

Prediction of enteric methane production, yield and intensity of beef cattle using an intercontinental database

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48 Abstract

Enteric methane (CH₄) production attributable to beef cattle contributes to global greenhouse gas 49 emissions. Reliably estimating this contribution requires extensive CH₄ emission data from beef 50 cattle under different management conditions worldwide. The objectives were to: 1) predict CH4 51 production (g d⁻¹ animal⁻¹), yield [g (kg dry matter intake; DMI)⁻¹] and intensity [g (kg average 52 daily gain)⁻¹] using an intercontinental database (data from Europe, North America, Brazil, 53 54 Australia and South Korea); 2) assess the impact of geographic region, and of higher- and lowerforage diets. Linear models were developed by incrementally adding covariates. A K-fold cross-55 56 validation indicated that a CH₄ production equation using only DMI that was fitted to all available 57 data had a root mean square prediction error (RMSPE; % of observed mean) of 31.2%. Subsets containing data with $\geq 25\%$ and $\leq 18\%$ dietary forage contents had an RMSPE of 30.8 and 34.2%, 58 59 with the all-data CH₄ production equation, whereas these errors decreased to 29.3 and 28.4%, respectively, when using CH₄ prediction equations fitted to these subsets. The RMSPE of the \geq 60 25% forage subset further decreased to 24.7% when using multiple regression. Europe- and North 61 62 America-specific subsets predicted by the best performing $\geq 25\%$ forage multiple regression equation had RMSPE of 24.5 and 20.4%, whereas these errors were 24.5 and 20.0% with region-63 specific equations, respectively. The developed equations had less RMSPE than extant equations 64 65 evaluated for all data (22.5 vs. 23.2%), for higher-forage (21.2 vs. 23.1%), but not for the lowerforage subsets (28.4 vs. 27.9%). Splitting the dataset by forage content did not improve CH₄ yield 66 or intensity predictions. Predicting beef cattle CH₄ production using energy conversion factors, as 67 applied by the Intergovernmental Panel on Climate Change, indicated that adequate forage 68 content-based and region-specific energy conversion factors improve prediction accuracy and are 69 preferred in national or global inventories. 70

71

- 72 Keywords: empirical modeling, geographical region, forage content, dietary variables, methane
- 73 emission

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1. Introduction

75 The livestock sector emits about 7.1 gigatonnes of CO₂ equivalents of greenhouse gases per year, which represented approximately 14.5% of total global anthropogenic greenhouse gas 76 emissions in 2005 (Gerber *et al.*, 2013). Cattle emitted 4.6 gigatonnes CO_2 equivalents, of which 77 2.5 gigatonnes originated from beef and 2.1 gigatonnes from dairy cattle, whereas small ruminants 78 79 and buffalos emitted 0.47 and 0.62 gigatonnes CO₂ equivalents, respectively. Methane from enteric fermentation contributed about 45% of the combined CO₂ equivalents emissions from the 80 two cattle types. World-wide beef cattle systems produced 35 million tonnes of meat, whereas 81 dairy cattle systems produced 27 million tonnes. Meat protein greenhouse gas emission intensity 82 83 from beef cattle, and combined meat and milk protein intensity from dairy cattle vary from about 200 to 1100, and 50 to 350 kg CO_2 equivalents per kg edible protein, respectively, depending on 84 the region of the world (Opio et al., 2013). Based on expected farming and consumer lifestyle 85 86 practices and the predicted world population growth, compared to 1995, global enteric CH_4 emissions are predicted to increase by 70% by 2055 (Popp et al., 2010). To offset this increase and 87 to deal with the highly variable and typically greater CH_4 emission intensity of beef cattle systems, 88 89 accurate prediction of beef cattle CH₄ emissions across regions are urgently required.

Various beef cattle CH₄ prediction equations, for which a variety of diet and animal characteristics were used as covariates, based on treatments means (*e.g.*, Ellis *et al.*, 2009; Escobar-Bahamondes *et al.*, 2017a) or individual animal data (Ellis *et al.*, 2007; Moraes *et al.*, 2014) have been published. Although the use of individual animal data as applied in the latter two studies contributes to more explained variation of CH₄ production due to dry matter intake (DMI) differences at the animal level, all previously mentioned studies only comprised data from specific geographical locations. In contrast to these equations, which may be appropriate for cattle systems

under similar regional conditions, the widely used Intergovernmental Panel on Climate Change 97 (IPCC) methodology recommends a generic CH_4 energy conversion factor (Y_m) without any 98 adjustment for different geographical locations (IPCC, 2014). The Y_m quantifies enteric CH₄ 99 100 emission as a fraction of the gross energy intake and discriminates between diets with forage contents of ≤ 10 and > 10% DM, with $Y_{\rm m}$ being 3.0% and 6.5% of the gross energy intake, 101 respectively. However, more complex equations accounting for dietary nutrient composition and 102 individual animal characteristics in addition to total feed intake may perform better than those that 103 ignore these covariates for various cattle categories (Ellis et al., 2007, 2009; Moraes et al., 2014; 104 105 Santiago-Suarez *et al.*, 2016). Therefore, more complex beef cattle CH_4 prediction equations that draw from databases with a broad range of diets and geographic conditions may more accurately 106 predict global CH₄ emissions. Publications of inventories that investigated cattle enteric CH₄ 107 emissions in certain countries or regions (e.g., Basarab et al., 2005; Kebreab et al., 2008; Bannink 108 et al., 2011; Castelan-Ortega et al., 2014; Charmley et al., 2016) compared to an intercontinental 109 evaluation (e.g., Niu et al., 2018) confirm the utility of the latter approach. 110

The objectives of the current study were: 1) to collate an intercontinental database of enteric CH₄ production of individual animal records of beef cattle; 2) to determine the key variables for predicting beef cattle enteric CH₄ production (g d⁻¹ animal⁻¹), yield [g (kg DMI)⁻¹] and intensity [g (kg average daily body weight gain)⁻¹] and their respective relationships; 3) to develop and crossvalidate intercontinental and region-specific models, and models for lower- and higher-forage diets.

117

- 118 **2.** Materials and Methods
- 119 *2.1 Database*

120 The 'GLOBAL NETWORK' project is an international collaborative initiative of animal scientists (http://animalscience.psu.edu/fnn; accessed May 16, 2017). All animal scientists with an 121 interest in greenhouse gas research and with access to CH₄ measurements from beef cattle were 122 invited to collaborate and contribute data to this collaborative CH_4 mitigation data analysis. The 123 resultant beef cattle CH₄ database that was developed from this initiative contains 2015 individual 124 beef cattle records from 52 studies conducted from 1969 to 2015 by research entities from Europe 125 (n = 869 from 18 studies), North America (n = 649 from 14 studies), Brazil (n = 313 from 12 studies)126 studies), Australia (n = 174 from 7 studies) and South Korea (n = 10 from 1 study). The European 127 128 studies were conducted in the UK (n = 313 from 7 studies), Switzerland (n = 96 from 1 study), Belgium (n = 72 from 4 studies), Ireland (n = 147 from 2 studies) and France (n = 241 from 4 129 studies). Eleven North American studies were from the United States (n = 492), and 3 were 130 conducted in Canada (n = 157). The database includes records of enteric CH₄ production along 131 with corresponding DMI, dietary gross energy, crude protein, ether extract (EE), neutral detergent 132 fiber (NDF), starch, ash and forage contents, average daily body weight gain (ADG) and body 133 weight (BW). The database comprised a broad variety of beef cattle that included growing and 134 finishing steers, bulls and heifers, pregnant heifers, and pregnant, non-pregnant, dry and lactating 135 136 beef cows. Various pure beef breeds and crossbreeds were included, viz., Aberdeen Angus, Blonde d'Aquitaine, Belgian Blue, Brahman, Brown Swiss × Limousin, Charolais, Hanwoo, Holstein × 137 Zebu, Hereford × Angus, Luing and Nellore. 138

The original studies in the database (complete data bibliography is provided in Supplementary information) investigated the impact of diet composition on enteric CH_4 production or cattle metabolism. However, some studies tested the effect of a specific feed additive, nutrient or the use of hormone supplementation, and the data from these treatments were excluded. The

excluded treatments included rapeseed cake and nitrate (Troy et al., 2015), limestone (Zanetti et 143 al., 2017), Acacia tannins, maca, garlic and lupine seeds (Staerfl et al., 2012), monensin (Caetano 144 et al., 2016, 2018), organosulfur compounds (garlic extracts) (Peiren et al., unpublished) and 145 essential oils (Castro Montoya et al., 2015), lipids (Duthie et al., 2015), dried corn distillers grains 146 (Hünerberg et al., 2013ab), linseed oil and protected fat (Fiorentini et al., 2014), soybean oil and 147 148 protected fat (Silva et al., 2018), glycerin (Lage et al., 2016), whole soybeans (Rossi et al., 2017), monensin (Hales et al., 2012, 2013, 2014 2015, unpublished; Berndt et al., unpublished), 149 diethylstilbestrol (Rumsey et al., 1981) and growth hormone-releasing factor (Lapierre et al., 150 151 1992). After removal of data associated with the aforementioned treatments, 1413 individual records were retained. 152

Records with missing CH₄ or DMI values were removed from the database; records from 153 respiration chambers in which two animals were housed simultaneously were combined by 154 averaging the CH₄ and DMI and all other variables regarding the two animals; records from 155 repeated measurements within the same experimental period were averaged over the individual 156 measurements recorded. In total, 1366 individual animal records were subsequently retained. In 157 addition, records from growing cattle with negative ADG, and a study for which DMI varied from 158 9.0 to 32.5 kg d⁻¹ (Rooke *et al.*, 2015, unpublished) were discarded from the dataset, leaving 1257 159 160 records retained. Finally, studies were screened on the basis of mean CH₄ yield after which two studies, for which the control treatments contained 60 and 82% forage had unrealistically low CH₄ 161 yields of 10.3 and 11.3 g (kg DMI)⁻¹ (San Vito et al., 2016; De Carvalho et al., 2016), respectively, 162 were considered outliers and removed from the dataset. This resulted in the retention of 1248 163 164 records.

165 2.2 Model development

Production, yield and intensity of CH₄ were predicted by fitting mixed-effects modelsaccording to:

168
$$y_{ij} = \beta_0 + \beta_1 x_{ij1} + \beta_2 x_{ij2} + \dots + \beta_k x_{ijk} + s_i + \varepsilon_{ij},$$

169 where y_{ij} denotes the *j*th response variable of CH₄ production (g d⁻¹ animal⁻¹), CH₄ yield [g (kg 170 DMI)⁻¹] or CH₄ intensity [g (kg ADG)⁻¹] from the *i*th study; β_0 denotes the fixed effect of intercept; 171 x_{ij1} to x_{ijk} denote the fixed effects of predictor variables and β_1 to β_k are the corresponding slopes; 172 s_i and ϵ_{ij} denote the random effect of study and residual error, respectively, distributed as $s_i \sim$ 173 $N(0, \sigma_s^2), \epsilon_{ij} \sim N(0, y_{ij}\sigma_e^2)$ for CH₄ production, and $\epsilon_{ij} \sim N(0, \sigma_e^2)$ for CH₄ yield and intensity; σ_s^2 174 is the between-study variance, σ_e^2 is the residual variance, and $y_{ij}\sigma_e^2$ is the residual error variance 175 being proportional to the dependent variable.

176 In order to provide equations that depend on various predictor variables, eight categories of CH₄ production models were developed, of which four used a fixed and another four a selected 177 combination of covariates: DMI only (DMI_C), DMI and dietary NDF content (DMI+NDF_C), 178 DMI and dietary starch content (DMI+STA_C), DMI and dietary EE content (DMI+EE_C); a 179 selection of DMI and the dietary NDF, starch, forage, EE, crude protein and ash contents (Diet_C), 180 the Diet_C covariates plus BW (Animal_C), the Animal_C covariates except DMI 181 (Animal_no_DMI_C), and DMI, the dietary NDF and crude protein contents, and BW (Global_C). 182 Global_C was exclusively associated with covariates that had few or no missing data points. In 183 184 addition to these eight categories, CH₄ production was predicted using Y_m only. The mixed-effects model to estimate $Y_{\rm m}$ of this GLOBAL NETWORK Tier 2 equation only included y_{ij} , β_0 , s_i , and 185 ϵ_{ij} of the previously shown model, with $\epsilon_{ij} \sim N(0, \sigma_e^2)$. According to the CH₄ production models, 186 187 six categories of CH₄ yield prediction models were developed: dietary NDF content only 188 (NDF_C), dietary starch content only (STA_C), dietary EE content only (EE_C); a selection of the dietary NDF, starch, forage, EE, crude protein and ash contents (Diet_no_DMI_C), the Animal_no_DMI_C covariates, and dietary NDF, forage and crude protein contents and BW (Global_no_DMI_C). Finally, eight categories of CH₄ intensity prediction models were developed: DMI_C, DMI+NDF_C, DMI+STA_C, DMI+EE_C, Diet_C, Animal_C, Animal_no_DMI_C and Global_C.

Covariates that play a key role in predicting CH₄ production were selected for Diet_C, 194 Diet_no_DMI_C, Animal_C, Animal_no_DMI_C, Global_C and Global_no_DMI_C using a 195 multistep selection approach. Model selection started with all potential covariates associated with 196 197 the particular model category. Subsequently, one or more next selection steps were performed if not all records without missing values for the selected covariates were used in the previous step. 198 A backward selection approach was applied throughout the different steps, *i.e.*, only covariates 199 200 selected in a previous step could be selected for the next step. The model selection procedure stopped when the selected covariates were the same as the ones selected in the previous step. With 201 this procedure, a model equation was constructed based on records that contained no missing 202 values for the final selection of covariates and data sufficiency was maximized for the development 203 of model equations throughout the different categories. 204

The Bayesian information criterion (BIC; *e.g.*, James *et al.*, 2014) was computed for all fitted models. The BIC is a well-known quantitative approach to model selection that favors more parsimonious models over more complex models by penalizing the number of parameters included in the model. Models with the smallest BIC were selected, as a smaller BIC indicates a better tradeoff between the goodness of fit and the number of model parameters. In addition, the presence of multicollinearity of fitted models was examined based on the variance inflation factor. The largest variance inflation factor among all predictor variables was considered as an indicator of multicollinearity (Kutner *et al.*, 2005). The identified predictor variables were removed from the
model one at a time using a stringent variance inflation factor cutoff value of 3 (Zuur *et al.*, 2010).
All models were fitted using the *lme* function (Pinheiro and Bates, 2000) of R language and
environment for statistical computing (R Core Team 2017; version 3.5.2).

