

*Spatial mixing of mitochondrial lineages and greater genetic diversity in some invasive populations of the American mink (*Neovison vison*) compared to native populations*

Article

Accepted Version

García, K., Melero, Y. ORCID: <https://orcid.org/0000-0002-4337-1448>, Palazón, S., Gosálbez, J. and Castresana, J. (2017) Spatial mixing of mitochondrial lineages and greater genetic diversity in some invasive populations of the American mink (*Neovison vison*) compared to native populations. *Biological Invasions*, 19 (9). pp. 2663-2673. ISSN 1387-3547 doi: <https://doi.org/10.1007/s10530-017-1475-4> Available at <https://centaur.reading.ac.uk/83099/>

It is advisable to refer to the publisher's version if you intend to cite from the work. See [Guidance on citing](#).

To link to this article DOI: <http://dx.doi.org/10.1007/s10530-017-1475-4>

Publisher: Springer

All outputs in CentAUR are protected by Intellectual Property Rights law, including copyright law. Copyright and IPR is retained by the creators or other copyright holders. Terms and conditions for use of this material are defined in

the [End User Agreement](#).

www.reading.ac.uk/centaur

CentAUR

Central Archive at the University of Reading

Reading's research outputs online

1 **Spatial mixing of mitochondrial lineages and greater genetic diversity in some invasive populations**
2 **of the American mink (*Neovison vison*) compared to native populations**

3

4 **Karla García^{1*}, Yolanda Melero², Santiago Palazón^{1,3}, Joaquim Gosálbez¹ and Jose Castresana⁴**

5

6 ¹Department of Evolutionary Biology, Ecology and Environmental Sciences, University of Barcelona,
7 08028 Barcelona, Spain. *Correspondence: karla.garcia.cl@gmail.com

8

9 ²CREAF, 08193 Cerdanyola del Vallès, Spain.

10

11 ³Fauna and Flora Service, Department of Territory and Sustainability, Government of Catalonia, 08017
12 Barcelona, Spain.

13

14 ⁴Institute of Evolutionary Biology (CSIC-Universitat Pompeu Fabra), 08003 Barcelona, Spain.

15

16

17 **Abstract**

18

19 The genetic characteristics of introduced populations have a relevant impact on their ability to
20 establish and spread. The American mink (*Neovison vison*), native to North America, is an important
21 invasive species in the Iberian Peninsula. Here, we used mitochondrial DNA sequences data to investigate
22 the genetic diversity and phylogeographic structure of invasive vs. native populations of this species. We
23 also evaluated whether genetic diversity in invasive populations could be explained by the genetic
24 characteristics of the native sources from which they derived. Phylogenetic analysis revealed two major
25 lineages in the native range, which indicated a clear separation between western and eastern populations.
26 On the contrary, we found no evidence of genetic structure in the invasive range. This was probably the
27 result of the diverse origins of the released specimens and the rapid expansion and encounters of the
28 introduced populations. We detected spatial mixing of both North American lineages in several sampling
29 localities of the north central area of the Iberian Peninsula, giving rise to high levels of genetic diversity
30 in some areas compared to North American populations. This could potentially lead to higher fitness of
31 these individuals and thus increase the population viability and invasiveness of this species. These results
32 point to the need to better study the populations in which lineages mix and, if necessary, intensify control
33 efforts in them.

34

35 **Keywords:** American mink, mitochondrial genes, genetic diversity, native range, Iberian Peninsula.

36 1. Introduction

37
38 Identifying the biological attributes of successful invaders is among the most pressing questions yet to
39 be answered (Sol et al. 2012). Although there is evidence that species differ in their invasion potential,
40 controversy exists about the characteristics that differentiate successful and unsuccessful invaders
41 (Blackburn et al. 2009). Other studies of introduced species show that genetics might play an integral role
42 in the success of an invasion, and they suggest that this role is determined by the quality of the genetic
43 variation introduced, rather than the total quantity of genetic variation (see Dlugosch et al. 2015). The
44 genetic diversity of founding populations, along with potential increase in genetic diversity from
45 subsequent introductions, may be important to the successful establishment and spread of introduced
46 species (Suarez and Tsutsui 2008; Sol et al. 2012). Nevertheless, there is no consensus on whether
47 invaders stemming from multiple native origins are more successful than those from single populations,
48 or whether demographic bottlenecks may limit the invasion success of a species (Edelaar et al. 2015;
49 Estoup 2016).

50
51 Overall, introduced populations lose genetic variation relative to their source populations (Dlugosch
52 and Parker 2008). However, numerous introduction events and a greater number of source populations
53 may buffer against such genetic losses (Kolbe et al. 2004; Dlugosch and Parker 2008; Uller and Leimu
54 2011). In fact, in some cases successful invaders show large increases in genetic diversity in the
55 introduced range, apparently caused by mixing of populations from different source regions (Genton et al.
56 2005; Novak and Mack 1993). Therefore, it is crucial to obtain information on the phylogeographic
57 structure and genetic diversity throughout the native and introduced ranges to understand the factors
58 affecting genetic diversity during invasion (e.g. multiple introduction events or admixture levels among
59 populations) and the interaction between genetic diversity and invasive potential (Edelaar et al. 2015).

60
61 The American mink (*Neovison vison*) is a successful invader with a native range restricted to North
62 America and with invasive populations distributed across Europe, Asia and South America (Bonesi and
63 Palazón 2007; Dunstone 1993). In Europe, this species was introduced for commercial fur farming during
64 the 20th century. Populations established by escaped individuals are now naturalised in fresh water and
65 coastal ecosystems (Bonesi and Palazón 2007), with detrimental consequences for the abundance of many
66 native species of prey and competitors of conservation importance (Bonesi et al. 2004; Harrington et al.
67 2009; Macdonald and Harrington 2003; Melero et al. 2012a), including critically endangered species such
68 as the European mink (*Mustela lutreola*) (Maran et al. 2016) and the Pyrenean desman (*Galemys*
69 *pyrenaicus*) (Fernandes et al. 2008). Fur farms were established in many countries and harboured mink
70 stocks from different parts of their native range. Because of the large scale and long-term nature of this
71 farming, the number of escapees, which started to breed in the wild and became the founder individuals
72 of feral populations, was very high (Bonesi and Palazón 2007). In Spain, mink farming started at the end
73 of the 1950s and the number of farms reached a peak in the 1980s with about 400 farms concentrated
74 mainly in the North West (Galicia) (Ruiz-Olmo et al. 1997). Today, feral mink populations are widely
75 distributed across almost all of the north and centre of the Iberian Peninsula. During the last few decades,
76 these populations experienced a rapid expansion and increased the number of nuclei in almost every

77 colonised area (Ruiz-Olmo et al. 1997; Santulli 2014). Although the number of mink farms in Spain was
78 considerably reduced at the end of the 1980s, mink escapes from the last active farms are still ongoing
79 (Bonesi and Palazón 2007).

