

*Spatial mixing of mitochondrial lineages and greater genetic diversity in some invasive populations of the American mink (Neovison vison) compared to native populations*

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1 **Spatial mixing of mitochondrial lineages and greater genetic diversity in some invasive populations**  
2 **of the American mink (*Neovison vison*) compared to native populations**

3

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15

16

17 **Abstract**

18

19           The genetic characteristics of introduced populations have a relevant impact on their ability to  
20 establish and spread. The American mink (*Neovison vison*), native to North America, is an important  
21 invasive species in the Iberian Peninsula. Here, we used mitochondrial DNA sequences data to investigate  
22 the genetic diversity and phylogeographic structure of invasive vs. native populations of this species. We  
23 also evaluated whether genetic diversity in invasive populations could be explained by the genetic  
24 characteristics of the native sources from which they derived. Phylogenetic analysis revealed two major  
25 lineages in the native range, which indicated a clear separation between western and eastern populations.  
26 On the contrary, we found no evidence of genetic structure in the invasive range. This was probably the  
27 result of the diverse origins of the released specimens and the rapid expansion and encounters of the  
28 introduced populations. We detected spatial mixing of both North American lineages in several sampling  
29 localities of the north central area of the Iberian Peninsula, giving rise to high levels of genetic diversity  
30 in some areas compared to North American populations. This could potentially lead to higher fitness of  
31 these individuals and thus increase the population viability and invasiveness of this species. These results  
32 point to the need to better study the populations in which lineages mix and, if necessary, intensify control  
33 efforts in them.

34

35 **Keywords:** American mink, mitochondrial genes, genetic diversity, native range, Iberian Peninsula.

## 36 1. Introduction

37  
38 Identifying the biological attributes of successful invaders is among the most pressing questions yet to  
39 be answered (Sol et al. 2012). Although there is evidence that species differ in their invasion potential,  
40 controversy exists about the characteristics that differentiate successful and unsuccessful invaders  
41 (Blackburn et al. 2009). Other studies of introduced species show that genetics might play an integral role  
42 in the success of an invasion, and they suggest that this role is determined by the quality of the genetic  
43 variation introduced, rather than the total quantity of genetic variation (see Dlugosch et al. 2015). The  
44 genetic diversity of founding populations, along with potential increase in genetic diversity from  
45 subsequent introductions, may be important to the successful establishment and spread of introduced  
46 species (Suarez and Tsutsui 2008; Sol et al. 2012). Nevertheless, there is no consensus on whether  
47 invaders stemming from multiple native origins are more successful than those from single populations,  
48 or whether demographic bottlenecks may limit the invasion success of a species (Edelaar et al. 2015;  
49 Estoup 2016).

50  
51 Overall, introduced populations lose genetic variation relative to their source populations (Dlugosch  
52 and Parker 2008). However, numerous introduction events and a greater number of source populations  
53 may buffer against such genetic losses (Kolbe et al. 2004; Dlugosch and Parker 2008; Uller and Leimu  
54 2011). In fact, in some cases successful invaders show large increases in genetic diversity in the  
55 introduced range, apparently caused by mixing of populations from different source regions (Genton et al.  
56 2005; Novak and Mack 1993). Therefore, it is crucial to obtain information on the phylogeographic  
57 structure and genetic diversity throughout the native and introduced ranges to understand the factors  
58 affecting genetic diversity during invasion (e.g. multiple introduction events or admixture levels among  
59 populations) and the interaction between genetic diversity and invasive potential (Edelaar et al. 2015).

60  
61 The American mink (*Neovison vison*) is a successful invader with a native range restricted to North  
62 America and with invasive populations distributed across Europe, Asia and South America (Bonesi and  
63 Palazón 2007; Dunstone 1993). In Europe, this species was introduced for commercial fur farming during  
64 the 20<sup>th</sup> century. Populations established by escaped individuals are now naturalised in fresh water and  
65 coastal ecosystems (Bonesi and Palazón 2007), with detrimental consequences for the abundance of many  
66 native species of prey and competitors of conservation importance (Bonesi et al. 2004; Harrington et al.  
67 2009; Macdonald and Harrington 2003; Melero et al. 2012a), including critically endangered species such  
68 as the European mink (*Mustela lutreola*) (Maran et al. 2016) and the Pyrenean desman (*Galemys*  
69 *pyrenaicus*) (Fernandes et al. 2008). Fur farms were established in many countries and harboured mink  
70 stocks from different parts of their native range. Because of the large scale and long-term nature of this  
71 farming, the number of escapees, which started to breed in the wild and became the founder individuals  
72 of feral populations, was very high (Bonesi and Palazón 2007). In Spain, mink farming started at the end  
73 of the 1950s and the number of farms reached a peak in the 1980s with about 400 farms concentrated  
74 mainly in the North West (Galicia) (Ruiz-Olmo et al. 1997). Today, feral mink populations are widely  
75 distributed across almost all of the north and centre of the Iberian Peninsula. During the last few decades,  
76 these populations experienced a rapid expansion and increased the number of nuclei in almost every