216 2.3 Data handling

The entire database contained a wide variety of dietary forage contents ($57.6 \pm 29.8\%$ DM; 217 average \pm SD), ranging from 8 to 100% DM. The database was split into a higher-forage subset 218 containing the records with $\geq 25\%$ forage, and a lower-forage subset containing all data with \leq 219 220 18% forage. No studies tested forage contents between 18 and 25%. Because of the small coefficient of variation, dietary starch could not be selected for the lower-forage Diet C, 221 Animal_C, and Animal_no_DMI_C equations. To explore the geographical impact of CH₄ 222 production, all European, North American and Brazilian higher-forage data were also used as 223 separate subsets. Because of the scarcity of data from Australia and South Korea, no specific 224 equations for the latter two regions were developed. Data from growing and finishing cattle for 225 which ADG was measured were selected for a growing cattle subset, which enabled the 226 development of CH₄ intensity $[g (kg ADG)^{-1}]$ equations. Other outliers were identified using the 227 228 interquartile range method (Zwillinger and Kokoska, 2000) based on all dependent and independent variables as in Niu et al. (2018). A factor of 1.5 for extremes was used in constructing 229 boundaries to identify outliers for dependent variables and a factor of 2.5 for independent variables. 230 Outliers were identified only for the complete database. The CH₄ intensity [g (kg ADG)⁻¹] values 231 were log transformed to stabilize normality before outlier identification. After removal of records 232 233 with interquartile range identified outliers in the CH₄ production and variables, 1021 records from 234 114 dietary treatments and 39 studies were retained. Of these records 882 were from 104 treatments

and 38 studies in the higher-forage subset, 139 from 10 treatments and 8 studies from the UK,
Ireland, France, Canada and Brazil in the low-forage subset, 307 from 28 treatments and 15 studies
in the European higher-forage subset, 394 from 36 treatments and 10 studies the North American
higher-forage subset, 104 from 17 treatments and 7 studies the Brazilian higher-forage subset, 72
from 22 treatments and 5 studies from Australia, and 5 from 1 treatment and 1 study from South
Korea.

The cleaned dataset used for analysis comprised measurements of enteric CH₄ emission 241 that were obtained from respiration chambers (n = 676), the GreenFeed system (n = 87), and the 242 243 sulfur hexafluoride (SF₆) tracer technique (n = 258). Animals were either kept in confinement or on pasture (n = 991 vs. 30, respectively). Types of forage frequently used in higher-forage diets 244 included fresh alfalfa, sugarcane, sugarcane bagasse, corn silage, barley straw, whole-crop barley 245 silage, whole-crop wheat silage, grass herbage, elephant grass, grass silage, grass seed hay, grass 246 hay wrapping, timothy and natural grassland hay. Types of forage frequently used in lower-forage 247 248 diets were barley straw, wheat straw, whole-crop wheat silage, corn silage and whole-crop barley silage. Concentrate ingredients in higher-forage and lower-forage diets included dried distillers 249 250 grains, barley, canola meal, soybean meal, soybean hulls, crude glycerin, corn grain, cereal by-251 products, dehydrated alfalfa, dehydrated beet pulp, citrus pulp, wheat distillers grains, whole grain oats and minerals. 252

253 2.4 Cross-validation and model evaluation

The predictive accuracy of the developed CH₄ prediction models was evaluated using a leave-one-out cross-validation (*e.g.*, James *et al.*, 2014), in which all individual studies were consecutively taken as the testing set for model evaluation, while all remaining studies were taken as the training set for model fitting. Currently, most national enteric CH₄ inventories are based on 258 energy conversion factors recommended by the IPCC (2006), which were evaluated, *i.e.*, not crossvalidated. The IPCC models and the developed models throughout all categories were, if 259 applicable, evaluated on the various (sub)sets using a combination of model evaluation metrics. 260 Furthermore, equations from Yan et al. (2000, 2009) based on data from Northern Ireland, Ellis et 261 al. (2007) based on data from North America, Ellis et al. (2009) based on data from Canada, Patra 262 263 (2017) based on data from Brazil, India, Australia and Zimbabwe, Escobar-Bahamondes et al. (2017a) based on data from North America, Europe, Australia, Japan and New Zealand, Charmley 264 et al. (2016) based on data from Australia, and the Mitscherlich equation from Mills et al. (2003) 265 266 based on data from the UK were evaluated given that the covariates used in these published equations were available in the present database. Of these previously published extant equations, 267 the equation that performed the best using our data and the single regression equation that only 268 269 depended on DMI and performed the best using our data were reported in the present study. Data from studies included in the present database used for the development of these extant equations 270 271 were excluded from evaluations of those extant equations to ensure independent evaluation.

First, the mean square prediction error (MSPE) was calculated according to Bibby andToutenburg (1977) as:

274 MSPE =
$$\frac{\sum_{i=1}^{n} (O_i - P_i)^2}{n}$$
,

where O_i and P_i denote the observed and predicted value of the response variable for the *i*th observation, respectively, and *n* denotes the number of observations. The square root of the mean square prediction error (RMSPE) was used to assess overall model prediction error. In the present study, RMSPE was expressed as a proportion of observed CH₄ production, yield or intensity means. The MSPE was decomposed into mean bias (MB), slope bias (SB) and random bias to identify systematic biases, of which the MB and SB were calculated as follows:

281 MB =
$$(\bar{O} - \bar{P})^2$$
,

$$SB = (s_p - rs_o)^2,$$

where \overline{O} and \overline{P} denote the predicted and observed means, s_p denotes the standard deviation of 283 predicted values, s_o denotes the standard deviation of observed values, and r denotes the Pearson 284 correlation coefficient. Second, the ratio of RMSPE and s_o , namely RMSPE-observations standard 285 deviation ratio (RSR), which accounts for the specific variability of the data used for evaluation 286 287 (Moriasi et al., 2007), was used to compare the performance of models based on data from different (sub)sets. Smaller values of RSR indicate less variation in the prediction error compared to the 288 standard deviation of the observations, with RSR = 1 indicating the RMSPE variance is equal to 289 observed data variance. If RSR > 1, \overline{O} is a better predictor than P_i . Third, the concordance 290 291 correlation coefficient (CCC; Lin, 1989), which quantifies both accuracy and precision based on 292 the bias correction factor (C_b) and r by comparing the best-fit line and observations to the identity 293 line (y = x), respectively, was calculated. The CCC is given as:

294
$$CCC = r \cdot C_b$$
,

The closer the CCC of a model to 1, the better the model performance.

Different forage proportion cutoff values with increments of 5% from 15 to 50% were tested to evaluate the effect of the cutoff for splitting the database into higher-forage and lowerforage subsets on equation performance. Cutoff values of 0, 15, 20, 25, 30, 35, 40, 45, and 50% forage DM were used for evaluation. Per cutoff value, an RSR weighted to the number of observations for the DMI_C equation was calculated for the higher-forage and lower-forage CH4 production equations, after which the optimal cutoff value could be determined.

302

303 3. Results

304 The inclusion criterion for dietary treatment had different effects on the variables means, viz., DMI (8.13 vs. 8.06 kg d⁻¹; cleaned vs. uncleaned averages, respectively), and NDF (35.0 vs. 305 35.0% of DM), starch (34.0 vs. 30.5% of DM), EE (3.02 vs. 3.52% of DM), ash (6.29 vs. 7.26% 306 307 of DM), and forage (51.0 vs. 58.1% of DM) content of the diet, BW (478 vs. 487 kg), CH₄ production (161 vs. 164 g d⁻¹ animal⁻¹), CH₄ yield [20.0 vs. 20.4 g (kg DMI)⁻¹], CH₄ intensity [145] 308 vs. 207 g (kg ADG)⁻¹] and $Y_{\rm m}$ (6.0 vs. 6.0 % of the gross energy intake). Summary statistics for the 309 (sub)sets of the present cleaned database that included intake, dietary nutrient composition, BW, 310 ADG and CH₄ variables are presented in Tables 1 and S1. 311

312 3.1 Methane production equations

The DMI_C all-data CH₄ production (g d⁻¹ animal⁻¹) equation indicated a positive 313 relationship of DMI with CH₄ production (Eq. 1; Table 2; regression coefficient $\pm 2.5E$ gives a 314 rough estimate of the 95% confidence interval boundaries that correspond to a *P*-value of 0.05, all 315 *P*-values < 0.05 were not reported). The DMI+NDF_C, DMI+STA_C and DMI+EE_C equations 316 had positive, negative and negative regression coefficients for dietary NDF, starch and EE in 317 relation to CH₄ production, respectively (Eqns. 2-4). The RSR, which is the most appropriate 318 statistic for evaluating equations based on different numbers of observations, for the DMI_C, 319 320 DMI+NDF_C, DMI+STA_C and DMI+EE_C equations indicated similar predictive performance, whereas the CCC indicated the DMI+NDF_C equation performed better than the DMI_C and 321 DMI+EE_C equations (0.63 vs. 0.60 and 0.61, respectively). Dietary forage content and DMI were 322 323 selected for the Diet_C and Animal_C equations (Eqns. 5-6), with BW also selected for the Animal_C equation. Dietary forage and ash and BW were selected for the Animal_no_DMI_C 324 325 equation (Eq. 7). The Animal_C was the best performing all-data equation developed in the present 326 analysis, with RSR and CCC of 0.61 and 0.76, respectively. Across the developed all-data equations, slope bias ranged from 1.01-12.7%, which was consistently associated with underprediction at the high end and over-prediction at the low end of production (Fig. 1). Overall, models with a higher number of covariates tended to have less slope bias and had less betweenstudy variance (σ_s^2 not shown).

The RSR of the all-data DMI_C CH₄ production equation was 0.71 (Table 3). Splitting the 331 database into higher-forage and lower-forage subsets at cutoffs of 15 to 50% resulted in very 332 similar weighted average RSR values of 0.68 to 0.69. The cutoff of 20% that was applied resulted 333 in an RSR of 0.94 for the lower-forage subset at this cutoff value, whereas the cutoff values from 334 25 to 50% had all lower RSR values for the lower-forage subset. This might suggest that the lower-335 336 forage subset is a better predictor at a higher cutoff. However, the prediction of the data associated with $\leq 20\%$ forage did not improve at cutoff values > 20% (results not shown), indicating that data 337 338 with > 20% forage decreased the RSR of the lower-forage subset, but not the data associated with $\leq 20\%$ forage. Based on these differences in performance and the fact that diets containing $\leq 20\%$ 339 340 forage are commonly fed to cattle in intense feedlot production systems, the data were split at 20% forage throughout the present study, which made all lower-forage data contain $\leq 18\%$ forage and 341 342 the higher-forage $\geq 25\%$ forage.

The higher-forage CH₄ production equations overlapped with the all-data equations, where DMI and dietary NDF, starch and EE in the DMI_C, DMI+NDF_C, DMI+STA_C and DMI+EE_C equations showed regression coefficients with the same sign (Eqns. 12-15; Table 4). Moreover, similar covariates were selected for the Diet_C, Animal_C and Animal_no_DMI_C equations as for the all-data equations, although the Animal_no_DMI_C equation did not contain dietary ash (Eqns. 16-18). The higher-forage equations predicted the higher-forage subset better than the all-data equations, with mean RSR of 0.62 *vs.* 0.66 and CCC of 0.70 *vs.* 0.68, respectively, for the DMI_C, DMI+NDF_C, DMI+STA_C, DMI+EE_C, Diet_C, Animal_C and Animal_no_DMI_C equations. The developed higher-forage equations under-predicted CH₄ production at the high end and over-predicted it at the low end of production, with the multiple regression equations having less slope bias than the DMI_C equation (Fig. 2). In line with the alldata equations, models with a higher number of covariates had less between-study variance.

355 In accordance with the all-data and the higher-forage equations, DMI was positively related to CH₄ production in the lower-forage DMI_C equation (Eq. 20; Table 5). The DMI+NDF_C, 356 DMI+STA C and DMI+EE C equations indicated no significant relationships between the 357 358 corresponding dietary NDF, starch and EE contents with CH₄ production (Eqns. 21-23; P-values of 0.14, 0.10 and 0.57, respectively). The lower-forage DMI C equation predicted the lower-359 forage subset better than the all-data equations based on RSR, whereas the highest CCC of 0.35 360 for the lower-forage subset were obtained from the all-data DMI+STA_C and Animal_C equations 361 (Eqns. 3, 6; Table 2). Systematic bias, that is the sum of mean and slope bias, was less than 5.75% 362 for these developed lower-forage equations (Table 5), except for the DMI+STA C equation that 363 had 3.70 and 20.18% mean and slope bias, respectively. The minor slope bias of the lower-forage 364 DMI_C equation ($\leq 0.03\%$) was due to under-prediction of CH₄ production at the high end and 365 366 over-prediction at the low end (Fig. 3).