80

81 In Europe, there is genetic information on feral American mink populations from several countries.
82 Previous studies in Spain using microsatellite data determined that American mink populations from
83 disjunct geographical areas appeared differentiated and not yet connected by gene flow (Lecis et al.
84 2008). In Scotland, Zalewski et al. (2009) and Fraser et al. (2013) found that landscape features restricted
85 gene flow within this species. Other studies in Poland found that multiple introductions determine the
86 genetic structure of American mink populations (Zalewski et al. 2010) and that these populations exhibit
87 high mitochondrial genetic diversity (Zalewski et al. 2011). A study in France using microsatellite data
88 suggested recent admixture between populations that had recently encountered each other (Bifulchi et al.
89 2010). Similarly, in Sweden it was found, also using microsatellite data, that genetic diversity decreased
90 over six consecutive years of mink culling as part of a population control programme (Zalewski et al.
91 2016).

92

93 Despite the importance of understanding the genetic diversity within the native populations for
94 investigating the evolutionary dynamics of invasions, no phylogeographic study has been performed so
95 far in the North American native range of the American mink. Therefore, many genetic studies performed
96 thus far are difficult to interpret. We used mitochondrial DNA (mtDNA) sequence data from populations
97 in the native range to elucidate whether invasive populations in the Iberian Peninsula have single or
98 multiple origins from the North American sources. We aimed to determine the phylogeographic structure
99 of native and invasive populations, and to evaluate whether mitochondrial genetic diversity observed in
100 established invasive populations can be explained by genetic characteristics of native populations. We
101 examined whether the genetic variability of the Iberian population represented only a small part of that in
102 its native range, as expected for many invasive species or, due to the introduction history and particular
103 origin of these populations, this genetic variability was larger. Overall, the study aims to provide
104 recommendations for mink control in the Iberian Peninsula with respect to determining key control areas.

105

106 **2. Materials and methods**

107

108 **2.1 Sample collection**

109

110 We used hair and tissue samples from invasive populations provided by different Autonomous
111 Communities of Spain. These samples were collected during population control operations conducted by
112 the local administrations between 2007 and 2016 as part of eradication programs and were conserved in
113 tubes containing absolute ethanol (ESM Appendix: Table S1). A total of 105 samples from Catalonia
114 (16), La Rioja (15), Galicia (12), Basque Country (20), Navarra (11), Castellon (7) and Castilla-Leon (24)
115 were included in the study.

116

117 Moreover, tissue samples of 21 specimens from North America (ESM Appendix: Table S2) were
118 obtained from different biological collections: University of Alaska Museum (12), Museum of
119 Southwester Biology (5) and Cornell University Museum of Vertebrates (4).

120

121 **2.2 DNA extraction, PCR and sequencing of mitochondrial DNA**

122

123 We extracted DNA using the QIAGEN DNeasy Blood and Tissue Kit, following the manufacturer's
124 instructions, in a final elution volume of 100 μ l. These extractions were carried out in a separated UV-
125 irradiated area with dedicated equipment. Samples were incubated in a lysis buffer with proteinase K at
126 56°C overnight to ensure maximum tissue lysis.

127

128 For each sample, we amplified a 573 bp fragment of the 5' distal part of the mitochondrial DNA
129 control region (D-loop). For this purpose, we designed novel primers using complete mitochondrial
130 sequences of several mustelids: Neovison_tRNA_pro (5'-ATC AGC ACC CAA AGC TGA CAT TCT
131 A-3') and Neovison_Dloop_556r (5'-TGT GTA TGT CCT GTG ACC ATT GAC T-3'). PCR reactions
132 were set up in a dedicated PCR clean room that is physically separated from post-PCR working areas and
133 regularly decontaminated by UV-irradiation. PCR reactions were performed in a final volume of 25 μ l,
134 containing 2–4 μ l of genomic DNA, 1 μ M of each primer, 0.2 mM dNTPs, 0.75 units of Promega GoTaq
135 DNA polymerase and 17.5 μ M of bovine serum albumin under the following conditions: an initial
136 denaturation of 2 min at 95°C, followed by 35 cycles of denaturation (30 s at 95°C), annealing (30 s at
137 54°C) and extension (30 s at 72°C). PCR products were visualised by electrophoresis in a 1% agarose
138 SYBR-Safe (Invitrogen) stained gel.

139

140 PCR products were purified using ExoSAP-It (Affymetrix), and both strands were sequenced using
141 the original PCR primers at Macrogen Europe (Amsterdam, The Netherlands). Sequences were inspected,
142 trimmed and assembled using Geneious Pro (Biomatters Ltd.). All sequences obtained in this study were
143 deposited in European Nucleotide Archive/GenBank under accession numbers LT854061-LT854186.

144

145 **2.3 Phylogenetic and genetic diversity analyses**

146

147 Consensus sequences were aligned using MAFFT version 7.130 (Katoh et al. 2002) with default
148 settings. Due to the presence of several gaps and ambiguous positions in a very specific region of the
149 alignment, we manually removed positions 92 to 110, thus leaving 554 bp in the alignment.

150

151 A maximum likelihood phylogenetic tree was calculated with RAxML version 8.0 (Stamatakis 2014)
152 using a general time reversible substitution model and rate heterogeneity modelled with a gamma
153 distribution plus a proportion of invariable sites, as recommended by the program. From this tree, a
154 haplotype genealogy was generated using Haploviewer 1.0 (Salzburger et al. 2011).

155

156 The nucleotide diversity (π) was estimated using the R package PEGAS (Population and Evolutionary

157 Genetics Analysis System, Paradis 2010). To analyse the variation in nucleotide diversity across the
158 Iberian Peninsula, we used a visualization method that does not require the pre-definition of populations.
159 The nucleotide diversity was estimated at each sampling location by using all samples present within 30
160 km of the location. This allowed the estimation of genetic diversity from a sufficient number of samples
161 at each point, yet the resolution was sufficient to distinguish regional differences in genetic diversity.
162 QGIS (<http://qgis.osgeo.org>) was used to plot the nucleotide diversity in the map. In the case of the North
163 American native range, it was not possible to apply this analysis due to the limited number of samples.
164 Therefore, we selected three areas with a sufficient number of samples: North West (British Columbia,
165 six samples), South East (Texas, three samples) and East (New York, four samples), in order to estimate
166 the nucleotide diversity in these specific areas.