77 colonised area (Ruiz-Olmo et al. 1997; Santulli 2014). Although the number of mink farms in Spain was  
78 considerably reduced at the end of the 1980s, mink escapes from the last active farms are still ongoing  
79 (Bonesi and Palazón 2007).

80

81 In Europe, there is genetic information on feral American mink populations from several countries.  
82 Previous studies in Spain using microsatellite data determined that American mink populations from  
83 disjunct geographical areas appeared differentiated and not yet connected by gene flow (Lecis et al.  
84 2008). In Scotland, Zalewski et al. (2009) and Fraser et al. (2013) found that landscape features restricted  
85 gene flow within this species. Other studies in Poland found that multiple introductions determine the  
86 genetic structure of American mink populations (Zalewski et al. 2010) and that these populations exhibit  
87 high mitochondrial genetic diversity (Zalewski et al. 2011). A study in France using microsatellite data  
88 suggested recent admixture between populations that had recently encountered each other (Bifulchi et al.  
89 2010). Similarly, in Sweden it was found, also using microsatellite data, that genetic diversity decreased  
90 over six consecutive years of mink culling as part of a population control programme (Zalewski et al.  
91 2016).

92

93 Despite the importance of understanding the genetic diversity within the native populations for  
94 investigating the evolutionary dynamics of invasions, no phylogeographic study has been performed so  
95 far in the North American native range of the American mink. Therefore, many genetic studies performed  
96 thus far are difficult to interpret. We used mitochondrial DNA (mtDNA) sequence data from populations  
97 in the native range to elucidate whether invasive populations in the Iberian Peninsula have single or  
98 multiple origins from the North American sources. We aimed to determine the phylogeographic structure  
99 of native and invasive populations, and to evaluate whether mitochondrial genetic diversity observed in  
100 established invasive populations can be explained by genetic characteristics of native populations. We  
101 examined whether the genetic variability of the Iberian population represented only a small part of that in  
102 its native range, as expected for many invasive species or, due to the introduction history and particular  
103 origin of these populations, this genetic variability was larger. Overall, the study aims to provide  
104 recommendations for mink control in the Iberian Peninsula with respect to determining key control areas.

105

## 106 **2. Materials and methods**

107

### 108 **2.1 Sample collection**

109

110 We used hair and tissue samples from invasive populations provided by different Autonomous  
111 Communities of Spain. These samples were collected during population control operations conducted by  
112 the local administrations between 2007 and 2016 as part of eradication programs and were conserved in  
113 tubes containing absolute ethanol (ESM Appendix: Table S1). A total of 105 samples from Catalonia  
114 (16), La Rioja (15), Galicia (12), Basque Country (20), Navarra (11), Castellon (7) and Castilla-Leon (24)  
115 were included in the study.

116

117 Moreover, tissue samples of 21 specimens from North America (ESM Appendix: Table S2) were  
118 obtained from different biological collections: University of Alaska Museum (12), Museum of  
119 Southwester Biology (5) and Cornell University Museum of Vertebrates (4).

120

## 121 **2.2 DNA extraction, PCR and sequencing of mitochondrial DNA**

122

123 We extracted DNA using the QIAGEN DNeasy Blood and Tissue Kit, following the manufacturer's  
124 instructions, in a final elution volume of 100  $\mu$ l. These extractions were carried out in a separated UV-  
125 irradiated area with dedicated equipment. Samples were incubated in a lysis buffer with proteinase K at  
126 56°C overnight to ensure maximum tissue lysis.