In contrast to the higher-forage equations, dietary NDF and starch contents in the European higher-forage DMI+NDF_C and DMI+STA_C equations were not related to CH₄ production (Eqns. 29-30, Table 6; *P*-values of 0.20 and 0.69, respectively). Furthermore, DMI, dietary NDF and EE were selected for the Diet_C equation (Eq. 32) with BW also being selected for the Animal_C equation (Eq. 33), whereas DMI and BW, and BW were selected for the Global_C and Animal_no_DMI_C equations, respectively (Eqns. 34-35). The North American higher-forage

equations were largely in line with the higher-forage equations. However, the Animal_ no DMI C 373 equation also contained dietary ash (Eq. 44; Table 7) as obtained for the all-data equation, and the 374 Global C equation also contained dietary crude protein (Eq. 45), The European higher-forage and 375 376 North American higher-forage equations under-predicted CH₄ at the high end and over-predicted it at the low end of production, except for the European higher-forage DMI+EE C equation, which 377 378 under-predicted CH₄ at the low end and over-predicted at the high end (Figs. 4-5). Dietary NDF and EE contents in the Brazilian higher-forage DMI+NDF_C and DMI+EE_C equations were not 379 significantly related to CH₄ production (Eqns. 49-50, Table S2; P-values of 0.28 and 0.05, 380 381 respectively), the Diet C equation contained DMI and dietary ash (Eq. 51), whereas the Animal no DMI C equation contained dietary forage (Eq. 52). Slope bias varied from 9.05 to 382 18.9% for the developed Brazilian higher-forage equations, except for the Animal_no_DMI_C 383 equation for which 32.9% slope bias was obtained. Equations under-predicted CH₄ production at 384 the low end and over-predicted at the high end, whereas the Animal_no_DMI_C equation showed 385 a negative observed vs. predicted correlation (Fig. S1). Compared to the higher-forage equations, 386 the European higher-forage, North American higher-forage and Brazilian higher-forage data were 387 more adequately predicted by the European higher-forage (mean RSR of 0.80 vs. 0.85, mean CCC 388 389 of 0.50 vs. 0.48; respectively; Tables 4, 6), North American higher-forage (mean RSR of 0.53 vs. 0.57, mean CCC of 0.80 vs. 0.77; respectively; Tables 4, 7) and Brazilian higher-forage (mean 390 RSR of 1.13 vs. 1.35, respectively; Tables 4, S2), although mean CCC indicated Brazilian higher-391 392 forage data was more adequately predicted using the higher-forage than the Brazilian higherforage equations (0.17 vs. 0.11, respectively; Tables 4, S2). 393

The IPCC (2006) Tier 2 higher-forage equation had an RSR of 0.68 and a CCC of 0.75 when evaluated using all data (Eq. 9; Table 2). Predicting the higher-forage subset with this

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396 equation resulted in RSR and CCC of 0.53 and 0.84, respectively (Eq. 9; Table 4). Despite this high accuracy of prediction of the Tier 2 approach, increased variance appeared along the unity 397 line of the predicted vs. observed plots (Figs. 1-2). The IPCC Tier 2 (2006) lower-forage equation 398 had an RSR of 1.38, a CCC of 0.17 and 59.6% mean bias for the lower-forage subset (Eq. 25; 399 Table 5). The GLOBAL NETWORK Tier 2 equations with Y_m of 6.1% and 6.3% (Eqns. 8, 19; 400 401 Tables 2, 4) performed slightly better than the IPCC Tier 2 (2006) equation for the all-data and higher-forage (sub)sets, respectively [note that the IPCC equations were validated, the GLOBAL 402 NETWORK equations were cross-validated], whereas the lower-forage GLOBAL NETWORK 403 404 Tier 2 equation with Y_m of 4.5% resulted in RSR of 0.90, a CCC of 0.43 and 0.47% of mean bias (Eq. 24) performed obviously better than the lower-forage IPCC Tier 2 equation. Although the 405 406 IPCC currently uses a 10% forage cutoff, a $Y_{\rm m}$ of 4.5% is still more accurate than a $Y_{\rm m}$ of 3.0% for the present data, with RSR being 0.98 and 1.51, and CCC being 0.40 and 0.16 for the GLOBAL 407 NETWORK and IPCC Tier 2 lower-forage equations, respectively (Eqns. 24-25). The European 408 409 higher-forage and North American higher-forage subsets were associated with RSR of 0.66 and 0.48, and CCC of 0.71 and 0.88 for the IPCC Tier 2 (2006) equation, respectively (Eq. 9; Tables 410 6-7), whereas RSR of 1.81 and CCC of 0.21 were obtained for the Brazilian higher-forage subset 411 412 (Eq. 9; Table S2). Compared to the latter equation, the GLOBAL NETWORK Tier 2 equations with Y_m of 6.6 and 6.3% performed similarly based on RSR and CCC for the European higher-413 414 forage and North American higher-forage subset (Eqns. 36, 46; Tables 6-7), whereas less mean 415 bias was obtained with 1.89 vs. 3.54% and 2.51 vs. 8.70%, respectively. The Brazilian higherforage subset was better predicted when using the GLOBAL NETWORK Tier 2 approach resulted 416 417 in a *Y*_m of 5.5%, an RSR of 1.29, and a CCC of 0.28 (Eq. 53; Table S2).

418 Equations developed by Ellis et al. (2009), Charmley et al. (2016) and Escobar-Bahamondes et al. (2017a) were among the best performing extant equations and outperformed 419 the Yan et al. (2000, 2009), Mills et al. (2003), Ellis et al. (2007) and Patra (2017) equations for 420 421 all (sub)sets. The best performing equation of Charmley et al. (2016) performed better than the all-data DMI C equation (Eqns. 1, 10; Table 2). The all-forage equation of Escobar-Bahamondes 422 423 et al. (2017a) appeared to perform most accurately among all of the equations (Eq. 11). However, only 646 data points were available for independent evaluation. Based on RSR, it did not 424 outperform the Animal C equation for these 646 data points. For the higher-forage subset, the best 425 426 Charmley et al. (2016) and the Escobar-Bahamondes et al. (2017a) equations performed the best based on CCC (Eqns. 20, 11; Table 4), but not on RSR. The Ellis et al. (2009) equation that also 427 depended on the NDF:starch ratio (Eq. 26; Table 5) performed the best for the lower-forage data 428 429 with RSR of 0.89 and CCC of 0.41. For the European higher-forage subset, the best Charmley et al. (2016) and the Escobar-Bahamondes et al. (2017a) equations (Eqns. 37, 11; Table 6) did not 430 perform better than the Animal_C equation when just considering RSR and CCC values, although 431 the Animal_C equation was evaluated using fewer data points. For the North American higher-432 forage subset, the best performing Charmley et al. (2016) equation (Eq. 47; Table 7) performed 433 434 similarly to the Global_C equation based on RSR, whereas the Charmley et al. (2016) equation performed even slightly better based on CCC. Despite the accuracy of the various equations of 435 Charmley et al. (2016) and in contrast to the Animal_C equations, the predicted vs. observed plots 436 437 showed increasing variation along the unity line for all-data in particular (Fig. 1). However, the best-performing equations that were developed, which was the Animal_C equation for most 438 439 subsets, did not show increasing variation along the unity line. This indicates that the best 440 performing equations that were developed explain variation that is not captured by the Charmley

et al. (2016) equations. These higher precisions obtained from the best performing equations is
also indicated by the correlation coefficients of predicted *vs.* observed values on which the CCC
is calculated (result not shown).

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3.2 Methane yield equations

Positive, negative and negative slope regression coefficients were obtained for the NDF C, 445 STA_C and EE_C all-data CH₄ yield [g (kg DMI)⁻¹] equations (Eqns. 54-56, Table S3), 446 respectively, which aligned with the all-data CH₄ production equations. The Diet_no_DMI_C and 447 Global_no_DMI_C equations selected dietary forage (Eqns. 57-58), whereas dietary EE and ash 448 449 were also selected for the Diet no DMI C equation, and dietary crude protein for the Global no DMI C equation. The NDF C, STA C, EE C, Diet no DMI C 450 and Global_no_DMI_C equations had RSR values of 0.98, 1.06, 1.01, 0.97 and 0.96, respectively. The 451 NDF_C, STA_C and EE_C higher-forage CH₄ yield equations indicated positive, negative and 452 negative relationships to CH₄ yield, respectively (Eqns. 59-61, Table S4), whereas only dietary 453 forage content was selected for the Diet_no_DMI_C equation (Eq. 62). The higher-forage CH₄ 454 yield was associated with RSR of 1.03 to 1.21 (Table S3) when predicted by the all-data equations, 455 whereas the higher-forage equations predicted CH₄ yield of this subset with RSR values from 0.98 456 457 to 1.04 (Table S4). The higher-forage equations reproduced the observed variation in CH₄ yield less adequately than the all-data equations, with even a negative observed vs. predicted relationship 458 for the higher-forage STA_C and EE_C equations (Figs. S2-S3). 459

460 3.3 Methane intensity equations

In contrast to the CH₄ production equations, the DMI regression coefficients in the all-data DMI_C and DMI+EE_C CH₄ intensity equations [g (kg ADG)⁻¹] contained zero in their confidence intervals (*P*-values of 0.14 and 0.22, respectively), whereas the DMI+NDF_C and 464 DMI+STA C equations had a positive regression coefficient for DMI (Eqns. 63-66; Table S5). In line with the CH₄ production equations, dietary NDF, starch and EE contents in the DMI+NDF_C, 465 DMI+STA C and DMI+EE C equations had positive, negative and negative relationships with 466 CH₄ intensity, respectively. Dietary forage content was selected for the Diet C, 467 Animal no DMI C and Global C equations (Eqns. 67-69), with DMI also being selected for the 468 Diet_C equation and BW also being selected for the Global_C equation. The Diet_C, 469 Animal_no_DMI_C and Global_C equations had RSR values of 0.99, 1.00 and 0.96, respectively, 470 and appeared to predict the variation in CH₄ intensity most adequately (Fig. S4), whereas the other 471 472 all-data CH₄ intensity equations had RSR greater than 1 and appeared to predict the variation in CH₄ intensity less adequately. 473

The higher-forage DMI_C, DMI+STA_C and DMI+EE_C equations did not indicate that 474 DMI was related to CH₄ intensity (Eqns. 70, 72-73, Table S6; P-values of 0.06, 0.52 and 0.93, 475 respectively). Dietary NDF was positively related to CH₄ intensity (Eq. 71), whereas dietary starch 476 and EE contents were not related to CH₄ intensity (Eqns. 72-73; P = 0.32). Dietary ash content 477 was selected for the Diet_C equation (Eq. 74), whereas BW were selected for the Animal_C and 478 Global_C equations (Eqns. 75-76), with dietary NDF also being selected for the Global_C 479 480 equation. All higher-forage CH₄ intensity equations had $RSR \ge 1.03$. Furthermore, as also obtained for the higher-forage CH₄ yield equations, the higher-forage CH₄ intensity equations did not 481 reproduce the observed variation in CH₄ intensity of the higher-forage subset more adequately than 482 483 the all-data CH₄ intensity equations (Figs. S4-S5), which was also indicated by the RSR and CCC values. 484

485

486 **4. Discussion**

487 Global applicability is an important attribute of prediction equations of beef cattle enteric CH₄ emission. Various beef cattle systems that are applied world-wide may fit in our analysis. For 488 more details about these beef cattle fattening systems, we refer to *e.g.*, De Vries *et al.* (2015), 489 Gerssen-Gondelach et al. (2017) and Drouillard (2018). Our database, in which data (1021 490 individual records) from a variety of geographical regions across the world is represented, 491 492 therefore, contributes to the overall robustness and global applicability of our all-data and higherforage equations in particular. Hence, CH₄ production of beef cattle will be accurately predicted 493 for data samples that represent a wider set of conditions throughout the world, which is a unique 494 495 feature of the present equations. Several CH_4 prediction equations for beef cattle have been published previously, but they were developed from relatively small databases and only for one 496 specific geographic region, such as Yan et al. (2009) using 108 individual animal records from 5 497 studies from Northern Ireland, Ellis et al. (2007) using 83 treatment means from 14 studies from 498 North America, Ellis et al. (2009) using 872 individual animal records from 12 studies from 499 500 Alberta (Canada), and Moraes et al. (2014) using individual records from 414 heifers and 458 steers housed at one research station in the United States. 501

In the present study, we collated a wide-ranging database that included a large number of 502 503 studies from Europe, North America, Brazil, Australia and South Korea, which represented diverse global beef production systems. Studies from tropical areas were, however, not predominant in the 504 505 present analysis, for which we refer to Charmley et al. (2016) who included studies from tropical 506 Australia, and Patra (2017) who included studies from India, Zimbabwe, Australia and Brazil. Furthermore, Escobar-Bahamondes et al. (2017a) had a database comprising 148 treatment means 507 from 38 studies with diets containing > 40% forage, and a database comprising 43 treatment means 508 509 from 17 studies with diets containing < 20% forage. Therefore, their analysis for lower-forage

510 diets, in particular, included more data from more studies than ours, but their cutoff values for lower and higher forage were based on differences in microbiome composition rather than the 511 prediction error used in the present analysis. Furthermore, their analysis did not explore 512 intercontinental variation in beef cattle CH₄ emissions and did not have the benefit of using 513 individual animal records. Other unique strengths of the present study are the development of CH_4 514 515 yield and intensity equations, whereas beef cattle studies are commonly limited to only total CH₄ production, and the inclusion of dietary forage content as a covariate of the three CH₄ emission 516 metrics. 517

518 Our database includes data obtained with different CH₄ (viz., respiration chambers, GreenFeed system, SF₆) and DMI (viz., weighing and estimating using marker techniques) 519 520 measurement methods. The different CH₄ measurement techniques have their strengths and 521 weaknesses (Hammond et al., 2016; Hristov et al., 2018), whereas directly weighing the amount of feed offered and refusals and their dry matter content is regarded as more accurate than the 522 523 ytterbium and *n*-alkane markers used for some studies in the present database, which may over- or underestimate DMI (Pérez-Ramírez et al., 2012). However, the development of a DMI_C CH4 524 production equation specific for respiration chamber, GreenFeed system and SF₆ subsets did not 525 526 consistently improve the RSR and CCC of the corresponding subsets compared with the all-data DMI_C equation (results not shown). Furthermore, including CH₄ measurement method as a 527 528 covariate in the statistical model of an all-data DMI_C equation did not improve the model fit. 529 Similar results were obtained for measurement method of DMI. Therefore, CH₄ and DMI measurement methods did not have a major effect on the performance of the equations developed 530 531 in the present analysis. However, the relatively high and low accuracies with which the region-532 specific subsets could be predicted may be related to the CH₄ measurement methods, because the percentage of use of respiration chambers in the European higher-forage, North American higherforage and Brazilian higher-forage subsets differed substantially (48, 95 and 0%, respectively).
Finally, statistically accounting for cattle breed or cattle type (*e.g.*, steers, heifers, cows) was
considered, but did not or not consistently improve the prediction of CH₄ production throughout
the subsets.

538 Non-linear CH₄ prediction equations such as the Mitscherlich equation were previously found to outperform linear equations in some studies (e.g., Mills et al., 2003; Patra, 2017). 539 However, for the present database, fitting non-linear equations, viz., Monomolecular, Exponential, 540 541 Mitscherlich and Power forms, did not result in improved prediction of CH₄ production compared to the linear DMI C equations (result not shown). The latter result is in line with the non-linear 542 Mills et al. (2003) and Patra (2017) equations that did not outperform the linear Charmley et al. 543 (2016) equations. This suggests that a multiple linear regression approach, as used for the 544 development of our Animal_C and Global_C equations rather than non-linear approaches, 545 improves the precision and accuracy of prediction of CH₄ production. The utility of ADG and 546 digestibility of EE, NDF, nitrogen, gross energy, DM and organic matter for predicting CH₄ was 547 also evaluated, but these covariates did not result in better prediction of CH₄ production than 548 549 achieved by the various equations that are presented.