167

168 **3. Results**

169

170 **3.1 Mitochondrial phylogeographic analysis**

171

172 A total of 16 different haplotypes of the mitochondrial control region were detected in 126 American
173 mink, 11 of which were recorded in North America and 7 in the Iberian Peninsula (ESM Appendix: Fig.
174 S1). The maximum likelihood phylogenetic tree showed two large clades found throughout western and
175 eastern North America, respectively (Fig. 1). The haplotype genealogy reconstructed from the maximum
176 likelihood tree revealed the same two large groups (Fig. 2a). As indicated in this genealogy, two
177 mutations unambiguously differentiate both clades. The map of samples clearly shows the geographical
178 separation of the two clades in North America (Fig. 2b). In the Iberian Peninsula, we found more samples
179 belonging to the western than to the eastern clade of North America (68 vs. 37 samples, respectively).
180 However, samples belonging to both clades were detected together in four regions of the north central
181 area of the Iberian Peninsula: Basque Country, La Rioja, Navarra and Avila (South of Castilla-Leon) (Fig.
182 2c).

183

184 In addition, the phylogeographic structure was also different at a finer scale in the native and invasive
185 ranges. While there was evidence of a geographical pattern for the individuals collected from the native
186 range, with individual haplotypes concentrated in a single locality or close localities (Fig. 1), many
187 haplotypes were widely dispersed in the Iberian Peninsula (Fig. 1, Fig. S1), showing that haplotypes are
188 not informative with regards to the geographical origin of any sample.

189

190 **3.2 Mitochondrial genetic diversity**

191

192 The global genetic diversity of the American mink estimated with the control region was relatively
193 low for all samples, with a value of nucleotide diversity of $\pi = 0.68\%$, whereas the nucleotide diversity
194 for North America and the Iberian Peninsula was 0.78% and 0.64%, respectively (haplotype diversity was
195 0.81, 0.92 and 0.76, respectively). However, the nucleotide diversity varied considerably when different
196 localities and populations were considered separately. Among the three areas of the native range in which

197 π was calculated separately, the population of British Columbia showed the highest nucleotide diversity
198 (0.53%) whereas New York (0.12%) and Texas (0%) registered lower values. In agreement with the
199 spatial mixing of clades in the central areas of the Iberian Peninsula, the map derived from the genetic
200 diversity values of the samples around each locality clearly showed maximum levels of genetic diversity
201 in these regions (Fig. 3). When values of several sampling localities were averaged, the Basque
202 population concentrated the highest nucleotide diversity (0.61%), followed by Avila (0.50%) and La
203 Rioja (0.25%). From these areas, genetic diversity decreased towards the remainder of the mink
204 distribution, reaching values of 0% in some areas.

205

206 **4. Discussion**

207

208 Our results show that the establishment of the invasive American mink population in the Iberian
209 Peninsula is the product of the introduction of individuals from the two main genetic lineages found in
210 western and eastern North America. In the invasive range, these lineages registered a complete lack of
211 phylogeographic structure. This was probably the result of the additive effect of multiple continuous
212 introductions, and the rapid expansion and encounters of the introduced populations. Overall, genetic
213 diversity was lower in the introduced than in the native populations, although in some introduced
214 populations diversity was higher to any one native population, at least with respect to the reduced native
215 area analysed so far. These results provide evidence for the potential of introduced populations to
216 maintain or increase the genetic diversity.

217

218 **4.1 Phylogeographic patterns**

219

220 Phylogenetic analysis revealed two large native phylogeographic lineages, which indicated a clear
221 separation between the western and eastern populations in North America. Therefore, the relationship
222 between the mtDNA haplotypes and the location from which they were sampled provides evidence of
223 phylogeographic structure among populations of American mink in their native range. Unfortunately, the
224 lack of information about sequence data of mtDNA of the American mink in North America did not allow
225 our results to be compared with those of similar studies. Other studies of plant and animal species from
226 North America have illustrated a complex phylogeographic history of the biota of this region (Shafer et
227 al. 2010; Soltis et al. 2006). The data currently available suggest that vicariance, dispersal and the
228 existence of refugia have worked in concert to produce the species' distribution and genetic divergence
229 patterns of the region (Brunsfield et al. 2001). Future research focused on filling the gaps by sampling at
230 locations between the two clades seen in this work is required to achieve a deeper understanding of the
231 evolutionary history of the American mink in its native range.

232

233 The pattern of spatial structure recorded in North America allowed for a more precise identification of
234 the origin of invasive populations in the Iberian Peninsula. Our analysis revealed that there were
235 introductions from both western and eastern regions in North America. Additionally, the distribution of
236 the invasive haplotypes in the Iberian Peninsula suggests that at least the north central area of the Iberian

237 Peninsula (populations of the Basque Country, La Rioja, Navarra and Avila) experienced multiple
238 introductions from both regions of North America, while the western and eastern Iberian populations
239 (Galicia, Catalonia and Castellon) originated from introductions from western North America, contrary to
240 what was previously believed (Palazón et al. 1997). This situation most likely occurred as a result of the
241 multiple origins of mink from different farms and the fast range expansion of the occupied area since
242 their introduction in the 1960s-1970s. It is also likely that farms received founder individuals from
243 different source populations in the native range.

244

245 The occurrence of the two main mitochondrial lineages in the north central area of the Iberian
246 Peninsula suggests that the intermixing of individuals of both lineages may potentially take place.
247 However, it will be necessary to perform analysis with multiple genomic markers in the future to
248 determine whether individuals belonging to these two lineages effectively interbreed and, if so, to
249 estimate admixture levels in these populations (Pritchard et al. 2000). It is also important to note that, due
250 to the limited information derived from the mitochondrial DNA, which only reflects the maternal lineage,
251 the existence of some admixture levels in other Iberian populations cannot be discarded until genomic
252 analyses are performed. In fact, a previous study in the Iberian Peninsula based on microsatellites showed
253 that the highest admixture levels between populations were found in the Basque Country population,
254 which in principle is consistent with our results (Lecis et al. 2008). However, since no native specimens
255 were available in that study, it was not possible to determine how the identified populations were related
256 to the source populations. This confirms that it is crucial to obtain samples from the native range to shed
257 light on the relationship between invasive and native populations and to identify potential source
258 populations.