127

128 For each sample, we amplified a 573 bp fragment of the 5' distal part of the mitochondrial DNA  
129 control region (D-loop). For this purpose, we designed novel primers using complete mitochondrial  
130 sequences of several mustelids: Neovison\_tRNA\_pro (5'-ATC AGC ACC CAA AGC TGA CAT TCT  
131 A-3') and Neovison\_Dloop\_556r (5'-TGT GTA TGT CCT GTG ACC ATT GAC T-3'). PCR reactions  
132 were set up in a dedicated PCR clean room that is physically separated from post-PCR working areas and  
133 regularly decontaminated by UV-irradiation. PCR reactions were performed in a final volume of 25  $\mu$ l,  
134 containing 2–4  $\mu$ l of genomic DNA, 1  $\mu$ M of each primer, 0.2 mM dNTPs, 0.75 units of Promega GoTaq  
135 DNA polymerase and 17.5  $\mu$ M of bovine serum albumin under the following conditions: an initial  
136 denaturation of 2 min at 95°C, followed by 35 cycles of denaturation (30 s at 95°C), annealing (30 s at  
137 54°C) and extension (30 s at 72°C). PCR products were visualised by electrophoresis in a 1% agarose  
138 SYBR-Safe (Invitrogen) stained gel.

139

140 PCR products were purified using ExoSAP-It (Affymetrix), and both strands were sequenced using  
141 the original PCR primers at Macrogen Europe (Amsterdam, The Netherlands). Sequences were inspected,  
142 trimmed and assembled using Geneious Pro (Biomatters Ltd.). All sequences obtained in this study were  
143 deposited in European Nucleotide Archive/GenBank under accession numbers LT854061-LT854186.

144

## 145 **2.3 Phylogenetic and genetic diversity analyses**

146

147 Consensus sequences were aligned using MAFFT version 7.130 (Katoh et al. 2002) with default  
148 settings. Due to the presence of several gaps and ambiguous positions in a very specific region of the  
149 alignment, we manually removed positions 92 to 110, thus leaving 554 bp in the alignment.

150

151 A maximum likelihood phylogenetic tree was calculated with RAxML version 8.0 (Stamatakis 2014)  
152 using a general time reversible substitution model and rate heterogeneity modelled with a gamma  
153 distribution plus a proportion of invariable sites, as recommended by the program. From this tree, a  
154 haplotype genealogy was generated using Haploviewer 1.0 (Salzburger et al. 2011).

155

156 The nucleotide diversity ( $\pi$ ) was estimated using the R package PEGAS (Population and Evolutionary

157 Genetics Analysis System, Paradis 2010). To analyse the variation in nucleotide diversity across the  
158 Iberian Peninsula, we used a visualization method that does not require the pre-definition of populations.  
159 The nucleotide diversity was estimated at each sampling location by using all samples present within 30  
160 km of the location. This allowed the estimation of genetic diversity from a sufficient number of samples  
161 at each point, yet the resolution was sufficient to distinguish regional differences in genetic diversity.  
162 QGIS (<http://qgis.osgeo.org>) was used to plot the nucleotide diversity in the map. In the case of the North  
163 American native range, it was not possible to apply this analysis due to the limited number of samples.  
164 Therefore, we selected three areas with a sufficient number of samples: North West (British Columbia,  
165 six samples), South East (Texas, three samples) and East (New York, four samples), in order to estimate  
166 the nucleotide diversity in these specific areas.

167

### 168 **3. Results**

169

#### 170 **3.1 Mitochondrial phylogeographic analysis**

171

172 A total of 16 different haplotypes of the mitochondrial control region were detected in 126 American  
173 mink, 11 of which were recorded in North America and 7 in the Iberian Peninsula (ESM Appendix: Fig.  
174 S1). The maximum likelihood phylogenetic tree showed two large clades found throughout western and  
175 eastern North America, respectively (Fig. 1). The haplotype genealogy reconstructed from the maximum  
176 likelihood tree revealed the same two large groups (Fig. 2a). As indicated in this genealogy, two  
177 mutations unambiguously differentiate both clades. The map of samples clearly shows the geographical  
178 separation of the two clades in North America (Fig. 2b). In the Iberian Peninsula, we found more samples  
179 belonging to the western than to the eastern clade of North America (68 vs. 37 samples, respectively).  
180 However, samples belonging to both clades were detected together in four regions of the north central  
181 area of the Iberian Peninsula: Basque Country, La Rioja, Navarra and Avila (South of Castilla-Leon) (Fig.  
182 2c).

183

184 In addition, the phylogeographic structure was also different at a finer scale in the native and invasive  
185 ranges. While there was evidence of a geographical pattern for the individuals collected from the native  
186 range, with individual haplotypes concentrated in a single locality or close localities (Fig. 1), many  
187 haplotypes were widely dispersed in the Iberian Peninsula (Fig. 1, Fig. S1), showing that haplotypes are  
188 not informative with regards to the geographical origin of any sample.