The linear regression equations of Charmley *et al.* (2016) that depended on DMI and outperformed our DMI_C equations were fitted using models that included more terms than just DMI, which resulted in nearly unbiased predictions of CH₄. Furthermore, the data Charmley *et al.* (2016) used were only from certain regions in Australia and may have been relatively homogeneous. The equations developed using these data may then result in accurate prediction of CH₄ production based on only DMI. Therefore, the prediction bias for our various Animal_C and

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556 Global_C equations and some potential overestimation of between-study variance that remained may vanish by the inclusion of even more covariates in the statistical model. The negligible bias 557 obtained for the European higher-forage Animal C equation, for which dietary crude protein and 558 EE contents were available for all individual animal records used for fitting this model, and 559 suggests that multiple regression equations are associated with less bias. Despite prediction biases 560 561 of mixed-effects models being associated with the inclusion of the random study effect (see also White et al., 2017), which applies to models with fewer covariates in particular, omission of the 562 random study effect will affect the inference made on the covariates and may result in type II errors 563 564 (St-Pierre, 2001). Therefore, for achieving unbiased predictions, mixed-effects models are ideally applied to datasets without missing values throughout the different covariates. Such datasets will 565 result in greater variation of the dependent variable explained by multiple fixed-effects terms and 566 less overestimation of the random study effect. 567

568 4.1 Key predictor variables

Dry matter intake was the most important predictor of enteric CH₄ production as it was 569 significantly and positively related to CH₄ production for all-data and the higher-forage, lower-570 forage, European higher-forage and North American higher-forage subsets. A positive relationship 571 572 between DMI and CH₄ production is in agreement with previous dairy and beef cattle studies (*e.g.*, Ellis et al., 2007; Hristov et al., 2013a; Richmond et al., 2015; Bell et al., 2016; Charmley et al., 573 574 2016; Niu *et al.*, 2018) and this is because more CH_4 is produced when more substrate is available 575 for microbial fermentation and in turn methanogenesis. In addition, all Diet_C and Animal_C models based on these five (sub)sets selected DMI for the prediction of CH₄ production, and the 576 577 Animal_no_DMI_C equations did not perform as well as the Animal_C equations, indicating the 578 importance of DMI relative to other covariates.

579 The positive relationship between the all-data CH₄ production and dietary NDF content also aligns with previous results (e.g., Ellis et al., 2007; Yan et al., 2009; Niu et al., 2018). The 580 coefficients of variation were 43.8, 45.4 and 30.3% for CH₄ production, and 32.0, 30.6 and 8.6% 581 for dietary NDF content for the all-data and the higher-forage and lower-forage subsets, 582 respectively. This decrease in variation is in line with the disappearance of this positive 583 relationship for the all-data and higher-forage vs. the lower-forage equations. Therefore, 584 developing subsets with limited variation in forage percentage seems to have masked the positive 585 relationship between CH₄ production and dietary NDF content. Furthermore, dietary nutrient 586 587 contents change at the expense of other nutrients. Dietary NDF content may increase at the expense of more rapidly fermentable carbohydrates, which is positively associated with CH₄ production 588 589 (Hatew et al., 2015). The latter hypothesis aligns with a model with DMI and dietary NDF and starch fitted to all data having regression coefficients that were positive, positive and not different 590 from zero for DMI and dietary NDF and starch, respectively (result not shown). However, the 591 lignin fraction of NDF being undegradable indicates that increased dietary NDF may not result in 592 more CH₄ production in case of high lignin contents. Warner et al. (2016) observed lower CH₄ 593 production but higher CH₄ yield per unit of digestible organic matter for dairy cattle fed grass 594 595 silage of high lignin and NDF content, compared with grass silage of low lignin and NDF content. The observation of Na et al. (2017) who found different CH₄ yields per unit of DMI for deer and 596 goats, but not per unit of digestible DMI may also support this hypothesis. 597

598 Dietary starch content is negatively related to CH_4 production as it typically increases 599 propionate production in the rumen, yielding less H_2 for the reduction of CO_2 to CH_4 (Martin *et* 600 *al.*, 2010; Grainger and Beauchemin, 2011). The effect of dietary starch on CH_4 production 601 appeared to be less pronounced for higher-forage diets in dairy cows (Van Gastelen *et al.*, 2015), 602 which may explain why no relationship between dietary starch content and CH₄ production was found for the European higher-forage subset, which had the highest forage content of all subsets. 603 Furthermore, it was suggested that a critical dietary content of starch is required to decrease CH₄ 604 production (Martin et al., 2010; Van Gastelen et al., 2015), possibly more than approximately 20% 605 of DM, and that slight differences in intakes of starch, and other major carbohydrates (e.g.,606 607 hemicellulose, cellulose and lignin) cannot explain the difference in CH₄ emissions of cattle (Moe and Tyrrell, 1979; Moate et al., 2018). This may also explain why no relationship between dietary 608 starch content and CH₄ production was obtained based on the European higher-forage subset. The 609 610 lack of a relationship between CH₄ production and dietary starch content for the lower-forage subset may be related to the small variation in starch content (coefficient of variation is 13.1%). 611

The positive relationship that was obtained between CH₄ production and dietary forage 612 aligns with previously published studies (e.g., Yan et al., 2000; Hristov et al., 2013) stating that 613 either increased forage or decreased concentrate proportion in the diet yielded more CH₄. Johnson 614 and Johnson (1995) referred to cattle fed more than 90% concentrate producing only half of the 615 CH₄ produced by cattle fed more common concentrate proportions, and Aguerre et al., 2011 616 observed a linear increase in CH_4 yield upon increasing dietary forage content from 47 to 68%. 617 618 Nevertheless, a modeling study by Sauvant and Giger-Reverdin (2009) predicted that a decrease 619 in CH₄ yield is only observed for dietary forage contents less than 65%. Despite the latter prediction, the frequent appearance of dietary forage in the equations developed in the present 620 621 study indicates dietary forage content is a decent predictor of CH₄ emission, possibly more robust than dietary NDF content that was less frequently selected for the developed equations. 622

Dietary lipid content is commonly negatively related to CH₄ production (Grainger and
Beauchemin, 2011). Lipids may inhibit cellulolytic bacteria, protozoal and archaeal activity,

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625 decrease NDF digestibility, and supply non-fermentable energy to the rumen, outcomes that can decrease CH₄ production (Maia et al., 2007; Beauchemin et al., 2008; Guyader et al., 2014). Long-626 chain saturated fatty acids may have a minimal inhibitive effect on archaeal activity and CH₄ 627 production, whereas fatty acids such as C12:0 and C18:3 were found to be relatively potent 628 reducers (Machmüller and Kreuzer, 1999; Patra, 2013). Therefore, the actual decrease in CH₄ 629 630 production obtained from lipids may depend on their fatty acid composition, although this is not confirmed by all in vivo studies (e.g., Grainger and Beauchemin, 2011). More importantly, the 631 removal of data associated with dietary lipid and oil supplements excluded data with higher dietary 632 633 EE contents, which more potently decrease CH_4 production (Patra, 2013), may explain why DMI+EE C equations did not perform better than the DMI C equations, despite the significant 634 relationships that were obtained for CH₄ production and yield with dietary EE content. 635

Dietary crude protein content being positively associated with the all-data CH₄ yield in the 636 present analysis aligns with the observation that dietary nitrogen content is positively related to 637 fiber digestibility (Dijkstra et al., 1996). However, decreased CH₄ production may only be 638 observed from cattle fed a diet that is deficient in rumen degradable protein (Sutter *et al.*, 2017). 639 In the present study, we did not observe any relationship between crude protein content and NDF 640 641 digestibility, r = 0.04. However, we did observe a correlation between crude protein content and organic matter and dry matter digestibility, r = 0.42 and r = 0.37, respectively. This is possibly due 642 to higher starch degradability, which could not be verified because of the lack of starch 643 644 degradability data. Van Lingen et al. (2018) applying a multivariate regression approach found that the methodological issues such as the structure of random-effects (co)variance matrices and 645 646 the combination of fixed-effects variables affect the statistical inference regarding the relationship 647 between dietary crude protein and CH₄ production or yield. Therefore, also based on dietary crude 648 protein selected for only one equation, the latter relationship may not be commonly strong as well 649 as it may not generally exist. Dietary crude protein may actually be associated with lesser CH₄ 650 production when sufficient rumen degradable protein is fed so as not to limit fermentation in the 651 rumen due to N shortage (Dijkstra *et al.*, 2011), and may be considered a less robust predictor of 652 CH₄ production than dietary NDF and starch.

A positive relationship between BW and CH₄ production observed in various equations in 653 the present analysis aligns with previous cattle research (Yan et al., 2009; Moraes et al., 2014; 654 Escobar-Bahamondes et al., 2017a). Demment and Van Soest (1985) and Smith and Baldwin 655 656 (1974) observed rumen volume and weight proportional to BW of animals. Consequently, smaller animals ingest less feed and emit less CH₄ (Hristov *et al.*, 2013b). In addition, empirical modeling 657 (Sauvant and Nozière, 2016) and mechanistic model simulations (Huhtanen et al., 2015, 2016) 658 659 indicated the DMI/BW ratio to be an important factor for CH₄ yield. At similar DMI, smaller cattle tend to produce less CH₄ as the passage rate from the rumen to the intestine may be higher due to 660 a greater DMI/BW ratio. This has been confirmed in sheep for which animals yielding less CH₄ 661 had smaller rumen size (Goopy et al., 2014). Therefore, BW influences DMI, and DMI and rumen 662 volume determine the passage rate of ruminal digesta, which affects feed digestibility, rumen 663 664 fermentation conditions, and ultimately CH₄ production and yield.

665 **4.2 Best performing equations**

Various equations and model categories for predicting beef cattle CH_4 emission have been applied on various subsets in the present study. The Y_m models have only one parameter and are the simplest models, the DMI_C models are still fairly simple, whereas the Animal_C models are potentially the most complex. The Animal_C model commonly performed best among all models and outperformed the GLOBAL NETWORK Tier 2 equations, except for the European higher671 forage subset. The DMI appeared to be the major predictor of enteric CH₄ production in beef cattle, but may not always be available for individual animals on commercial farms, which points to the 672 value of the Animal no DMI C models. Using dietary forage content and BW as a covariate 673 commonly improved the prediction of CH₄ compared with a DMI C equation. Therefore, the on-674 farm availability of all previously mentioned variables is recommended. This availability also 675 676 enables the evaluation of the effect of dietary nutrient composition on CH₄ production. Moreover, the DMI+STA_C equation (Eq. 3) appeared to perform well, and might also be used for the 677 prediction of beef cattle CH₄ production, although this equation was based on fewer observations. 678 679 Nevertheless, dietary starch content was never selected for the Diet_C, Animal_C, Animal_no_DMI_C and Global_C equations, whereas NDF content was, indicating that 680 DMI+STA_C equations may be slightly less robust than Diet_C and Animal_C equations. 681

If dietary forage content is known to be > 25%, we recommend the use of the higher-forage 682 equations, because the RSR and CCC of these equations are lower and higher, respectively, 683 compared to the higher-forage subset evaluation of the all-data equations. Based on their predictive 684 performance, the higher-forage Animal_C and the Escobar-Bahamondes et al. (2017a) equations 685 (Eqns. 17, 11; Table 4) are specifically recommended. Despite its lower precision, the Charmley 686 687 et al. (2016) equation (Eq. 20) will still give an accurate estimate of CH₄ production if only DMI is available. If dietary forage percentage is $\leq 18\%$, we recommend the Ellis *et al.* (2009) equations 688 (Eqns. 26-27; Table 5). In addition, we recommend the lower-forage DMI_C or all-data 689 690 DMI+NDF_C, DMI+STA_C and Animal_C equations that performed relatively well (Eqns. 20, 2-3, 6; Tables 2, 5). If dietary forage content is between 18 and 25%, we suggest an all-data 691 692 equation that includes dietary forage, or dietary NDF or starch, because of the forage content that 693 is commonly related to the latter two carbohydrate fractions. The European higher-forage and North American higher-forage equations performed somewhat better on RSR and CCC than the higher-forage equations for the European higher-forage and North American higher-forage subsets, and less systematic bias was obtained for the region-specific equations. Therefore, we most strongly recommend the Diet_C, Animal_C and Charmley *et al.* (2016) equations (Eq. 32-33, 37; Table 6). For North American higher-forage data, the Global_C and Charmley *et al.* (2016) equations are recommended in particular (Eqns. 45, 47; Table 7).

Models that assumed a fixed Y_m, such as the IPCC and GLOBAL NETWORK Tier 2 700 equations, performed nearly as good as the developed more complex best performing equations in 701 702 most cases. The Tier 2 equations may, therefore, have a high potential for predicting beef cattle CH₄ production as well, in particular for higher-forage diets, although the higher variance along 703 704 the unity lines of the predicted vs. observed plots indicates a lack of precision. Moreover, the 705 substantial mean bias that was obtained for the lower-forage subset in particular emphasizes the importance of an accurate estimate of $Y_{\rm m}$. In cases where dietary forage contents are not close to 706 the means of the present data (sub)sets, we do not recommend the use of the Y_m equations 707 considered in the present study, but an equation that contains dietary forage, NDF or starch. A Y_m 708 of 4.5% that was obtained for lower-forage diets may be fairly accurate given a $Y_{\rm m}$ of 5.2% that 709 710 was reported for 42 treatments means with $\leq 17\%$ forage (Escobar-Bahamondes *et al.*, 2017b), and a $Y_{\rm m}$ of 3.8% for 34 treatments means with $\leq 18\%$ forage (Escobar-Bahamondes *et al.*, 2017a). 711 712 Both of these studies reported 9.5% forage on average for studies collected from multiple 713 continents. These $Y_{\rm m}$ values are all higher than the 3.0%, which the IPCC uses for $\leq 10\%$ forage diets. The $\leq 10\%$ forage records in the present analysis, which also had a $Y_{\rm m}$ of 4.5% suggests that 714 715 the $Y_{\rm m}$ value for lower-forage diets used by the IPCC needs to be reconsidered. However, practices 716 such as feeding steam-flaked corn (Hales et al., 2012) and dietary supplementation with monensin 717 (Appuhamy *et al.*, 2013) may require alternative prediction as these diets may have a $Y_{\rm m}$ value of 3.0%. This also applies to fat supplemented diets (Grainger and Beauchemin, 2011; Patra, 2013). 718 For CH₄ yield predictions, the all-data NDF_C, Diet_no_DMI_C and Global_no_DMI_C 719 720 CH₄ yield equations (Eqns. 54, 57-58) had RSR values < 1 and are suitable for use if dietary forage content is unknown. The Global no DMI C equations may also be used if dietary forage content 721 722 is known to be $\leq 18\%$. The higher-forage Diet_no_DMI_C CH₄ yield equation is the only equation to consider for forage contents > 25% (Eq. 62; Table S4), based on RSR values > 1 for the other 723 higher-forage CH₄ yield equations. For forage contents between 18 and 25% we recommend an 724 725 all-data equation with RSR < 1 and the highest CCC value when evaluated with all data, which is the Global_no_DMI_C yield equation (Eq. 58). Given that all CH₄ intensity equations were 726 associated with an RSR value > 1 for the higher- and lower-forage subsets, we recommend the 727 728 observed average values of 108 and 161 [g (kg ADG)⁻¹] for dietary forage contents of \leq 18% and \geq 25%, respectively. For dietary contents between 18 and 25% or if forage content is unknown we 729 recommend the all-data Global C equation (Eq. 69; Table S5). 730