259

260 **4.2 Genetic diversity in the native versus invasive range**

261

262 Based on mitochondrial DNA control region sequence data, our results provide evidence that the
263 introduced American mink populations in the Iberian Peninsula have overall less genetic diversity than
264 native mink populations in North America. These results are in agreement a priori with the “genetic
265 paradox” (Allendorf and Lundquist 2003; Estoup et al. 2016), which predicts that low genetic variability
266 can be expected in invasive species because of the loss of rare alleles, as a result of the founder effect and
267 population bottlenecks (Allendorf and Lundquist 2003; Grapputo et al. 2005). Nevertheless, when
268 analysed at the population level within each range, nucleotide diversity in the population of the Basque
269 Country and La Rioja in the Iberian Peninsula was found to be greater than that recorded for North
270 American populations. The pattern registered here is consistent with other studies (e.g. Edelaar et al.
271 2015; Genton et al. 2005; Roman and Darling 2007), providing another example of an invasive species
272 retaining genetic diversity. Similarly, previous studies have demonstrated that the genetic diversity of
273 invasive populations in their new range may be increased through the admixture of lineages from multiple
274 native populations (Keller and Taylor 2010; Kolbe et al. 2008). As already indicated, the pattern observed
275 in the Iberian Peninsula could be the result of multiple introductions of mink deriving from different local
276 lineages of North America, their rapid expansion, and the encounter of lineages in the north central area

277 of the Iberian Peninsula that may potentially hybridise. However, we cannot discard that cross-breeding
278 of the different North American subspecies or populations took place within farms (Belliveau et al. 1999;
279 Dunstone 1993).

280

281 High values of genetic diversity have also been registered in some mink population of Poland
282 (Zalewski et al. 2011). In this case, mean π for all feral mink analysed was 0.94%, with values for
283 different populations ranging between 0 and 1.41%. Although the mitochondrial control region fragment
284 used in that work was different to ours and a direct comparison cannot be performed, the values found in
285 some populations in Poland were much higher than in others. It is therefore likely that the same processes
286 suggested for the Iberian Peninsula (i.e. multiple introductions, rapid expansion and encounters of
287 individuals of different lineages) can be applied to other areas with invasive populations of American
288 mink.

289

290 **4.3 Implications for management**

291

292 This study provides an understanding of the genetic structure and diversity of the populations of the
293 American mink in the Iberian Peninsula, which can be used to identify key control areas. Thus, one way
294 to improve the effectiveness of management programmes of an invasive species such as the American
295 mink is by incorporating molecular-based information into these programs (Hampton et al. 2004). In this
296 regard, the current spatial mixing of individuals from two different native lineages in the north central
297 area of the Iberian Peninsula (Basque Country, La Rioja, Navarra and Avila populations) and the
298 potential interbreeding between such lineages that may take place could translate, among other outcomes,
299 into a higher fitness in these individuals (e.g. Vandewoestijne et al. 2008). In fact, previous studies have
300 found a positive correlation between genetic diversity and colonization success as measured by the
301 reproduction rate (e.g. Crawford and Whitney 2010). Interestingly, Melero et al. (2012b) found that mink
302 of the Basque Country were bigger in size than mink of the rest of the Iberian Peninsula. According to
303 Melero et al. (2012b), this phenotypic variation could be due to a mixing of individuals from different
304 origins, among other possibilities. If the link between spatial mixing of mitochondrial lineages, genome
305 admixture and higher fitness is demonstrated for some populations, then higher efforts to control and
306 eradicate the American mink should be undertaken in such populations in order to impede or minimise
307 their expansion.

308

309 Current controls of the American mink in the Iberian Peninsula focus on reducing their effects on
310 ecosystems and biodiversity (i.e., areas with local populations of threatened species, wetlands where they
311 can cause high mortalities in waterfowl and protected natural areas, etc.). In this regard, it is important to
312 consider that the areas of the Iberian Peninsula with presence of both American mink lineages coincide
313 with critical areas for the conservation of two of the most threatened mammals of the Iberian Peninsula,
314 the European mink and the Pyrenean desman. The American mink compete successfully for habitat and
315 prey with the European mink (Sidorovich et al. 1999; Sidorovich and MacDonald 2001), whose
316 population of the North of Spain and South West of France is highly isolated from the rest of the

317 distribution (Michaux et al. 2005). On the other hand, the populations of Pyrenean desman of the Central
318 System have experienced a strong reduction in the last few years and are threatened with extinction.
319 Although the reasons for this decline are not well known, predation by the American mink may have
320 contributed to it (Fernandes et al. 2008). Thus, the presence of the two native lineages of the American
321 mink in the areas with these threatened species should be taken into account in future conservation plans
322 due the putative higher expansive potential of the American mink in them.

323
324 The results of this work are in line with previous studies suggesting that mink control should also be
325 directed to minimise the contact between genetically differentiated populations (Bifulchi et al. 2010;
326 Fraser et al. 2013; Lecis et al. 2008; Robertson and Gemmell 2004), and so counteract a possible increase
327 in mink fecundity and dispersal (Melero et al. 2015; Oliver et al. 2016). The present study has shown how
328 molecular-based information obtained from the invasive range, together with that from the native range,
329 was critical to characterise the sources from which the Iberian populations of the American mink
330 originated. Further studies are now necessary to obtain more detailed information, particularly in areas
331 where admixture of lineages may potentially occur, to help prioritise efforts to control and manage this
332 invasive species.

333

334 5. Acknowledgements

335

336 This work was financially supported by the research Project CGL2014-53968-P of the "Plan Nacional
337 I+D+I del Ministerio de Economía y Competitividad" (Spain) to J.C., cofinanced with FEDER funds.
338 K.G. was funded by a "Becas Chile-CONICYT" PhD studentship. We acknowledge the following
339 institutions and administrations for samples: University of Alaska Museum (Mammal Collection),
340 Museum of Southwestern Biology (Division of Mammals), Cornell University Museum of Vertebrates
341 (Mammal Collection), Generalitat de Catalunya, Gobierno de La Rioja, Junta de Castilla y León, Xunta
342 de Galicia (Dirección Xeral de Conservación da Natureza), Generalitat Valenciana (Centro de
343 Recuperación de Fauna La Granja de El Saler) and Gobierno de Navarra. We wish to thank to all
344 technicians of administrations involved in sample collection and management, in particular: Asun Gómez,
345 Rafael Romero, Juan del Nido, Olga Alarcia, Gabriel de Pedro and Iñigo Zuberogoitia, Gloria Giralda
346 and José María Gil. We also thank Oliver Hawlitschek and Marina Querejeta for helpful comments on the
347 manuscript.