189

#### 190 **3.2 Mitochondrial genetic diversity**

191

192 The global genetic diversity of the American mink estimated with the control region was relatively  
193 low for all samples, with a value of nucleotide diversity of  $\pi = 0.68\%$ , whereas the nucleotide diversity  
194 for North America and the Iberian Peninsula was 0.78% and 0.64%, respectively (haplotype diversity was  
195 0.81, 0.92 and 0.76, respectively). However, the nucleotide diversity varied considerably when different  
196 localities and populations were considered separately. Among the three areas of the native range in which



197  $\pi$  was calculated separately, the population of British Columbia showed the highest nucleotide diversity  
198 (0.53%) whereas New York (0.12%) and Texas (0%) registered lower values. In agreement with the  
199 spatial mixing of clades in the central areas of the Iberian Peninsula, the map derived from the genetic  
200 diversity values of the samples around each locality clearly showed maximum levels of genetic diversity  
201 in these regions (Fig. 3). When values of several sampling localities were averaged, the Basque  
202 population concentrated the highest nucleotide diversity (0.61%), followed by Avila (0.50%) and La  
203 Rioja (0.25%). From these areas, genetic diversity decreased towards the remainder of the mink  
204 distribution, reaching values of 0% in some areas.

205

#### 206 **4. Discussion**

207

208 Our results show that the establishment of the invasive American mink population in the Iberian  
209 Peninsula is the product of the introduction of individuals from the two main genetic lineages found in  
210 western and eastern North America. In the invasive range, these lineages registered a complete lack of  
211 phylogeographic structure. This was probably the result of the additive effect of multiple continuous  
212 introductions, and the rapid expansion and encounters of the introduced populations. Overall, genetic  
213 diversity was lower in the introduced than in the native populations, although in some introduced  
214 populations diversity was higher to any one native population, at least with respect to the reduced native  
215 area analysed so far. These results provide evidence for the potential of introduced populations to  
216 maintain or increase the genetic diversity.

217

#### 218 **4.1 Phylogeographic patterns**

219

220 Phylogenetic analysis revealed two large native phylogeographic lineages, which indicated a clear  
221 separation between the western and eastern populations in North America. Therefore, the relationship  
222 between the mtDNA haplotypes and the location from which they were sampled provides evidence of  
223 phylogeographic structure among populations of American mink in their native range. Unfortunately, the  
224 lack of information about sequence data of mtDNA of the American mink in North America did not allow  
225 our results to be compared with those of similar studies. Other studies of plant and animal species from  
226 North America have illustrated a complex phylogeographic history of the biota of this region (Shafer et  
227 al. 2010; Soltis et al. 2006). The data currently available suggest that vicariance, dispersal and the  
228 existence of refugia have worked in concert to produce the species' distribution and genetic divergence  
229 patterns of the region (Brunsfield et al. 2001). Future research focused on filling the gaps by sampling at  
230 locations between the two clades seen in this work is required to achieve a deeper understanding of the  
231 evolutionary history of the American mink in its native range.

232

233 The pattern of spatial structure recorded in North America allowed for a more precise identification of  
234 the origin of invasive populations in the Iberian Peninsula. Our analysis revealed that there were  
235 introductions from both western and eastern regions in North America. Additionally, the distribution of  
236 the invasive haplotypes in the Iberian Peninsula suggests that at least the north central area of the Iberian

237 Peninsula (populations of the Basque Country, La Rioja, Navarra and Avila) experienced multiple  
238 introductions from both regions of North America, while the western and eastern Iberian populations  
239 (Galicia, Catalonia and Castellon) originated from introductions from western North America, contrary to  
240 what was previously believed (Palazón et al. 1997). This situation most likely occurred as a result of the  
241 multiple origins of mink from different farms and the fast range expansion of the occupied area since  
242 their introduction in the 1960s-1970s. It is also likely that farms received founder individuals from  
243 different source populations in the native range.