731

732 **5.** Conclusion

Our analysis is based on the large GLOBAL NETWORK dataset comprising data from several continents and a wide variety of forage contents. As observed previously, DMI is the key factor for predicting beef cattle enteric CH₄ production. Non-linear models with DMI as the only independent variable did not outperform their counterpart linear models. However, linear models depending on DMI and dietary forage content or these two covariates plus BW commonly had an improved predictive ability. Separate equations for lower-forage ($\leq 18\%$) and higher-forage (\geq 25%) data also improved predictive ability. Model evaluation specific to European higher-forage,

North American higher-forage and Brazilian higher-forage diets compared with that of 740 intercontinental higher-forage diet models suggests that overall enteric CH₄ production is more 741 accurately predicted by region-specific models, although in many cases the best intercontinental 742 743 and region-specific models may perform similarly. The equations developed in the present study commonly had higher precision and less prediction error with similar accuracy compared to the 744 extant equations that were evaluated. Evaluation of CH₄ emission conversion factors indicated that 745 region-specific and in particular dietary forage content-based $Y_{\rm m}$ values are required for adequately 746 predicting beef cattle CH₄ production in national or global inventories. 747

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770 **References**

- Aguerre, M.J., Wattiaux, M.A., Powell, J.M., Broderick, G.A., Arndt, C., 2011. Effect of forageto-concentrate ratio in dairy cow diets on emission of methane, carbon dioxide, and
 ammonia, lactation performance, and manure excretion. J. Dairy Sci. 94, 3081-3093.
- 774 <u>https://doi:10.3168/jds.2010-4011</u>
- Appuhamy, J.A.D.R.N., Strathe, A.B., Jayasundara, S., Wagner-Riddle, C., France, J., Dijkstra, J.,
 Kebreab E., 2013. Anti-methanogenic effects of monensin in dairy and beef cattle: A metaanalysis. J. Dairy Sci. 96, 5161-5173. https://doi.org/10.3168/jds.2012-5923
- Bannink, A., Schijndel, M.W. van, Dijkstra, J., 2011. A model of enteric fermentation in dairy 778 cows to estimate methane emission for the Dutch National Inventory Report using the 779 603-618. 780 IPCC Tier 3 approach. Anim. Feed Sci. Technol. 166-167, https://doi.org/10.1016/j.anifeedsci.2011.04.043 781
- 782 Basarab, J.A., Okine, E.K., Baron, V.S., Marx, T., Ramsey, P., Ziegler, K., Lyle, K., 2005.
- 783 Methane emissions from enteric fermentation in Alberta's beef cattle population. Can. J.
- 784 Anim. Sci. 85, 501-512. <u>https://doi.org/10.4141/A04-069</u>
- Beauchemin, K.A., Kreuzer, M., O'Mara, F.O., McAllister, T.A., 2008. Nutritional management
 for enteric methane abatement: a review. Austr. J. Exp. Agric. 48, 21-27.
 https://doi.org/10.1071/EA07199
- Bell, M., Eckard, R., Moate, P.J., Yan, T., 2016. Modelling the effect of diet composition on enteric
- methane emissions across sheep, beef cattle and dairy cows. Animals, 6, 54.
 https://doi.org/10.3390/ani6090054

791	Benchaar, C., Hassanat, F., Martineau, R., Gervais, R., 2015. Linseed oil supplementation to dairy
792	cows fed diets based on red clover silage or corn silage: Effects on methane production,
793	rumen fermentation, nutrient digestibility, N balance, and milk production. J. Dairy Sci.
794	98, 7993-8008. https://doi.org/10.3168/jds.2015-9398

- Bibby, J., Toutenburg, T., 1977. Prediction and Improved Estimation in Linear Models. John 785
 Wiley Sons, Chichester.
- Castelan-Ortega, O.A., Ku-Vera, J.C., Estrada-Flores, J.G., 2014. Modeling methane emissions
 and methane inventories for cattle production systems in Mexico. Atmosfera 27, 185-191.
- 799 https://doi.org/10.1016/S0187-6236(14)71109-9
- 800 Charmley, E., Williams, S.R.O., Moate, P.J., Hegarty, R.S., Herd, R.M., Oddy, V.H., Reyenga, P.,
- Staunton, K.M., Anderson, A., Hannah, M.C. 2016. A universal equation to predict
 methane production of forage-fed cattle in Australia. Anim. Prod. Sci. 56, 169-180.
 https://doi.org/10.1071/AN15365
- Demment, M.W., Van Soest, P.J., 1985. A nutritional explanation for body-size patterns of
 ruminant and non-ruminant herbivores. American Naturalist 125, 641-672.
 https://doi.org/10.1086/284369
- Dijkstra, J., France, J., Assis, A.G., Neal, H.D.St.C., Campos, O.F., Aroeira, L.J.M., 1996.
 Simulation of digestion in cattle fed sugarcane: prediction of nutrient supply for milk
 production with locally available supplements. J. Agric. Sci. 127, 247-260.
 https://doi.org/10.1017/S0021859600078023

- Dijkstra, J., Oenema, O., Bannink, A., 2011. Dietary strategies to reducing N excretion from cattle:
 Implications for methane emissions. Curr. Opin. Environ. Sustain. 3, 414-422.
 https://doi.org/10.1016/j.cosust.2011.07.008
- B14 Drouillard, J.S., 2018. Current situation and future trends for beef production in the United States
- 815 of America A review. Asian-Australas. J. Anim. Sci. 31, 1007-1016.
 816 https://doi.org/10.5713/ajas.18.0428
- 817 Ellis, J.L., Kebreab, E., Odongo, N.E., McBride, B.W., Okine, E.K., France, J., 2007. Prediction
- 818 of methane production from dairy and beef cattle. J. Dairy Sci. 90, 3456-3467.
 819 https://doi.org/10.3168/jds.2006-675
- 820 Ellis, J.L., Kebreab, E., Odongo, N.E., Beauchemin, K.A., McGinn, S., Nkrumah, J.D., Moore,
- 821 S.S., Christopherson, R., Murdoch, G.K., McBride, B.W., Okine, E.K., France, J., 2009.
- 822 Modeling methane production from beef cattle using linear and nonlinear approaches. J.

823 Anim. Sci. 87, 1334-1345. <u>https://doi.org/10.2527/jas.2007-0725</u>

- 824 Escobar-Bahamondes, P., Oba, M., Beauchemin, K.A., 2017a. Universally applicable methane
- prediction equations for beef cattle fed high- or low-forage diets. Can. J. Anim. Sci. 97,
 83-94. https://doi.org/10.1139/cjas-2016-0042
- Escobar-Bahamondes, P., Oba, M., Beauchemin, K.A., 2017b. An evaluation of the accuracy and
 precision of methane prediction equations for beef cattle fed high-forage and high-grain
 diets. Animal 11, 68-77. https://doi.org/10.1017/S175173111600121X
- Gastelen, S. van, Antunes-Fernandes, S.C., Hettinga, K.A., Klop, G., Alferink, S.J.J., Hendriks,
 W.H., Dijkstra, J., 2015. Enteric methane production, rumen volatile fatty acid
 concentrations, and milk fatty acid composition in lactating Holstein-Friesian cows fed

- grass silage- or corn silage-based diets. J. Dairy Sci. 98, 1915-1927.
 https://doi.org/10.3168/jds.2014-8552
- 835 Gerber, P.J., Steinfeld, H., Henderson, B., Mottet, A., Opio, C., Dijkman, J., Falcucci, A., Tempio, G,
- 836 2013. Tackling climate change through livestock A global assessment of emissions and
- 837 *mitigation opportunities*. Food and Agriculture Organization of the United Nations (FAO),
 838 Rome.
- Gerssen-Gondelach, S.J., Lauwerijssen, R.B.G., Havlík, P., Herrero, M., Valin, H., Faaij, A.P.C.,
 Wicke, B., 2017. Intensification pathways for beef and dairy cattle production systems:
- 841 Impacts on GHG emissions, land occupation and land use change. Agric. Ecosyst. Environ.
- 842 240, 135-147. <u>https://doi.org/10.1016/j.agee.2017.02.012</u>
- Goopy, J., Donaldson, A., Hegarty, R., Vercoe, P., Haynes, F., Barnett, M., Oddy, V., 2014. Lowmethane yield sheep have smaller rumens and shorter rumen retention time. Brit. J. Nutr.
- 845 111, 578-585. <u>https://doi.org/10.1017/S0007114513002936</u>
- Grainger, C., Beauchemin, K.A., 2011. Can enteric methane emissions from ruminants be lowered
- without lowering their production? Anim. Feed Sci. Technol. 166-167, 308-320.
 <u>https://doi.org/10.1016/j.anifeedsci.2011.04.021</u>
- Guyader, J., Eugène, M., Nozière, P., Morgavi, D., Doreau, M., Martin, C., 2014. Influence of
 rumen protozoa on methane emission in ruminants: A meta-analysis approach. Animal 8,
- 851 1816-1825. https://doi.org/10.1017/S1751731114001852
- Hammond, K.J., Crompton, L.A., Bannink, A., Dijkstra, J., Yáñez-Ruiz, D.R., O'Kiely, P.,
 Kebreab, E., Eugène, M.A., Yu, Z., Shingfield, K.J., Schwarm, A., Hristov, A.N.,
- 854 Reynolds, C.K., 2016. Review of current *in vivo* measurement techniques for quantifying

- enteric methane emission from ruminants. Anim. Feed Sci. Technol. 219, 13-30.
 https://doi.org/10.1016/j.anifeedsci.2016.05.018
- Hales, K.E., Cole, N.A., MacDonald, J.C., 2012. Effects of corn processing method and dietary
 inclusion of wet distillers grains with solubles on energy metabolism, carbon-nitrogen
 balance, and methane emissions of cattle. J. Anim. Sci. 90, 3174-3185.
 https://doi.org/10.2527/jas.2011-4441
- Hatew, B., Cone, J.W., Pellikaan, W.F., Podesta, S.C., Bannink, A., Hendriks, W.H., Dijkstra, J.,
- 862 2015. Relationship between in vitro and in vivo methane production measured
 863 simultaneously with different dietary starch sources and starch levels in dairy cattle. Anim.
- 864
 Feed Sci. Technol. 202, 20-31. <u>https://doi.org/10.1016/j.anifeedsci.2015.01.012</u>
- 865 Hristov, A.N., Oh, J., Firkins, J., Dijkstra, J., Kebreab, E., Waghorn, G., Makker, M.P.S.,
- Adesogan, A.T., Yang, W., Lee, C., Gerber, P.J., Henderson, B., Tricarico, J.M., 2013a.
- 867 SPECIAL TOPICS Mitigation of methane and nitrous oxide emissions from animal
- 868 operations: I. A review of enteric methane mitigation options. J. Anim. Sci. 91, 5045-5069.
- 869 <u>https://doi.org/10.2527/jas.2013-6583</u>
- 870 Hristov, A.N., Ott, J., Tricarico, J.M., Rotz, A., Waghorn, G., Adesogan, A.T., Dijkstra, J., Montes,
- F., Oh, J., Kebreab, E., Oosting, S.J., Gerber, P.J., Henderson, B., Makker, H.P.S., Firkins,
- J., 2013b. SPECIAL TOPICS Mitigation of methane and nitrous oxide emissions from
- animal operations: III. A review of animal management mitigation options. J. Anim. Sci.
- 874 91, 5095-5113. <u>https://doi.org/10.2527/jas.2013-6585</u>
- Hristov, A.N., Kebreab, E., Niu, M., Oh, J., Bannink, A., Bayat, A.R., Boland, T., Brito, A.F.,
 Casper, D., Crompton, L.A., Dijkstra, J., Eugène, M.A., Garnsworthy, Ph.C., Haque, M.N.,

877	Hellwing, A.L.F., Huhtanen, P., Kreuzer, M., Kuhla, B., Lund, P., Madsen, J., Martin, C.,
878	Moate, P.J., Muetzel, S., Muñoz, C., Peiren, N., Powell, J.M., Reynolds, C.K., Schwarm,
879	A., Shingfield, K.J., Storlien, T.M., Weisbjerg, M.R., Yáñez-Ruiz, D.R., Yu, Z., 2018.
880	Symposium review: Uncertainties in enteric methane inventories, measurement techniques,
881	and prediction models. J. Dairy Sci. 101, 6655-6674. https://doi.org/10.3168/jds.2017-
882	<u>13536</u>

- Huhtanen, P., Ramin, M., Udén, P., 2015. Nordic dairy cow model Karoline in predicting methane
 emissions: 1. Model description and sensitivity analysis. Livest. Sci. 178, 71-80.
 https://doi.org/10.1016/j.livsci.2015.05.009
- Huhtanen, P., Ramin, M., Cabezas-Garcia, E.H., 2016. Effects of ruminal digesta retention time
 on methane emissions: a modelling approach. Anim. Prod. Sci. 56, 501-506.
 https://doi.org/10.1071/AN15507
- 889 IPCC (2006. 2006 IPCC Guidelines for National Greenhouse Gas Inventories. Intergovernmental
- 890 Panel on Climate Change, IGES, Kanagawa, Japan.
- 891 IPCC (2014. 2013 Revised Supplementary Methods and Good Practice Guidance Arising from the
- *Kyoto Protocol*, Hiraishi, T., Krug, T., Tanabe, K., Srivastava, N., Baasansuren, J., Fukuda,
 M., Troxler, T.G. (eds) IPCC, Switzerland.
- James, G., Witten, D., Hastie, T., Tibshirani, R., 2014. *An introduction to statistical learning: With applications in R.* Springer New York, US.
- Johnson, K.A., Johnson, D.E., 1995. Methane emissions from cattle. J. Anim. Sci. 73, 2483-2492.
 https://doi.org/10.2527/1995.7382483x