348

349 6. References

350

- 351 Allendorf FW, Lundquist LL (2003) Population Biology, Evolution, and Control of Invasive Species.
352 *Conserv Biol* 17:24–30. doi: 10.1046/j.1523-1739.2003.02365.x
- 353 Belliveau AM, Farid A, O'Connell M, Wright JM (1999) Assessment of genetic variability in captive and
354 wild American mink (*Mustela vison*) using microsatellite markers. *Can J Anim Sci* 79:7–16. doi:
355 10.4141/A97-115
- 356 Bifulchi A, Picard D, Lemaire C, Cormier JP, Pagano A (2010) Evidence of admixture between

357 differentiated genetic pools at a regional scale in an invasive carnivore. *Conserv Genet* 11:1–9. doi:
358 10.1007/s10592-008-9780-1

359 Blackburn TM, Cassey P, Lockwood JL (2009) The role of species traits in the establishment success of
360 exotic birds. *Glob Chang Biol* 15:2852–2860. doi: 10.1111/j.1365-2486.2008.01841.x

361 Bonesi L, Chanin P, Macdonald D (2004) Competition between Eurasian otter *Lutra lutra* and American
362 mink *Mustela vison* probed by niche shift. *Oikos* 106:19–26. doi: 10.1111/j.0030-
363 1299.2004.12763.x

364 Bonesi L, Palazón S (2007) The American mink in Europe: Status, impacts, and control. *Biol Conserv*
365 134:470–483. doi: 10.1016/j.biocon.2006.09.006

366 Brunsfeld SJ, Sullivan J, Soltis DE, Soltis PS (2001) Comparative phylogeography of north- western
367 North America: a synthesis. In: in J. Silvertown and J. Antonovics [Eds.], *Integrating ecology and*
368 *evolution in a spatial context*, Blackwell Science, Oxford, UK. pp 319 – 339

369 Crawford KM, Whitney KD (2010) Population genetic diversity influences colonization success. *Mol*
370 *Ecol* 19:1253–1263. doi: 10.1111/j.1365-294X.2010.04550.x

371 Dlugosch KM, Parker IM (2008) Founding events in species invasions: Genetic variation, adaptive
372 evolution, and the role of multiple introductions. *Mol Ecol* 17:431–449. doi: 10.1111/j.1365-
373 294X.2007.03538.x

374 Dlugosch KM, Anderson SR, Braasch J, Cang FA, Gillette HD (2015) The devil is in the details: genetic
375 variation in introduced populations and its contributions to invasion. *Mol Ecol* 24:2095–2111. doi:
376 10.1111/mec.13183

377 Dunstone N (1993). *The Mink*. T&AD Poyser Ltd., London

378 Edelaar P, Roques S, Hobson EA, Gonçalves Da Silva A, Avery ML, Russello MA, Senar JC, Wright TF,
379 Carrete M, Tella JL (2015) Shared genetic diversity across the global invasive range of the monk
380 parakeet suggests a common restricted geographic origin and the possibility of convergent
381 selection. *Mol Ecol* 24:2164–2176. doi: 10.1111/mec.13157

382 Estoup A, Ravign V, Hufbauer R, Vitalis R, Gautier M, Facon B (2016) Is There A Genetic Paradox of
383 Biological Invasion? *Annu Rev Ecol Evol Syst* 47:51–72. doi: 10.1146/annurev-ecolsys-121415

384 Fernandes M, Herrero J, Aulagnier S, Amori G (2008) *Galemys pyrenaicus*. The IUCN Red List of
385 Threatened Species: e.T8826A12934876.
386 <http://dx.doi.org/10.2305/IUCN.UK.2008.RLTS.T8826A12934876.en>

387 Fraser EJ, Macdonald DW, Oliver MK, Piertney S, Lambin X (2013) Using population genetic structure
388 of an invasive mammal to target control efforts - An example of the American mink in Scotland.
389 *Biol Conserv* 167:35–42. doi: 10.1016/j.biocon.2013.07.011

390 Genton BJ, Shykoff JA, Giraud T (2005) High genetic diversity in French invasive populations of
391 common ragweed, *Ambrosia artemisiifolia*, as a result of multiple sources of introduction. *Mol Ecol*
392 14:4275–4285. doi: 10.1111/j.1365-294X.2005.02750.x

393 Grapputo A, Boman S, Lindström L, Lyytinen A, Mappes J (2005) The voyage of an invasive species
394 across continents: Genetic diversity of North American and European Colorado potato beetle
395 populations. *Mol Ecol* 14:4207–4219. doi: 10.1111/j.1365-294X.2005.02740.x

396 Hampton JO, Spencer PBS, Alpers DL, Twigg LE, Woolnough AP, Doust J, Higgs T, Pluske J (2004)

397 Molecular techniques, wildlife management and the importance of genetic population structure and
398 dispersal: a case study with feral pigs. *J Appl Ecol* 41:735–743. doi: 10.1111/j.0021-
399 8901.2004.00936.x

400 Harrington LA, Harrington AL, Moorhouse T, Gelling M, Bonesi L, Macdonald DW (2009) American
401 mink control on inland rivers in southern England: An experimental test of a model strategy. *Biol*
402 *Conserv* 142:839–849. doi: 10.1016/j.biocon.2008.12.012

403 Katoh K, Misawa K, Kuma K, Miyata T (2002) MAFFT: a novel method for rapid multiple sequence
404 alignment based on fast Fourier transform. *Nucleic Acids Res* 30:3059–3066. doi:
405 10.1093/nar/gkf436

406 Keller SR, Taylor DR (2010) Genomic admixture increases fitness during a biological invasion. *J Evol*
407 *Biol* 23:1720–1731. doi: 10.1111/j.1420-9101.2010.02037.x

408 Kolbe JJ, Glor RE, Rodríguez Schettino L, Lara AC, Larson A, Losos JB (2004) Genetic variation
409 increases during biological invasion by a Cuban lizard. *Nature* 431:177–181. doi:
410 10.1038/nature02807

411 Kolbe JJJ, Larson A, Losos JBB, de Queiroz K (2008) Admixture determines genetic diversity and
412 population differentiation in the biological invasion of a lizard species. *Biol Lett* 4:434–437. doi:
413 10.1098/rsbl.2008.0205

414 Lecis R, Ferrando A, Ruiz-Olmo J, Mañas S, Domingo-Roura X (2008) Population genetic structure and
415 distribution of introduced American mink (*Mustela vison*) in Spain, based on microsatellite
416 variation. *Conserv Genet* 9:1149–1161. doi: 10.1007/s10592-007-9428-6

417 Macdonald DW, Harrington LA (2003) The American mink: The triumph and tragedy of adaptation out
418 of context. *N.Z. J. Zool.* 30(4):421–441. doi: <http://dx.doi.org/10.1080/03014223.2003.9518350>