244

245 The occurrence of the two main mitochondrial lineages in the north central area of the Iberian  
246 Peninsula suggests that the intermixing of individuals of both lineages may potentially take place.  
247 However, it will be necessary to perform analysis with multiple genomic markers in the future to  
248 determine whether individuals belonging to these two lineages effectively interbreed and, if so, to  
249 estimate admixture levels in these populations (Pritchard et al. 2000). It is also important to note that, due  
250 to the limited information derived from the mitochondrial DNA, which only reflects the maternal lineage,  
251 the existence of some admixture levels in other Iberian populations cannot be discarded until genomic  
252 analyses are performed. In fact, a previous study in the Iberian Peninsula based on microsatellites showed  
253 that the highest admixture levels between populations were found in the Basque Country population,  
254 which in principle is consistent with our results (Lecis et al. 2008). However, since no native specimens  
255 were available in that study, it was not possible to determine how the identified populations were related  
256 to the source populations. This confirms that it is crucial to obtain samples from the native range to shed  
257 light on the relationship between invasive and native populations and to identify potential source  
258 populations.

259

#### 260 **4.2 Genetic diversity in the native versus invasive range**

261

262 Based on mitochondrial DNA control region sequence data, our results provide evidence that the  
263 introduced American mink populations in the Iberian Peninsula have overall less genetic diversity than  
264 native mink populations in North America. These results are in agreement a priori with the “genetic  
265 paradox” (Allendorf and Lundquist 2003; Estoup et al. 2016), which predicts that low genetic variability  
266 can be expected in invasive species because of the loss of rare alleles, as a result of the founder effect and  
267 population bottlenecks (Allendorf and Lundquist 2003; Grapputo et al. 2005). Nevertheless, when  
268 analysed at the population level within each range, nucleotide diversity in the population of the Basque  
269 Country and La Rioja in the Iberian Peninsula was found to be greater than that recorded for North  
270 American populations. The pattern registered here is consistent with other studies (e.g. Edelaar et al.  
271 2015; Genton et al. 2005; Roman and Darling 2007), providing another example of an invasive species  
272 retaining genetic diversity. Similarly, previous studies have demonstrated that the genetic diversity of  
273 invasive populations in their new range may be increased through the admixture of lineages from multiple  
274 native populations (Keller and Taylor 2010; Kolbe et al. 2008). As already indicated, the pattern observed  
275 in the Iberian Peninsula could be the result of multiple introductions of mink deriving from different local  
276 lineages of North America, their rapid expansion, and the encounter of lineages in the north central area

277 of the Iberian Peninsula that may potentially hybridise. However, we cannot discard that cross-breeding  
278 of the different North American subspecies or populations took place within farms (Belliveau et al. 1999;  
279 Dunstone 1993).

280

281 High values of genetic diversity have also been registered in some mink population of Poland  
282 (Zalewski et al. 2011). In this case, mean  $\pi$  for all feral mink analysed was 0.94%, with values for  
283 different populations ranging between 0 and 1.41%. Although the mitochondrial control region fragment  
284 used in that work was different to ours and a direct comparison cannot be performed, the values found in  
285 some populations in Poland were much higher than in others. It is therefore likely that the same processes  
286 suggested for the Iberian Peninsula (i.e. multiple introductions, rapid expansion and encounters of  
287 individuals of different lineages) can be applied to other areas with invasive populations of American  
288 mink.

289

### 290 **4.3 Implications for management**

291

292 This study provides an understanding of the genetic structure and diversity of the populations of the  
293 American mink in the Iberian Peninsula, which can be used to identify key control areas. Thus, one way  
294 to improve the effectiveness of management programmes of an invasive species such as the American  
295 mink is by incorporating molecular-based information into these programs (Hampton et al. 2004). In this  
296 regard, the current spatial mixing of individuals from two different native lineages in the north central  
297 area of the Iberian Peninsula (Basque Country, La Rioja, Navarra and Avila populations) and the  
298 potential interbreeding between such lineages that may take place could translate, among other outcomes,  
299 into a higher fitness in these individuals (e.g. Vandewoestijne et al. 2008). In fact, previous studies have  
300 found a positive correlation between genetic diversity and colonization success as measured by the  
301 reproduction rate (e.g. Crawford and Whitney 2010). Interestingly, Melero et al. (2012b) found that mink  
302 of the Basque Country were bigger in size than mink of the rest of the Iberian Peninsula. According to  
303 Melero et al. (2012b), this phenotypic variation could be due to a mixing of individuals from different  
304 origins, among other possibilities. If the link between spatial mixing of mitochondrial lineages, genome  
305 admixture and higher fitness is demonstrated for some populations, then higher efforts to control and  
306 eradicate the American mink should be undertaken in such populations in order to impede or minimise  
307 their expansion.