- Kebreab, E., Johnson, J.A., Archibeque, S.L., Pape, D., Wirth, T., 2008. Model for estimating
 enteric methane emissions from United States dairy and feedlot cattle. J. Anim. Sci. 86,
 2738-2748. https://doi.org/10.2527/jas.2008-0960
- 901 Kutner, M.H., Nachtsheim, C., Neter, J., Li, W., 2005. *Applied linear statistical models*. Boston:
 902 McGraw-Hill Irwin. New York, US.
- Lin, L.I., 1989. A concordance correlation coefficient to evaluate reproducibility. Biometrics, 45,
 255-268. https://doi.org/10.2307/2532051
- 205 Lingen H.J., van, Fadel, J.G., Bannink, A., Dijkstra, J., Tricarico, J., Pacheco, D., Casper, D.P.,
- Kebreab, E., 2018. Multi-criteria evaluation of dairy cattle feed resources and animal
 characteristics for nutritive and environmental impacts. Animal 12, s310-s320.
 https://doi.org/10.1017/S1751731118001313
- 909 Machmüller, A., Kreuzer, M., 1999. Methane suppression by coconut oil and associated effects on
- 910 nutrient and energy balance in sheep. Can. J. Anim. Sci. 79, 65-72.
 911 https://doi.org/10.4141/A98-079
- 912 Maia, M.R.G., Chaudhary, L.C., Figueres, L., Wallace, R.J., 2007. Metabolism of polyunsaturated
- 913 fatty acids and their toxicity to the microflora of the rumen. Anton. Leeuw. 91, 303-314.
 914 https://doi.org/10.1007/s10482-006-9118-2
- Martin, C., Morgavi, D.P., Doreau, M., 2010. Methane mitigation in ruminants: from microbe to
 the farm scale. Animal 4, 351-365. https://doi.org/10.1017/S1751731109990620
- 917 Mills, J.A.N., Kebreab, E., Yates, C.M., Crompton, L.A., Cammell, S.B., Dhanoa, M.S., Agnew,
- 918 R.E., France, J, 2003. Alternative approaches to predicting methane emission from dairy
- 919 cows. J. Anim. Sci. 81, 3141-3150. <u>https://doi.org/10.2527/2003.81123141x</u>

920	Moate, P.J., Williams, S.R.O., Deighton, M.H., Hannah, M.C., Ribaux, B.E., Morris, G.L., Jacobs,
921	J.L., Hill, J., Wales, W.J., 2019. Effects of feeding wheat or corn and of rumen fistulation
922	on milk production and methane emissions of dairy cows. Anim. Prod. Sci. 59, 891-905.
923	https://doi.org/10.1071/AN17433

- Moe, P.W., Tyrrell, H.F., 1979. Methane production in dairy cows. J. Dairy Sci. 62, 1583-1586.
 https://doi.org/10.3168/jds.S0022-0302(79)83465-7
- Moraes, L.E., Strathe, A.B., Fadel, J.G., Casper, D.P., Kebreab, E., 2014. Prediction of enteric
 methane emissions from cattle. Glob. Change Biol. 20, 2140-2148.
 https://doi.org/10.1111/gcb.12471
- Moriasi, D.N., Arnold, J.G., Van Liew, M.W., Bingner, R.L., Harmel, R.D., Veith, T.L., 2007.
 Model evaluation guidelines for systematic quantification of accuracy in watershed
 simulations. T. ASABE 50, 885-900. <u>https://doi.org/10.13031/2013.23153</u>
- Na, Y., Li, D.H., Lee, S.R., 2017. Effects of dietary forage-to-concentrate ratio on nutrient digestibility and enteric methane production in growing goats (*Capra hircus hircus*) and
 Sika deer (*Cervus nippon hortulorum*). Asian-Australas. J. Anim. Sci. 7, 967-972.
 https://doi.org/10.5713/ajas.16.0954
- 936 Niu, M., Kebreab, E., Hristov, A.N., Oh, J., Arndt, C., Bannink, A., Bayat, A.R., Brito, A.F.,
- Boland, T., Casper, D.P., Crompton, L.A., Dijkstra, J., Eugène, M.A., Garnsworthy, Ph.C.,
- Haque, M.N., Hellwing, A.L.F., Huhtanen, P., Kreuzer, M., Kuhla, B., Lund, P., Maden,
- J., Martin, C., McClelland, S.C., McGee, M., Moate, P.J., Muetzel, S., Muñoz, C., O'Kiely,
- 940 P., Peiren, N., Reynolds, C.K., Schwarm, A., Shingfield, K.J., Storlien, T.M., Weisbjerg,
- 941 M.R., Yáñez-Ruiz, D.R., Yu, Z., 2018. Prediction of enteric methane production, yield and

- 942 intensity in dairy cattle using an intercontinental database. Glob. Change Biol. 24, 3368943 3389. <u>https://doi.org/10.1111/gcb.14094</u>
- 944 Opio, C., Gerber, P., Mottet, A., Falcucci, A., Tempio, G., MacLeod, M., Vellinga, T., Henderson,
- B., Steinfeld, H., 2013. Greenhouse gas emissions from ruminant supply chains A global
- 946 life cycle assessment. Food and Agriculture Organization of the United Nations (FAO),947 Rome.
- Patra, A.K., 2013. The effect of dietary fats on methane emissions, and its other effects on
 digestibility, rumen fermentation and lactation performance in cattle: A meta-analysis.
- 950 Livest. Prod. Sci. 155, 244-254. <u>https://doi.org/10.1016/j.livsci.2013.05.023</u>
- Patra, A.K., 2017. Prediction of enteric methane emission from cattle using linear and non-linear
 statistical models in tropical production systems. Mitig. Adapt. Strateg. Glob. Change 22,
 629-650. https://doi.org/10.1007/s11027-015-9691-7
- Pérez-Ramírez, E., Peyraud, J.L., Delagarde, R., 2012. N-alkanes v. ytterbium/faecal index as two 954 methods for estimating herbage intake of dairy cows fed on diets differing in the 955 herbage:maize feeding 956 silage ratio and level. Animal 6, 232-244. https://doi.org/10.1017/S1751731111001480 957
- Pinheiro, J.C., Bates, D.M., 2000. *Mixed-effects models in S and S-PLUS*. Springer Verlag, New
 York, Inc.
- Popp, A., Lotze-Campen, H., Bodirsky, B., 2010. Food consumption, diets shifts and associated
 non-CO₂ greenhouse gases from agricultural production. Glob. Environ. Change 20, 451-
- 962 462. <u>https://doi.org/10.1016/j.gloenvcha.2010.02.001</u>

- 963 R Core Team 2017. R: A language and environment for statistical computing. R Foundation for
 964 Statistical Computing, Vienna, Austria. URL <u>https://www.R-project.org/</u>.
- Richmond, A.S., Wylie, A.R.G., Laidlaw, A.S., Lively, F.O., 2015. Methane emissions from beef
 cattle grazing on semi-natural upland and improved lowland grasslands. Animal 9, 130-
- 967 137. https://doi.org/10.1017/\$1751731114002067
- 968 Santiago-Suarez, B., Moraes, L.E., Appuhamy, J.A.D.R.N., Pellikaan, W.F., Casper, D.P.,
- 969 Tricarico, J., Kebreab, E., 2016. Prediction and evaluation of enteric methane emissions
- 970 from lactating dairy cows using different levels of covariate information. Anim. Prod. Sci. 56,
- 971 557-564. <u>https://doi.org/10.1071/AN15496</u>
- Sauvant, D., Giger-Reverdin, S., 2009. Modélisation des interactions digestives et de la production de
 méthane chez les ruminants. (In French.) INRA Prod. Anim. 22, 375-384.
- Sauvant, D., Nozière, P., 2016. Quantification of the main digestive processes in ruminants: the
 equations involved in the renewed energy and protein feed evaluation systems. Animal 10,
- 976 755-770. <u>https://doi.org/10.1017/S1751731115002670</u>
- 977 St-Pierre N.R., 2001. Invited Review: Integrating quantitative findings from multiple studies using
 978 mixed model methodology. J. Dairy Sci., 84, 741-755. <u>https://doi.org/10.3168/jds.S0022-</u>
 979 0302(01)74530-4
- Sutter, F., Schwarm, A., Kreuzer, M., 2017. Development of nitrogen and methane losses in the
 first eight weeks of lactation in Holstein cows subjected to deficiency of utilisable crude
 protein under restrictive feeding conditions. Arch. Anim. Nutr. 71, 1-20.
- 983 <u>https://doi.org/10.1080/1745039X.2016.1258880</u>

984	Vries, M. de, Middelaar, C.E. van, Boer, I.J.M. de, 2015. Comparing environmental impacts of
985	beef production systems: A review of life cycle assessments. Livest. Sci. 178, 279-288.
986	https://doi.org/10.1016/j.livsci.2015.06.020

- Warner, D., Hatew, B., Podesta, S.C., Klop, G., van Gastelen, S., van Laar, H., Dijkstra, J., 987 Bannink, A., 2016. Effects of nitrogen fertilisation rate and maturity of grass silage on 988 methane emission by lactating dairy Animal 10. 34-43. 989 cows. 990 https://doi.org/10.1017/S1751731115001640
- 991 White, R.R., Roman-Garcia, Y., Firkins, J.L., VanDeHaar, M.J., Armentano, L.E., Weiss, W.P.,
- 992 McGill, T., Garnett, R, Hanigan, M.D., 2017. Evaluation of the National Research Council
- 993 (2001) dairy model and derivation of new prediction equations. 1. Digestibility of fiber,
 994 fat, protein, and non-fiber carbohydrate. J. Dairy Sci. 100, 3591-3610.
- 995 <u>https://doi.org/10.3168/jds.2015-10800</u>
- 996 Yan, T., Agnew, R.E., Gordon, F.J., Porter, M.G., 2000. Prediction of methane energy output in
- 997 dairy and beef cattle offered grass silage-based diets. Livest. Prod. Sci. 64, 253-263.
 998 https://doi.org/10.1016/S0301-6226(99)00145-1
- Yan, T., Porter, M.G., Mayne, C.S., 2009. Prediction of methane emission from beef cattle using
 data measured in indirect open-circuit respiration calorimeters, Animal 3, 1455-1462.
 <u>https://doi.org/10.1017/S175173110900473X</u>
- 1002 Zuur, A.F., Ieno, E.N., Elphick, C.S., 2010. A protocol for data exploration to avoid common
- 1003
 statistical problems. Methods Ecol. Evol. 1, 3-14. https://doi.org/10.1111/j.2041-1004

 1004
 210X.2009.00001.x

1005 Zwillinger, D., Kokoska, S., 2000. CRC Standard Probability and Statistics Tables and Formulae,

1006 CRC Press, Boca Raton, US.

		All data (n = 1021)	Hig	gher-fora	ge ($n = 8$	82)	Lo	wer-forag	ge $(n = 1)$	39)		EUR-HF	(n = 307))	Ν	JrAm-HF	r (n = 394	.)
Item*	Mean	Min	Max	SD	Mean	Min	Max	SD	Mean	Min	Max	SD	Mean	Min	Max	SD	Mean	Min	Max	SD
DMI (kg d ⁻¹)	8.13	2.26	17.5	2.82	7.81	2.26	17.5	2.86	10.1	4.77	15.6	1.89	9.46	3.17	15.7	2.12	6.16	2.26	14.1	2.51
GEI (MJ d ⁻¹)	150	42.6	317	53.7	144	42.6	317	53.1	191	88.2	300	37.1	177	57.8	299	42.0	114	42.6	254	45.4
Diet composition (% of DM	()																		
СР	14.6	6.19	22.5	2.56	14.6	6.19	21.3	2.60	14.6	11.4	22.5	2.35	14.3	7.80	19.2	1.84	15.6	10.0	21.3	2.52
EE	3.02	0.372	7.02	1.20	2.87	0.372	7.02	1.05	3.90	0.377	5.63	1.57	3.65	0.372	5.80	1.40	2.58	0.669	5.50	0.825
Ash	6.29	3.22	13.7	2.11	6.52	3.22	13.7	2.13	4.85	3.50	8.00	1.24	6.06	3.40	11.4	1.80	6.51	3.22	13.7	2.29
NDF	35.0	17.2	73.9	11.2	36.6	17.2	73.9	11.2	24.7	19.8	33.3	3.12	37.5	26.1	68.4	7.25	32.9	17.5	67.8	9.76
ADF	19.3	6.92	50.8	8.13	20.6	7.50	50.8	8.06	11.6	6.92	14.5	1.99	21.7	14.0	40.3	5.15	17.5	7.50	36.5	7.28
STA	34.0	2.50	64.1	13.6	32.2	2.50	64.1	14.3	42.0	32.0	56.9	4.52	25.3	2.50	40.3	10.3	41.2	16.8	64.1	12.6
For	51.0	8.0	100	27.7	57.7	25.0	100	24.0	9.8	8.0	18.1	3.18	64.6	31.0	100	16.9	47.9	25.0	100	23.6
ADG (kg d ⁻¹)	1.25	0.060	3.38	0.431	1.19	0.060	3.38	0.438	1.46	0.552	2.22	0.330	1.22	0.088	1.99	0.348	NA	NA	NA	NA
BW (kg)	478	133	791	148	454	133	791	144	625	376	734	76.5	571	133	791	128	391	196	699	116
Methane emissions	3																			
CH4 (g d ⁻¹)	161	37.0	372	70.5	162	37.0	372	73.5	153	45.1	310	46.4	215	40.9	372	71.8	125	37.0	313	57.5
CH4/DMI (g kg ⁻¹)	20.0	6.29	35.1	5.05	20.7	6.29	35.1	4.75	15.2	7.50	30.9	4.29	22.5	6.64	35.1	5.19	20.3	6.29	33.3	4.21
CH4/ADG (g kg ⁻																				
1) J	4.98	3.31	6.68	0.522	5.08	3.31	6.68	0.528	4.68	3.84	5.71	0.362	5.20	3.31	6.68	0.517	NA	NA	NA	NA
$Y_{\rm m}$ (% of GEI)§	6.0	1.9	10.4	1.5	6.3	1.9	10.4	1.4	4.5	2.3	8.7	1.2	6.7	2.0	10.3	1.5	6.2	1.9	10.4	1.3

1008 \leq 18%), European higher-forage (EUR-HF) and North American higher-forage (NrAm-HF) entries of the GLOBAL NETWORK beef cattle database.

Table 1. Variable summary statistics for all data, higher-forage (data associated with a forage content $\geq 25\%$), lower-forage (data associated with a forage content

1009 *DM = dry matter, DMI = dry matter intake, GEI = gross energy intake, CP = dietary crude protein, EE = dietary ether extract, NDF = dietary neutral detergent fiber,

1010 ADF = dietary acid detergent fiber, STA = dietary starch, For = dietary forage, ADG = average daily body weight gain, BW = body weight.

1011 ‡Min = minimum, Max = maximum, SD = standard deviation.

1012 *I* In transformed values.