419 Maran T, Skumatov D, Gomez A, Pödra M, Abramov AV, Dinets V (2016) *Mustela lutreola*. The IUCN
420 Red List of Threatened Species 2016: e.T14018A45199861.
421 <http://dx.doi.org/10.2305/IUCN.UK.2016-1.RLTS.T14018A45199861.en>

422 Melero Y, Plaza M, Santulli G, Saavedra D, Gosàlbez J, Ruiz-Olmo J, Palazón S (2012a) Evaluating the
423 effect of American mink, an alien invasive species, on the abundance of a native community: is
424 coexistence possible? *Biodivers Conserv* 21:1795–1809. doi: 10.1007/s10531-012-0277-3

425 Melero Y, Santulli G, Gómez A, Gosàlbez J, Rodriguez-Refojos C, Palazón S (2012b) Morphological
426 variation of introduced species: The case of American mink (*Neovison vison*) in Spain. *Mamm Biol*
427 77:345–350. doi: 10.1016/j.mambio.2012.02.001

428 Melero Y, Robinson E, Lambin X (2015) Density- and age-dependent reproduction partially compensates
429 culling efforts of invasive non-native American mink. *Biol Invasions* 17(9): 2645-2657. doi:
430 10.1007/s10530-015-0902-7

431 Michaux JR, Hardy OJ, Justy F, Fournier P, Kranz A, Cabria M, Davison A, Rosoux R, Libois R (2005)
432 Conservation genetics and population history of the threatened European mink *Mustela lutreola*,
433 with an emphasis on the west European population. *Mol Ecol* 14:2373–2388. doi: 10.1111/j.1365-
434 294X.2005.02597.x

435 Novak S, Mack R (1993) Genetic variation in *Bromus tectorum* (Poaceae): comparison between native
436 and introduced populations. *Heredity (Edinb)* 71:167–176. doi: 10.1038/hdy.1993.121

437 Oliver MK, Piernrey SB, Zalewski A, Melero Y, Lambin X (2016) The compensatory potential of
438 increased immigration following intensive American mink population control is diluted by male-
439 biased dispersal. *Biol Invasions* 18:3047–3061. doi: 10.1007/s10530-016-1199-x

440 Palazón S, Ruiz-Olmo J, Bueno F, Bueno F, Jordan G, Palomero G, Palomero G, Munilla I, Romero R,
441 Gimenez YJ (1997) El visón americano en España. In: El visón europeo y el visón americano en
442 España. (Eds.) Palazón, S. and Ruiz-Olmo, J. Colección Técnica, Ministerio de Medio Ambiente.
443 Madrid. pp 107–114

444 Paradis E (2010) PEGAS: An R package for population genetics with an integrated-modular approach.
445 *Bioinformatics* 26:419–420. doi: 10.1093/bioinformatics/btp696

446 Pritchard JK, Stephens M, Donnelly P (2000) Inference of population structure using multilocus genotype
447 data. *Genetics* 155:945–959. doi: 10.1111/j.1471-8286.2007.01758.x

448 Reid F, Schiaffini M, Schipper J (2016) *Neovison vison*. The IUCN Red List of Threatened Species 2016:
449 e.T41661A45214988. <http://dx.doi.org/10.2305/IUCN.UK.2016-1.RLTS.T41661A45214988.en>.

450 Robertson BC, Gemmell NJ (2004) Defining eradication units to control invasive pests. *J Appl Ecol*
451 41:1042–1048. doi: 10.1111/j.0021-8901.2004.00984.x

452 Roman J, Darling JA (2007) Paradox lost: genetic diversity and the success of aquatic invasions. *Trends*
453 *Ecol Evol* 22:454–464. doi: 10.1016/j.tree.2007.07.002

454 Ruiz-Olmo J, Munilla I, Romero R (1997) Distribution, Status and Colonization of the American Mink
455 *Mustela vison* in Spain. *J Wildl Res* 2:30–36.

456 Salzburger W, Ewing GB, Von Haeseler A (2011) The performance of phylogenetic algorithms in
457 estimating haplotype genealogies with migration. *Mol Ecol* 20:1952–1963. doi: 10.1111/j.1365-
458 294X.2011.05066.x

459 Santulli G (2014) Occupancy, abundance, potential distribucion and spatial competition of the critically
460 endangered european mink (*Mustela lutreola*) and the invasive non-native american mink
461 (*Neovison vison*) in the Iberian Peninsula. Dissertation, Universitat de Barcelona.

462 Shafer ABA, Cullingham CI, Côté SD, Coltman DW (2010) Of glaciers and refugia: a decade of study
463 sheds new light on the phylogeography of northwestern North America. *Mol Ecol* 19:4589–4621.
464 doi: 10.1111/j.1365-294X.2010.04828.x

465 Sidorovich V, Kruuk H, Macdonald D (1999) Body size, and interactions between European and
466 American mink (*Mustela lutreola* and *M. vison*) in Eastern Europe. *J Zool* 248:521–527. doi:
467 10.1111/j.1469-7998.1999.tb01051.x

468 Sidorovich V, MacDonald DW (2001) Density dynamics and changes in habitat use by the European
469 mink and other mustelids in connection with the American mink expansion in Belarus. *Netherlands*
470 *J Zool* 51:107–126. doi: 10.1163/156854201750210878

471 Sol D, Maspons J, Vall-Llosera M, Bartomeus I, García-Peña GE, Piñol J, Freckleton RP (2012)
472 Unraveling the life history of successful invaders. *Science* 337:580–583. doi:
473 10.1126/science.1221523

474 Soltis DE, Morris AB, McLachlan JS, Manos PS, Soltis PS (2006) Comparative phylogeography of
475 unglaciated eastern North America. *Mol Ecol* 15(14):4261–4293. doi:10.1111/j.1365-
476 294X.2006.03061.x