308

309 Current controls of the American mink in the Iberian Peninsula focus on reducing their effects on  
310 ecosystems and biodiversity (i.e., areas with local populations of threatened species, wetlands where they  
311 can cause high mortalities in waterfowl and protected natural areas, etc.). In this regard, it is important to  
312 consider that the areas of the Iberian Peninsula with presence of both American mink lineages coincide  
313 with critical areas for the conservation of two of the most threatened mammals of the Iberian Peninsula,  
314 the European mink and the Pyrenean desman. The American mink compete successfully for habitat and  
315 prey with the European mink (Sidorovich et al. 1999; Sidorovich and MacDonald 2001), whose  
316 population of the North of Spain and South West of France is highly isolated from the rest of the

317 distribution (Michaux et al. 2005). On the other hand, the populations of Pyrenean desman of the Central  
318 System have experienced a strong reduction in the last few years and are threatened with extinction.  
319 Although the reasons for this decline are not well known, predation by the American mink may have  
320 contributed to it (Fernandes et al. 2008). Thus, the presence of the two native lineages of the American  
321 mink in the areas with these threatened species should be taken into account in future conservation plans  
322 due the putative higher expansive potential of the American mink in them.

323  
324 The results of this work are in line with previous studies suggesting that mink control should also be  
325 directed to minimise the contact between genetically differentiated populations (Bifulchi et al. 2010;  
326 Fraser et al. 2013; Lecis et al. 2008; Robertson and Gemmell 2004), and so counteract a possible increase  
327 in mink fecundity and dispersal (Melero et al. 2015; Oliver et al. 2016). The present study has shown how  
328 molecular-based information obtained from the invasive range, together with that from the native range,  
329 was critical to characterise the sources from which the Iberian populations of the American mink  
330 originated. Further studies are now necessary to obtain more detailed information, particularly in areas  
331 where admixture of lineages may potentially occur, to help prioritise efforts to control and manage this  
332 invasive species.

333

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335

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348

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509 **Fig. 1** Maximum-likelihood phylogenetic tree derived from the partial mitochondrial control region  
510 sequences for 126 American mink, with eastern and western clades indicated. The number of individuals  
511 per locality is shown in brackets. Sequences from the native range are underlined. The scale indicates  
512 number of substitutions per position.

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515 **Fig. 2** Phylogeographic analysis of the mitochondrial control region sequences of *Neovison vison*. (A)  
516 Haplotype genealogy based on the maximum-likelihood tree. Circle sizes are proportional to the  
517 haplotype frequency and small filled circles are inferred haplotypes not present in the sample. (B, C) Map  
518 of native (B) and invasive (C) ranges showing the 126 samples of *N. vison* used in this study. Red and  
519 blue colours represent the eastern and western clades, respectively. Grey areas represent the original  
520 distribution of *N. vison* downloaded from the IUCN Red List of Threatened Species website (Reid et al.  
521 2016) (B) and the occupancy area of mink in the Peninsula (MAGRAMA, 2013) (C). AL: Alaska, BC:  
522 British Columbia, MT: Montana, WY: Wyoming, TX: Texas, NY: New York, GA: Galicia, CL: Castilla-  
523 Leon, AV: Avila, LR: La Rioja, BC: Basque Country, NAV: Navarra, CAS: Castellon, CAT: Catalonia

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526 **Fig. 3** Map plotting genetic diversity ( $\pi$ ) in different sampling localities of *Neovison vison* in the Iberian  
527 Peninsula. Colours of the sample points indicate the nucleotide diversity according to the provided scale.