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1013 §Methane conversion factor (%): energy of CH_4 as a proportion of GEI; the specific energy of CH_4 is 55.65 MJ kg⁻¹.

Table 2. All-data CH₄ emission (g d⁻¹ animal⁻¹) prediction equations for various categories and model performance across the data (sub)sets
based on root mean square prediction error (RMSPE; % of mean), RMSPE-observations-standard-deviation-ratio (RSR), mean and slope bias (MB
and SB; % of mean square prediction error), and concordance correlation coefficient (CCC).

	Ν	Iodel development				Model perfo	ormance	;		
Eq.	Category§	Prediction equation*	n^{\dagger}	(Sub)set‡	p^{\ddagger}	RMSPE,	RSR	MB,	SB,	CCC
						%		%	%	
[1]	DMI_C	54.2 (7.6) + 12.6 (0.6) × DMI	1021	All-data	991	31.2	0.71	0.69	12.55	0.60
				Higher-forage	852	30.8	0.67	3.86	27.26	0.64
				Lower-forage	139	33.9	1.12	34.96	0.59	0.27
[2]	DMI+NDF_C	-16.4 (9.0) + 12.1 (0.6) × DMI +	1021	All-data	991	31.4	0.71	0.92	1.99	0.63
		$2.10(0.16) \times \text{NDF}$		Higher-forage	852	31.8	0.69	1.82	2.81	0.65
				Lower-forage	139	28.7	0.94	3.45	2.10	0.32
[3]	DMI+STA_C	126 (11) + 11.5 (0.9) × DMI –	704	All-data	704	28.9	0.71	6.09	1.01	0.65
		1.75 (0.16) × STA		Higher-forage	575	28.7	0.70	13.87	1.80	0.68
				Lower-forage	129	30.1	0.96	15.77	0.64	0.35
[4]	DMI+EE_C	$83.0(9.8) + 11.9(0.6) \times DMI -$	754	All-data	754	29.4	0.71	1.35	8.83	0.61
		7.31 (1.69) × EE		Higher-forage	644	29.2	0.67	0.08	19.59	0.64
				Lower-forage	110	30.3	1.21	37.59	4.20	0.25
[5]	Diet_C	$-0.767(7.493) + 12.0(0.5) \times DMI$	1021	All-data	991	29.5	0.67	2.39	1.24	0.70
		+ 1.12 (0.06) × For		Higher-forage	852	29.5	0.64	1.30	2.12	0.72
				Lower-forage	139	29.8	0.98	17.57	0.30	0.32
[6]	Animal_C, Global_C		1003	All-data	991	26.9	0.61	2.20	1.37	0.76
				All-data √ ,♯	646	22.5	0.52	3.05	0.41	0.84

		$-28.3(8.3) + 10.3(0.6) \times DMI +$		Higher-forage	852	26.8	0.58	2.00	1.83	0.78
		$1.12 (0.06) \times For + 0.0885$		Lower-forage	139	27.8	0.91	3.72	0.65	0.35
		$(0.0150) \times BW$								
[7]	Animal_no_DMI_C	6.03 (10.40) + 1.25 (0.07) × For –	992	All-data	991	30.9	0.70	0.55	1.98	0.65
		2.29 (0.77) × Ash + 0.212 (0.015)		Higher-forage	852	30.9	0.67	0.23	3.04	0.68
		imes BW		Lower-forage	139	31.3	1.03	5.86	3.26	0.11
[8]	GLOBAL	$[0.061 \ (0.001) \times GEI] \ / \ 0.05565$	1021	All-data	991	28.5	0.64	0.59	0.27	0.75
	NETWORK Tier 2			All-data ♪ , ↓	991	28.3	0.64	0.23	0.87	0.76
[9]	IPCC Tier 2 (2006) ¶	$(0.065 \times \text{GEI}) / 0.05565$	-	All-data 	991	29.9	0.68	7.97	3.04	0.75
[10]	Charmley et al. (2016)	$-6.10 + 20.6 \times DMI$		All-data ♪ ,∥	939	28.9	0.66	0.00	1.15	0.74
[11]	Escobar-Bahamondes et al. (2017a)	$-35.0 + 0.08 \times BW + 1.2 \times For -$ $69.8 \times EEI^3 + 3.14 \times GEI$		All-data ,#	646	23.2	0.54	11.66	2.12	0.85

1017 § Category acronyms (*e.g.*, DMI_C) are explained in the 'Model development' subsection of the 'Methods and Materials' section.

1018 * Equations are presented with regression coefficient standard errors in parenthesis; DMI = dry matter intake (kg d⁻¹), NDF = dietary neutral

1019 detergent fiber (% of DM), STA = dietary starch (% of DM), EE = dietary ether extract (% of DM), Ash = dietary ash (% of DM), For = dietary

1020 forage (% of DM), BW = body weight (kg), GEI = gross energy intake (MJ d⁻¹), EEI = ether extract intake (kg d⁻¹).

- 1021 $\dagger n$ = number of observations used to fit model equations
- 1022 \ddagger All-data = all data collected for analysis, Higher-forage = data associated with a forage content \ge 25%, Lower-forage = data associated with a
- 1023 forage content $\leq 18\%$.
- 1024 $\ddagger p$ = numbers of observations used for model evaluation.
- 1025 ¶IPCC = Intergovernmental Panel on Climate Change.
- 1026 Performance was evaluated, not cross-validated.

- 1027 *INo* independent evaluation.
- 1028 The 991 data points minus data from Tomkins *et al.* (2011) and Kennedy and Charmley (2012) to ensure independent evaluation.
- 1029 #The 991 data points minus data from Pinares-Patiño et al. (2003), Chaves et al. (2006), McGeough et al. (2010ab), Doreau et al. (2011), Staerfl
- 1030 *et al.* (2012), Hünerberg *et al.* (2013ab) and Troy *et al.* (2015) to ensure independent evaluation.

Table 3. Root mean square prediction error-standard-deviation-ratio (RSR) of the DMI_C CH₄ production (g d⁻¹) equations based on lower-forage (\leq forage content cutoff) and higher-forage (> forage content cutoff) subsets, their number of observations (*n*), and the average RSR weighted to the number of higher- and lower-forage observations (All) for various diet forage content cutoff values to split the entire dataset into lower-forage and higher-forage subsets.

Forage content cutoff (% of DM) (Sub)set 0 15 20 25 30 35 40 45 50 Higher-forage NA 0.65 0.64 0.65 0.68 0.68 0.69 0.70 0.62 474 NA 902 882 783 672 664 602 579 п Lower-forage NA 0.95 0.94 0.78 0.69 0.70 0.68 0.68 0.73 NA 119 139 238 349 357 419 547 442 п All 0.71 0.68 0.68 0.68 0.69 0.69 0.69 0.68 0.69

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1036	Table 4. Higher-forage CH_4 emission (g d ⁻¹ animal ⁻¹) prediction equations for various categories and model performance across the data subsets
1037	based on root mean square prediction error (RMSPE; % of mean), RMSPE-observations-standard-deviation-ratio (RSR), mean and slope bias (MB
1038	and SB; % of mean square prediction error), and concordance correlation coefficient (CCC).

		Model development	Model performance								
Eq.	Category§	Prediction equation*	n^{\dagger}	(Sub)set‡	p^{\ddagger}	RMSPE,	RSR	MB,	SB,	CCC	
						%		%	%		
[12]	DMI_C	52.8 (7.5) + 13.8 (0.6) × DMI	882	Higher-forage	852	29.3	0.64	0.28	22.78	0.68	
				EUR-HF	307	29.3	0.88	26.42	15.91	0.43	
				NrAm-HF	394	27.3	0.59	16.37	27.26	0.75	
				BRZ-HF	75	26.1	1.40	31.70	22.96	0.23	
[13]	DMI+NDF_C	23.8 (9.1) + 13.5 (0.6) × DMI +	882	Higher-forage	852	29.1	0.64	0.20	17.83	0.69	
		0.844 (0.165) × NDF		EUR-HF	307	29.0	0.87	27.08	14.11	0.45	
				NrAm-HF	394	25.2	0.55	13.46	21.70	0.80	
				BRZ-HF	75	24.9	1.34	24.24	26.05	0.25	
[14]	DMI+STA_C	$83.4(11.4) + 13.6(0.8) \times DMI -$	575	Higher-forage	575	26.6	0.65	1.42	16.37	0.68	
		$0.594 (0.161) \times STA$		EUR-HF	273	26.0	0.93	21.41	4.37	0.35	
				NrAm-HF	269	24.7	0.56	16.08	20.02	0.78	
				BRZ-HF	14	29.2	2.47	77.05	9.49	-0.04	
[15]	DMI+EE_C	66.4 (9.5) + 13.3 (0.6) × DMI –	644	Higher-forage	644	27.8	0.64	1.32	15.45	0.69	
		3.69 (1.56) × EE		EUR-HF	122	26.1	1.00	43.22	6.45	0.39	
				NrAm-HF	394	28.4	0.61	18.29	28.79	0.73	
				BRZ-HF	104	24.0	1.33	30.48	19.06	0.23	
[16]	Diet_C		882	Higher-forage	852	27.9	0.61	0.49	15.52	0.72	

		23.4 (8.1) + 13.2 (0.5) × DMI +		EUR-HF	307	27.9	0.84	25.38	12.04	0.49
		$0.571 (0.080) \times For$		NrAm-HF	394	23.1	0.50	8.40	17.91	0.83
				BRZ-HF	75	22.6	1.21	14.48	25.25	0.28
[17]	Animal_C, Global_C	-6.41 (8.31) + 11.3 (0.6) × DMI +	864	Higher-forage	852	24.6	0.54	0.80	14.74	0.80
		$0.557 (0.077) \times For + 0.0996$		Higher-forage♯	567	21.2	0.47	0.11	10.43	0.86
		$(0.0142) \times BW$		EUR-HF	307	24.5	0.73	18.25	11.11	0.61
				NrAm-HF	394	20.3	0.44	3.82	11.92	0.88
				BRZ-HF	75	21.2	1.14	1.36	27.11	0.22
[18]	Animal_no_DMI_C	17.9 (10.4) + 0.732 (0.091) × For +	864	Higher-forage	852	30.8	0.67	0.01	13.23	0.65
		$0.226 (0.015) \times BW$		EUR-HF	307	26.3	0.79	14.80	11.89	0.52
				NrAm-HF	394	33.1	0.72	18.90	5.82	0.65
				BRZ-HF	75	27.0	1.45	31.74	25.94	-0.14
[19]	GLOBAL	$[0.063 (0.002) \times \text{GEI}] / 0.05565$	882	Higher-forage	852	24.6	0.54	2.12	1.37	0.82
	NETWORK Tier 2			Higher-forage ♪ , ↓	852	24.0	0.52	0.14	0.62	0.83
[9]	IPCC Tier 2 (2006) ¶	$(0.065 \times GEI) / 0.05565$	-	Higher-forage.	852	24.1	0.53	1.19	0.06	0.84
[20]	Charmley <i>et al</i> . (2016)	$21.0 \times DMI$		Higher-forage ♪,∥	829	25.4	0.57	0.10	0.05	0.81
[11]	Escobar-Bahamondes et al. (2017a)	$\begin{array}{c} -35.0+0.08\times BW+1.2\times For-\\ 69.8\times EEI^{3}+3.14\times GEI \end{array}$		Higher-forage √ ,#	567	23.1	0.51	9.61	1.56	0.86

1039 § Category acronyms (*e.g.*, DMI_C) are explained in the 'Model development' subsection of the 'Methods and Materials' section.

1040 * Equations are presented with regression coefficient standard errors in parenthesis; DMI = dry matter intake (kg d⁻¹), NDF = dietary neutral

1041 detergent fiber (% of DM), STA = dietary starch (% of DM), EE = dietary ether extract (% of DM), For = dietary forage (% of DM), BW = body

1042 weight (kg), GEI = gross energy intake (MJ d⁻¹), EEI = ether extract intake (kg d⁻¹).

1043 $\dagger n$ = number of observations used to fit model equations.

- 1044 \ddagger Higher-forage = data associated with a forage content $\ge 25\%$, EUR-HF = European data associated with a forage content $\ge 25\%$, NrAm-HF =
- 1045 North American data associated with a forage content \geq 25%, BRZ-HF = Brazilian data associated with a forage content \geq 25%.
- 1046 $\ddagger p =$ numbers of observations used for model evaluation.
- 1047 ¶IPCC = Intergovernmental Panel on Climate Change.
- 1048 Performance was evaluated, not cross-validated.
- 1049 **J**No independent evaluation.
- 1050 The 852 data points minus data from Tomkins *et al.* (2011) and Kennedy and Charmley (2012) to ensure independent evaluation.
- 1051 #The 852 data points minus data from Pinares-Patiño et al. (2003), Chaves et al. (2006), McGeough et al. (2010ab), Doreau et al. (2011), Staerfl
- 1052 *et al.* (2012), Hünerberg *et al.* (2013ab) and Troy *et al.* (2015) to ensure independent evaluation.

Table 5. Lower-forage CH₄ emission (g d⁻¹ animal⁻¹) prediction equations and model performance using the lower-forage subset based on root
 mean square prediction error (RMSPE; % of mean), RMSPE-observations-standard-deviation-ratio (RSR), mean and slope bias (MB and SB; % of
 mean square prediction error), and concordance correlation coefficient (CCC).

