model. *Environ. Sci. Policy* 8, 601613
 (2005).
- 478 36. Smith, S. J. *et al.* Anthropogenic sulfur dioxide emissions: 1850-2005. *Atmos. Chem. Phys.*479 **11**, 11011116 (2011).
- 480 37. Granier, C. *et al.* Evolution of anthropogenic and biomass burning emissions of air
 481 pollutants at global and regional scales during the 1980-2010 period, *Clim. Change*, 109,
 482 163-190 (2011).
- 483

484







a. Ozone mortality



b. PM_{2.5} mortality



a. Ozone mortality

(deaths yr⁻¹)



b. PM_{2.5} mortality (deaths yr¹)