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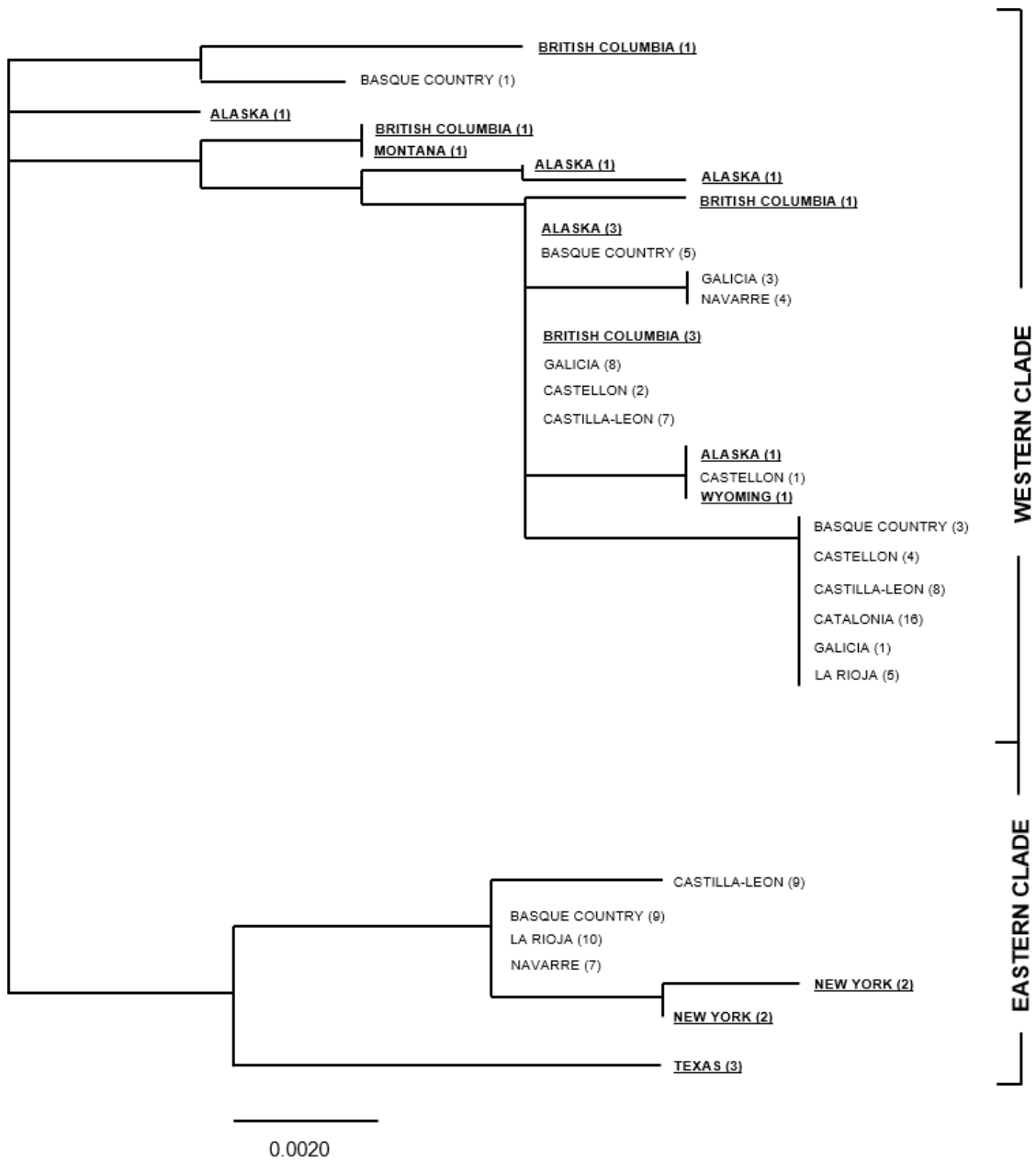
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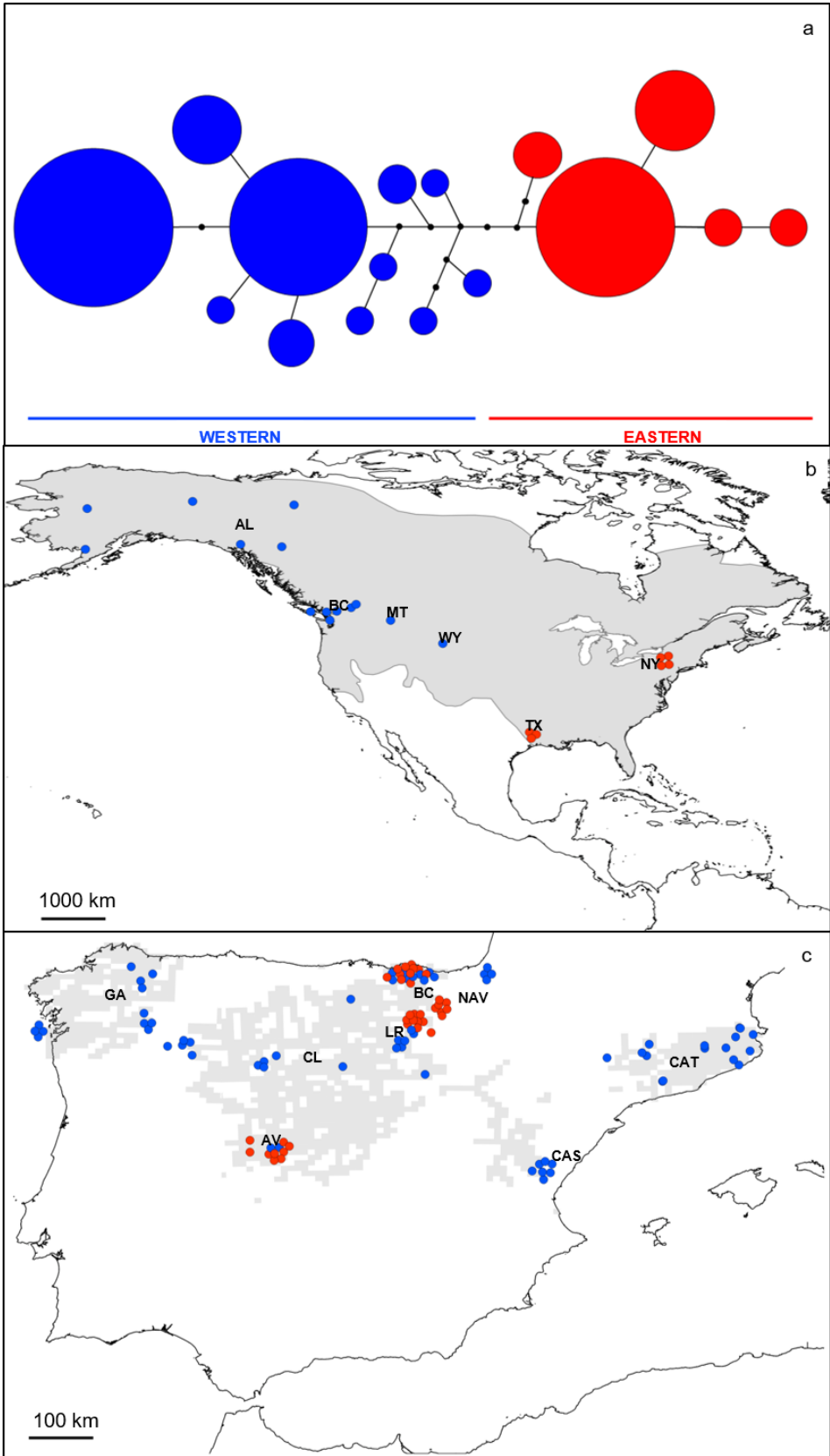
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