		Model development		Model performance									
Eq.	Category§	Prediction equation*	n^{\dagger}	(Sub)set‡	p^{\ddagger}	RMSPE,	RSR	MB,	SB,	CCC			
						%		%	%				
[20]	DMI_C, Diet_C,	46.6 (19.4) + 9.54 (1.80) × DMI	139	Lower-forage	139	28.4	0.94	4.81	0.03	0.26			
	Animal_C,												
	Global_C												
[21]	DMI+NDF_C	112 (47) + 9.46 (1.79) × DMI – 2.58	139	Lower-forage	139	29.3	0.96	4.74	1.34	0.25			
		$(1.72) \times \text{NDF}$											
[22]	DMI+STA_C	$42.0(38.9) + 9.85(1.88) \times DMI +$	129	Lower-forage	129	34.5	1.11	3.70	20.18	0.23			
		0.0331 (0.7546) × STA											
[23]	DMI+EE_C	$57.0(18.1) + 8.84(1.74) \times DMI -$	110	Lower-forage	110	24.1	0.96	4.54	1.22	0.26			
		1.17 (2.03) × EE											
[24]	GLOBAL	$[0.045 \ (0.002) \times GEI] \ / \ 0.05565$	139	Lower-forage	139	27.9	0.92	3.13	3.02	0.39			
	NETWORK Tier 2			Lower-forage ∫ , ∫	139	27.3	0.90	0.47	3.99	0.43			
				Lower-forage ♪ ,∥	101	25.2	0.98	2.27	12.91	0.41			
[25]	IPCC Tier 2 (2006)	(0.030 × GEI) / 0.05565	-	Lower-forage♪	139	42.1	1.38	59.60	0.08	0.17			
	Lower-forage¶			Lower-forage ∫ ,∥	101	39.0	1.51	64.08	0.35	0.16			
[26]	Ellis et al. (2009);	$48.2 + 14.1 \times DMI - 20.5 \times$		Lower-forage♪	129	27.8	0.89	0.26	2.04	0.41			
[27]	Eq. N Ellis at al. (2009) :	(STA/NDF)		Lower forage	130	27.0	0.02	6 10	0.15	0.34			
[27]	Ems er ur. (2009),	$41.2 \pm 12.0 \times DIVII$		Lower-Iorage	139	21.7	0.92	0.19	0.15	0.54			
	Eq. A												

- 1056 §Category acronyms (*e.g.*, DMI_C) are explained in the 'Model development' subsection of the 'Methods and Materials' section; no
- 1057 Animal_no_DMI_C equation available.
- 1058 *Equations are presented with regression coefficient standard errors in parenthesis; DMI = dry matter intake (kg d⁻¹), NDF = dietary neutral
- 1059 detergent fiber (% of DM), STA = dietary starch (% of DM), EE = dietary ether extract (% of DM), GEI = gross energy intake (MJ d⁻¹).
- 1060 $\dagger n$ = number of observations used to fit model equations.
- 1061 \ddagger Lower-forage = data associated with a forage content $\le 18\%$.
- 1062 $\ddagger p =$ numbers of observations used for model evaluation.
- 1063 ¶IPCC = Intergovernmental Panel on Climate Change.
- 1064 Performance was evaluated, not cross-validated.
- 1065 *INo* independent evaluation.
- 1066 ||A subset containing $\leq 10\%$ forage records only was used (as recommended by the IPCC, 2006)

1067Table 6. European higher-forage CH_4 emission (g d⁻¹ animal⁻¹) prediction equations for various categories and model performance based on root1068mean square prediction error (RMSPE; % of mean), RMSPE-observations-standard-deviation-ratio (RSR), mean and slope bias (MB and SB; % of1069mean square prediction error), and concordance correlation coefficient (CCC).

		Model development	Model performance								
Eq.	Category§	Prediction equation*	n^{\dagger}	(Sub)set‡	p^{\ddagger}	RMSPE,	RSR	MB,	SB,	CCC	
						%		%	%		
[28]	DMI_C	60.5 (16.4) + 15.0 (1.4) × DMI	307	EUR-HF	307	26.3	0.79	4.86	15.69	0.48	
[29]	DMI+NDF_C	38.1 (23.3) + 14.9 (1.4) × DMI + 0.598	307	EUR-HF	307	25.9	0.77	5.21	13.15	0.51	
		$(0.470) \times \text{NDF}$									
[30]	DMI+STA_C	92.4 (21.7) + 11.7 (2.0) × DMI + 0.113	273	EUR-HF	273	25.6	0.92	9.92	2.20	0.30	
		$(0.285) \times STA$									
[31]	DMI+EE_C	133 (34) + 14.5 (2.0) × DMI – 18.4 (6.6)	122	EUR-HF	122	23.7	0.91	1.86	13.87	0.54	
		\times EE									
[32]	Diet_C	$-20.9 (43.6) + 14.3 (2.0) \times DMI + 4.04$	122	EUR-HF	122	18.4	0.70	5.57	0.97	0.70	
		$(1.06) \times NDF - 15.4 (3.8) \times EE$									
[33]	Animal_C	$-102 (40.5) + 11.6 (2.1) \times DMI + 3.74$	122	EUR-HF	122	16.7	0.64	3.49	0.00	0.75	
		$(0.79) \times NDF - 11.1 (3.0) \times EE + 0.164$		EUR_HF♯	109	15.5	0.58	1.31	0.13	0.79	
		$(0.054) \times BW$									
[34]	Animal_no_DMI_C	34.1 (18.7) + 0.287 (0.028) × BW	307	EUR-HF	307	27.0	0.81	10.37	3.73	0.50	
[35]	Global_C	24.3 (17.7) + 9.37 (2.06) × DMI + 0.153	307	EUR-HF	307	24.5	0.73	8.88	11.91	0.58	
		$(0.040) \times BW$									
[36]	GLOBAL	[0.066~(0.003) imes GEI] / 0.05565	307	EUR-HF	307	22.9	0.69	7.03	2.04	0.68	
	NETWORK Tier 2			EUR-HF♪,,	307	21.8	0.65	1.89	1.34	0.71	

[9]	IPCC Tier 2, 2006¶	$(0.065 \times \text{GEI}) / 0.05565$	-	EUR-HF ♪	307	22.0	0.66	3.54	1.61	0.71
[37]	Charmley et al.	$-15.3 + 24.7 \times DMI$		EUR-HF ♪	307	21.9	0.66	0.61	0.16	0.72
	(2016)				122	19.7	0.75	2.09	3.18	0.66
[11]	Escobar-	$-35.0 + 0.08 \times BW + 1.2 \times For - 69.8 \times$		EUR-HF ♪ ,♯	109	16.0	0.60	0.61	0.37	0.77
	Bahamondes et al.	$EEI^3 + 3.14 \times GEI$								
	(2017a)									

1070 §Category acronyms (*e.g.*, DMI_C) are explained in the 'Model development' subsection of the 'Methods and Materials' section.

1071 *Equations are presented with regression coefficient standard errors in parenthesis; DMI = dry matter intake (kg d⁻¹), NDF = dietary neutral

1072 detergent fiber (% of DM), STA = dietary starch (% of DM), EE = dietary ether extract (% of DM), BW = body weight (kg), GEI = gross energy

- 1073 intake (MJ d^{-1}), EEI = ether extract intake (kg d^{-1}).
- 1074 $\dagger n$ = number of observations used to fit model equations.
- 1075 \ddagger EUR-HF = European data associated with a forage content $\ge 25\%$.
- 1076 $\ddagger p$ = numbers of observations used for model evaluation.
- 1077 ¶IPCC = Intergovernmental Panel on Climate Change.
- 1078 Performance was evaluated, not cross-validated.
- 1079 *INo* independent evaluation.

- 1080 # The 307 data points minus data from Pinares-Patiño et al. (2003), McGeough et al. (2010ab), Doreau et al. (2011), Staerfl et al. (2012) and Troy
- *et al.* (2015) to ensure independent evaluation.

Table 7. North American higher-forage CH₄ emission (g d⁻¹ animal⁻¹) prediction equations for various categories and model performance based 1082 1083 on root mean square prediction error (RMSPE; % of mean), RMSPE-observations-standard-deviation-ratio (RSR), mean and slope bias (MB and ediction error) and concordance correlation coefficient (CCC). CD, 0/af.....

1084	SB; % of mean	square prediction	n error), and	l concordance	correlation	coefficien
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	Model development			Model performance						
Eq.	Category§	Prediction equation*	n^{\dagger}	(Sub)set‡	p^{\ddagger}	RMSPE,	RSR	MB,	SB,	CCC
						%		%	%	
[38]	DMI_C	33.9 (7.7) + 14.7 (0.6) × DMI	394	NrAm-HF	394	25.3	0.55	0.00	26.45	0.78
[39]	DMI+NDF_C	1.58 (8.30) + 14.2 (0.6) × DMI + 1.05	394	NrAm-HF	394	23.8	0.52	0.02	16.29	0.82
		$(0.16) \times \text{NDF}$								
[40]	DMI+STA_C	$89.7 (10.7) + 14.2 (0.8) \times DMI - 1.17$	269	NrAm-HF	269	21.4	0.49	1.90	9.73	0.84
		$(0.17) \times STA$								
[41]	DMI+EE_C	43.7 (8.8) + 14.7 (0.6) × DMI – 3.72	394	NrAm-HF	394	25.0	0.54	0.01	25.47	0.78
		(1.52) × EE								
[42]	Diet_C	7.41 (7.12) + 14.1 (0.6) × DMI + 0.632	394	NrAm-HF	394	22.2	0.48	0.01	11.10	0.85
		$(0.069) \times \text{For}$								
[43]	Animal_C,	$-15.1(7.5) + 12.7(0.6) \times DMI + 0.644$	394	NrAm-HF	394	20.1	0.43	0.02	6.12	0.88
		$(0.066) \times \text{For} + 0.0779 \ (0.0134) \times \text{BW}$								
[44]	Animal_no_DMI_C	$14.0(12.3) + 0.965(0.104) \times For +$	394	NrAm-HF	394	32.2	0.70	0.32	5.45	0.63
		$0.207 (0.018) \times BW - 3.02 (0.95) \times Ash$								
[45]	Global_C	-38.8 (10.9) + 12.7 (0.6) × DMI + 0.605	394	NrAm-HF	394	20.0	0.43	0.04	3.03	0.89
		$(0.066) \times \text{For} + 1.61 \ (0.56) \times \text{CP} +$								
		0.0779 (0.0133) × BW								
[46]	GLOBAL	[0.063 (0.003) × GEI] / 0.05565	394	NrAm-HF	394	21.9	0.48	0.81	0.03	0.87
	NETWORK Tier 2			NrAm-HF∫. √	394	21.4	0.46	2.51	0.02	0.88
[9]	IPCC Tier 2, 2006¶	$(0.065 \times \text{GEI}) / 0.05565$	-	NrAm-HF♪	394	22.2	0.48	8.70	0.50	0.88
[47]	Charmley <i>et al</i> .	$20.5 \times \text{DMI}$		NrAm-HF♪	394	20.0	0.43	0.45	0.02	0.90
[.,]	(2016)				271	20.0	0.10	0.10	0.02	0.00

§Category acronyms (e.g., DMI_C) are explained in the 'Model development' subsection of the 'Methods and Materials' section. 1085

- 1086 *Equations are presented with regression coefficient standard errors in parenthesis; DMI = dry matter intake (kg d⁻¹), NDF = dietary neutral
- 1087 detergent fiber (% of DM), STA = dietary starch (% of DM), EE = dietary ether extract (% of DM), For = dietary forage (% of DM), GEI = gross
- 1088 energy intake (MJ d^{-1}), BW = body weight (kg).
- 1089 $\dagger n$ = number of observations used to fit model equations.
- 1090 \ddagger NrAm-HF = North American data associated with a forage content $\ge 25\%$.
- 1091 $\ddagger p$ = numbers of observations used for model evaluation.
- 1092 ¶IPCC = Intergovernmental Panel on Climate Change.
- 1093 Performance was evaluated, not cross-validated.
- 1094 *INo independent evaluation.*

Figure 1. Observed vs. predicted plots for all-data methane emission (g d⁻¹ animal⁻¹) prediction 1095 equations for the different categories, viz., dry matter intake (DMI_C), dry matter intake and 1096 neutral detergent fiber (DMI+NDF C), dry matter intake and starch (DMI+STA C), dry matter 1097 intake and ether extract (DMI+EE C), diet (Diet C), animal (Animal C), animal without DMI 1098 (Animal no DMI C), GLOBAL NETWORK Tier 2, IPCC Tier 2 (2006), and the extant 1099 1100 Charmley et al. (2016) and Escobar-Bahamondes et al. (2017) equations. The gray and black solid lines represent the fitted regression line for the relationship between observed and predicted 1101 values, and the identity line (y = x), respectively. 1102

1103

1104 Figure 2. Observed *vs.* predicted plots for higher-forage methane emission (g d^{-1} animal⁻¹)

1105 prediction equations for the different categories, *viz.*, dry matter intake (DMI_C), dry matter

1106 intake and neutral detergent fiber (DMI+NDF_C), dry matter intake and starch (DMI+STA_C),

1107 dry matter intake and ether extract (DMI+EE_C), diet (Diet_C), animal (Animal_C), animal

1108 without DMI (Animal_no_DMI_C), GLOBAL NETWORK Tier 2, IPCC Tier 2 (2006), and the

1109 extant Charmley et al. (2016) and Escobar-Bahamondes et al. (2017) equations. The gray and

1110 black solid lines represent the fitted regression line for the relationship between observed and

1111 predicted values, and the identity line (y = x), respectively.

1112

1113 Figure 3. Observed vs. predicted plots for lower-forage methane emission (g d^{-1} animal⁻¹)

1114 prediction equations for the different categories, *viz.*, dry matter intake (DMI_C), dry matter

1115 intake and neutral detergent fiber (DMI+NDF_C), dry matter intake and starch (DMI+STA_C),

1116 dry matter intake and ether extract (DMI+EE_C), GLOBAL NETWORK Tier 2, IPCC Tier 2

1117 (2006), and the extant Ellis *et al.* (2009) equations. The gray and black solid lines represent the

62

1118 fitted regression line for the relationship between observed and predicted values, and the identity 1119 line (y = x), respectively.

1120

1121	Figure 4. Observed vs. predicted plots for European higher-forage methane emission (g d ⁻¹
1122	animal ⁻¹) prediction equations for the different categories, <i>viz.</i> , dry matter intake (DMI_C), dry
1123	matter intake and neutral detergent fiber (DMI+NDF_C), dry matter intake and starch
1124	(DMI+STA_C), dry matter intake and ether extract (DMI+EE_C), dietary (Diet_C), animal
1125	(Animal_C), animal without DMI (Animal_no_DMI_C), global (Global_C), GLOBAL
1126	NETWORK Tier 2, and IPCC Tier 2 (2006), and the extant Charmley et al. (2016) and Escobar-
1127	Bahamondes et al. (2017) equations. The gray and black solid lines represent the fitted
1128	regression line for the relationship between observed and predicted values, and the identity line
1129	(y = x), respectively.

1130

1131 Figure 5. Observed *vs*. predicted plots for North American higher-forage methane emission (g d⁻¹

animal⁻¹) prediction equations for the different categories, *viz.*, dry matter intake (DMI_C), dry

1133 matter intake and neutral detergent fiber (DMI+NDF_C), dry matter intake and starch

1134 (DMI+STA_C), dry matter intake and ether extract (DMI+EE_C), dietary (Diet_C), animal

1135 (Animal_C), animal without DMI (Animal_no_DMI_C), global (Global_C), GLOBAL

1136 NETWORK Tier 2 (2006), IPCC Tier 2 (2006), and the extant Charmley *et al.* (2016) equation.

- 1137 The gray and black solid lines represent the fitted regression line for the relationship between
- 1138 observed and predicted values, and the identity line (y = x), respectively.

63