477 Stamatakis A (2014) RAxML version 8: A tool for phylogenetic analysis and post-analysis of large
478 phylogenies. *Bioinformatics* 30:1312–1313. doi: 10.1093/bioinformatics/btu033
479 Suarez A V., Tsutsui ND (2008) The evolutionary consequences of biological invasions. *Mol Ecol*
480 17:351–360. doi: 10.1111/j.1365-294X.2007.03456.x
481 Uller T, Leimu R (2011) Founder events predict changes in genetic diversity during human-mediated
482 range expansions. *Glob Chang Biol* 17:3478–3485. doi: 10.1111/j.1365-2486.2011.02509.x
483 Vandewoestijne S, Schtickzelle N, Baguette M (2008) Positive correlation between genetic diversity and
484 fitness in a large, well-connected metapopulation. *BMC Biol* 6:46–55. doi: 10.1186/1741-7007-6-
485 46
486 Zalewski A, Piertney SB, Zalewska H, Lambin X (2009) Landscape barriers reduce gene flow in an
487 invasive carnivore: geographical and local genetic structure of American mink in Scotland. *Mol*
488 *Ecol* 18:1601–1615. doi: 10.1111/j.1365-294X.2009.04131.x
489 Zalewski A, Michalska-Parda A, Bartoszewicz M, Kozakiewicz M, Brzeziński M (2010) Multiple
490 introductions determine the genetic structure of an invasive species population: American mink
491 *Neovison vison* in Poland. *Biol Conserv* 143:1355–1363. doi: 10.1016/j.biocon.2010.03.009
492 Zalewski A, Michalska-Parda A, Ratkiewicz M, Kozakiewicz M, Bartoszewicz M, Brzeziński M (2011)
493 High mitochondrial DNA diversity of an introduced alien carnivore: Comparison of feral and ranch
494 American mink *Neovison vison* in Poland. *Divers Distrib* 17:757–768. doi: 10.1111/j.1472-
495 4642.2011.00767.x
496 Zalewski A, Zalewska H, Lunneryd SG, André C, Mikusiński G (2016) Reduced genetic diversity and
497 increased structure in American mink on the Swedish coast following invasive species control.
498 *PLoS One* 11(6):e0157972. doi: 10.1371/journal.pone.0157972
499
500
501
502
503
504
505
506
507
508

509 **Fig. 1** Maximum-likelihood phylogenetic tree derived from the partial mitochondrial control region
510 sequences for 126 American mink, with eastern and western clades indicated. The number of individuals
511 per locality is shown in brackets. Sequences from the native range are underlined. The scale indicates
512 number of substitutions per position.

513

514

515 **Fig. 2** Phylogeographic analysis of the mitochondrial control region sequences of *Neovison vison*. (A)
516 Haplotype genealogy based on the maximum-likelihood tree. Circle sizes are proportional to the
517 haplotype frequency and small filled circles are inferred haplotypes not present in the sample. (B, C) Map
518 of native (B) and invasive (C) ranges showing the 126 samples of *N. vison* used in this study. Red and
519 blue colours represent the eastern and western clades, respectively. Grey areas represent the original
520 distribution of *N. vison* downloaded from the IUCN Red List of Threatened Species website (Reid et al.
521 2016) (B) and the occupancy area of mink in the Peninsula (MAGRAMA, 2013) (C). AL: Alaska, BC:
522 British Columbia, MT: Montana, WY: Wyoming, TX: Texas, NY: New York, GA: Galicia, CL: Castilla-
523 Leon, AV: Avila, LR: La Rioja, BC: Basque Country, NAV: Navarra, CAS: Castellon, CAT: Catalonia

524

525

526 **Fig. 3** Map plotting genetic diversity (π) in different sampling localities of *Neovison vison* in the Iberian
527 Peninsula. Colours of the sample points indicate the nucleotide diversity according to the provided scale.