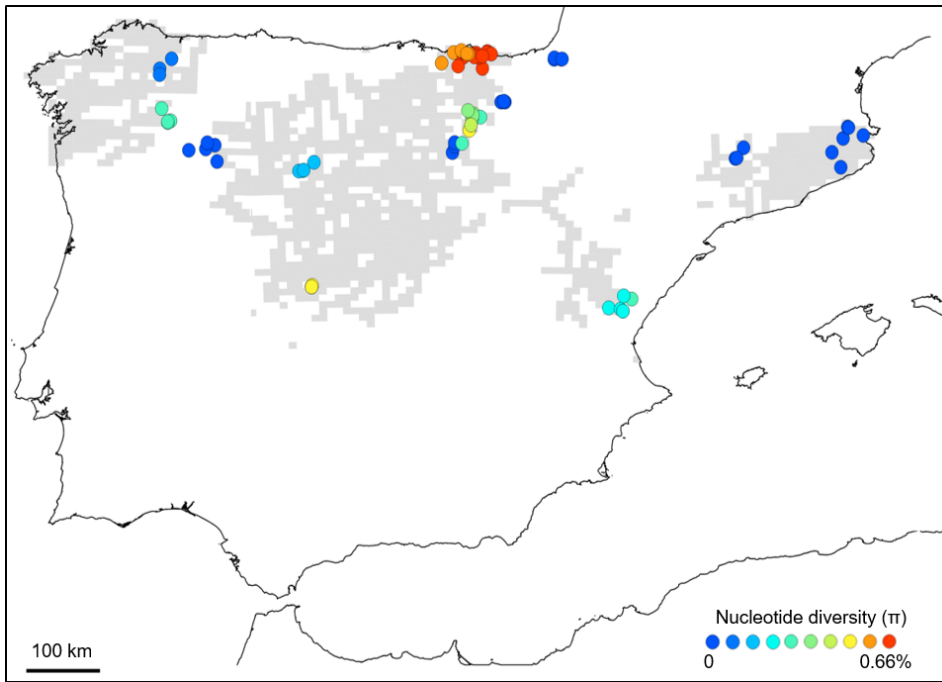
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Fig. 1



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Fig. 2



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Fig. 3