528

529

530

531

532

533

534

535

536

537

538

539

540

541

542

543

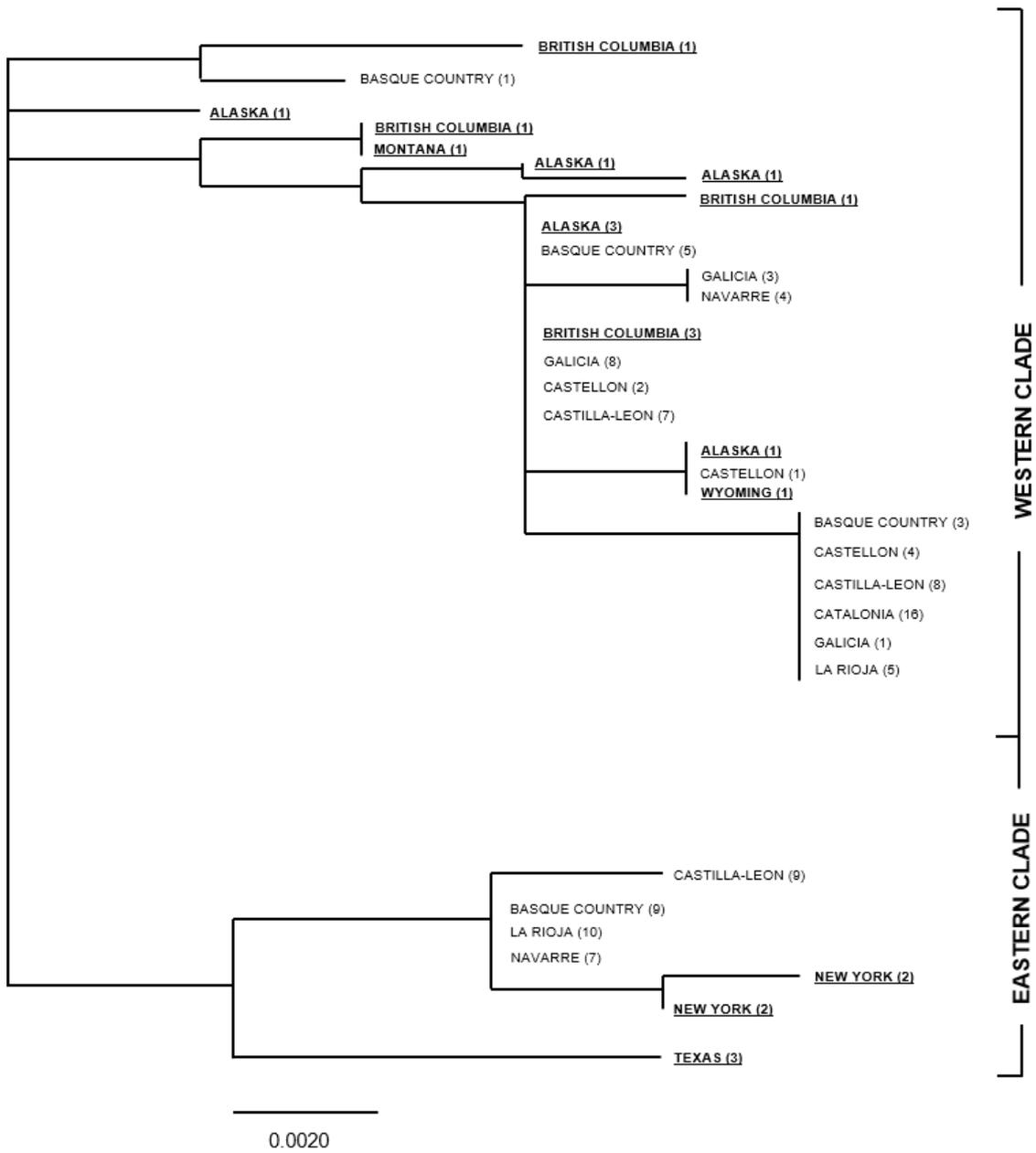
544

545

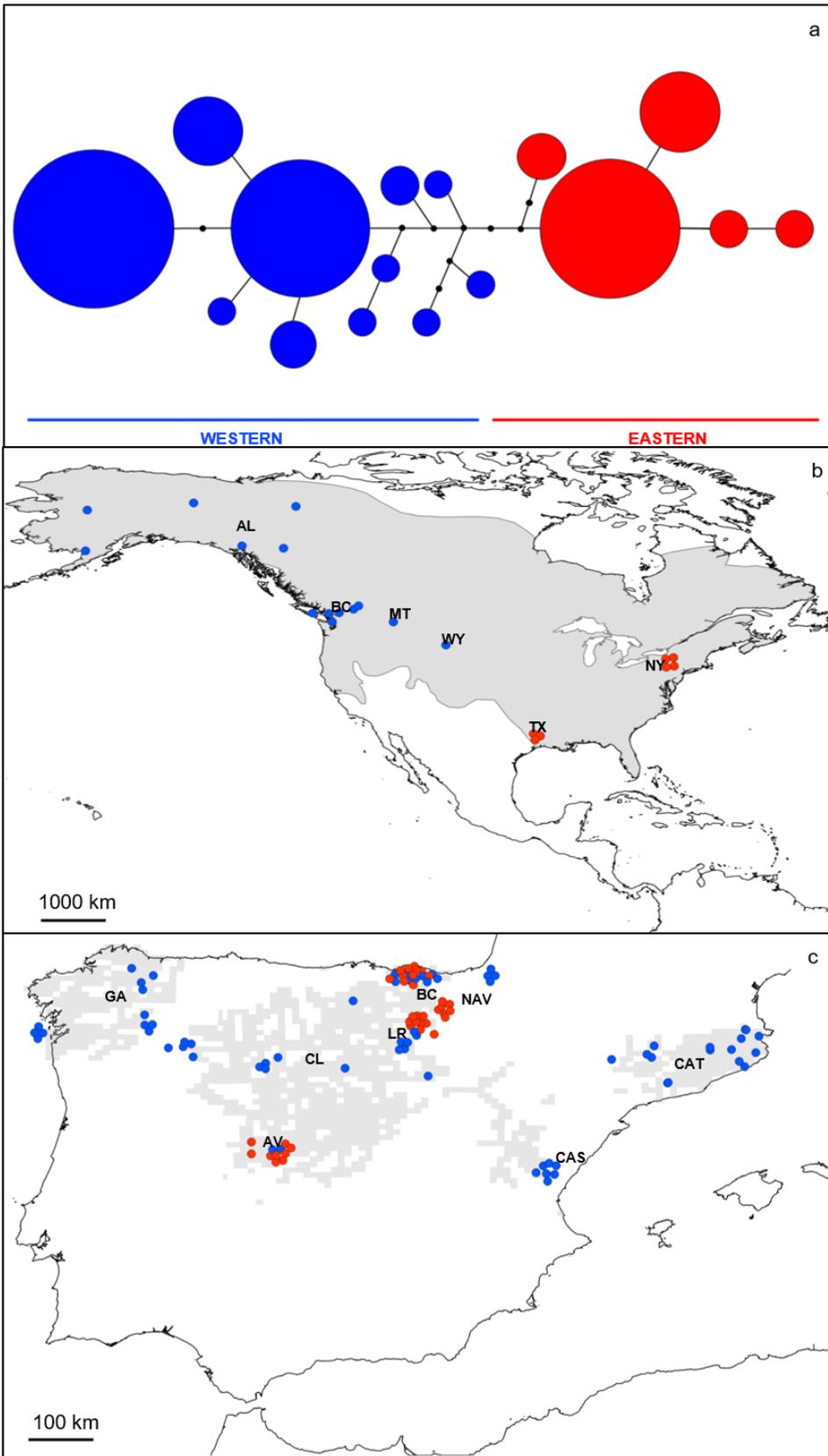
546

547

548

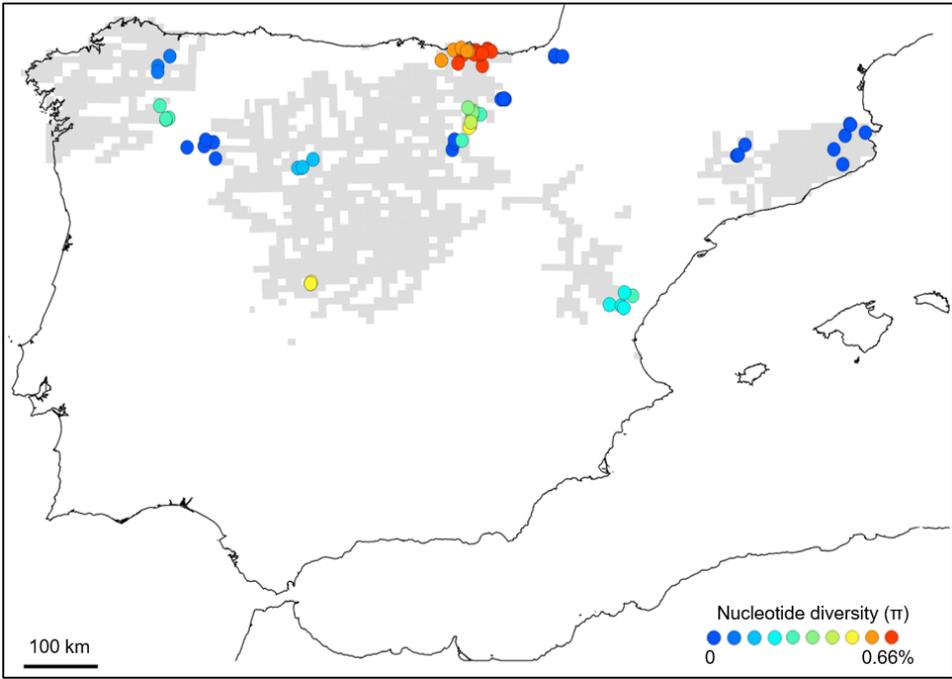


549
 550 Fig. 1
 551
 552
 553
 554
 555
 556
 557
 558
 559
 560



561
 562
 563
 564
 565

Fig. 2



566
567
568

Fig. 3