

*Infrared radiative performance of urban trees: spatial distribution and interspecific comparison among ten species in the UK by in-situ spectroscopy*

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3 **Infrared radiative performance of urban trees: spatial distribution**  
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5 **spectroscopy**

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17 **Infrared radiative performance of urban trees: spatial distribution**  
18 **and interspecific comparison among ten species in the UK by in-situ**  
19 **spectroscopy**

20

21 **Abstract**

22 Understanding the ways in which tree species interact with solar radiation has  
23 previously focused on transmission and reflection of sunlight, typically by examining  
24 individual leaves. Here we used a tree crown spectroscopy measurement method to  
25 conduct in-situ tests on the radiative performance of ten commonly planted tree species  
26 in the UK. Tree crown transflectance (comprehensive effect of transmission and  
27 reflection) was examined to determine i), how radiative performance of individual trees  
28 varies spatially within a species, and ii), how infrared radiative performance differs  
29 between tree species. Our results show that tree crown transflectance depends on the  
30 combination of tree crown morphology, local foliage distribution (leaf density, gaps in  
31 crown foliage contour, concave or convex crown shapes), solar altitude and leaf size.  
32 Spatially, the strongest tree crown transflection was found primarily towards sky on the  
33 sunlit side of trees rather than towards the zenith, meaning that infrared transflection  
34 towards surrounding buildings and pedestrians is substantial. For all ten species, the  
35 tree crown transflectance in the frontal sunlit area was linearly correlated with solar  
36 altitude on sunny days. Hence, a solar altitude of 45° was chosen as the benchmark  
37 condition for comparing interspecific differences. Interspecific comparison indicated

38 that interspecific differences in the infrared radiative performance levels were strongly  
39 dependent on leaf size when no obvious gaps or concave shapes were present within  
40 the tree crowns. Our findings provide insights for understanding radiative interactions  
41 between urban trees and surrounding built environment, as well as for tree species  
42 selection in urban heat stress mitigation.

43

44 *Keywords:* Urban trees; Infrared radiative performance; Tree crown spectroscopy; Tree  
45 species; Urban microclimate

47 **List of symbols**

<b>Nomenclature</b>	
$A_{foliage}$	net foliage area of a measuring patch with foliage gaps, $m^2$
$A_{patch}$	total area of a measuring patch in viewing vision of the fiber-optic tip, $m^2$
$A_{void}$	void area of a measuring patch with foliage gaps, $m^2$
$D_{measured}$	sampling distance from the fiber-optic tip to a measuring patch on tree crown contours
$VA$	viewing angle of the spectrometer fiber-optic tip relative to horizontal plane, $^{\circ}$
$VR$	void ratio of a measuring patch with foliage gaps, –
$\bar{x}_i$	the $i$ -th average component value of the statistical mean spectrum, –
<b>Greek symbols</b>	
$\alpha$	solar altitude, $^{\circ}$
$\lambda$	wavelength, nm
$\tau R$	transflectance (comprehensive effect of transmission and reflection) of tree crown contour or canopy patch, –
$\tau R_{mean,800-900}$	mean transflectance in the near infrared wavelength range of 800–900 nm, –
$\tau R(\lambda)$	spectral transflectance at wavelength $\lambda$ nm, –
$\tau R_{foliage}$	net transflectance of foliage excluding foliage gaps in a measuring patch, –
$\tau R_{meas\_bgd}$	background noise signal in sampling transflectance spectrum, –
$\sigma_{A,x}$	type A standard error (uncertainty) of a variable $x$ , unit is the same as $x$
<b>Abbreviations</b>	
$IR$	infrared
$LAI$	leaf area index
$PCA$	principal component analysis
$SAz$	solar azimuth (direction)
$SVF$	sky view factor
$UHI$	urban heat island
$VIS$	visible

## 1 Introduction

Heat waves frequently hit many cities throughout the world and record temperatures are being experienced in many regions in recent years due to the increasing temperatures associated with greenhouse gas emissions [1][2]. It is documented that heat waves have remarkably negative influences on the health of urban inhabitants and contribute significantly to mortality of residents [1][3][4]. Furthermore, anthropogenic global warming is increasing the frequency, duration and intensity of heat waves [5], with the urban heat island (UHI) effect intensifying heat stress under extreme hot climates [6]. In the pursuit of creating benign urban environments to improve human health and well-being, it is imperative to seek effective solutions or strategies for mitigating heat waves and adapting to climate change.

Trees and green spaces enable the provision of better ecosystem services to urban environments [7]. For example, urban trees help to mitigate heat waves through radiative shading [8] and evapotranspiration [9], resulting in lowered air temperatures that help to regulate outdoor thermal comfort [10][11][12]. Gillner *et al.* [13] demonstrated that street trees played a key role in mitigating effects of heat and drought at highly sealed urban sites. Numerical simulations have shown that urban green coverage (trees and grasslands) helped to mitigate human heat stress under different climates [14][15], and the same point was argued in a systematic review of urban greening [16]. Spatial configuration of trees (individual trees versus different types of spacing and/or aggregations) may show significant, but inconsistent results (including

both positive and negative effects) of urban heat mitigation in cities with different climatic conditions [17][18]. Zhou *et al.* [17] presented that percent cover of trees was more important than their spatial configuration in predicting land surface temperature in Baltimore, while the opposite was found in Sacramento. Reasonable urban tree design approaches have a good performance in mitigating daytime and nighttime UHI effects in urban environments [19][20], and an appropriate combination of vegetation and urban geometry can help to mitigate the adverse effects of UHI and provide a better pedestrian thermal comfort [21][22]. More than that, the cooling effects of urban trees or green roofs through radiative shading and evapotranspiration also contribute to building energy savings [12][23] [24] [25]. It is clear that urban forests and trees make important contributions to cities by providing a multitude of benefits [7]. However, tree species have different physiological responses to heat waves and extreme heat events [26][27][28], depending on heat stress adaptability and water availability for the tree species [29], meaning that some trees cope better with high urban temperatures than others [26][30].

To understand urban tree cooling effects and physiological responses, researchers have been focusing on metrics of assessing outdoor thermal comfort (e.g. surface temperature, mean radiant temperature, physiological equivalent temperature), tree physiological indices (e.g. leaf or crown temperature, leaf area index, stomatal conductance, evapotranspiration rate), determinants and quantification of tree cooling capacity, tree shade effects and so on. Surface temperatures of trees and green spaces



are typically 10–20 °C lower than those of sealed ground or built surfaces exposed to sunlight in summer, leading to a significant reduction of mean radiant temperature [13][30][31]. Leuzinger *et al.* [30] reported tree crown temperatures of ten common tree species planted in Central European cities and declared that tree surface temperatures were circa -1 to +4 °C higher than the ambient temperature. They found trees in parks were significantly cooler than those surrounded by sealed ground and small-leaved trees remained cooler than large-leaved trees. Furthermore, surface temperatures of different tree species varied considerably, and the SVF (sky view factor – the ratio of the amount of the sky that can be seen from a given point on a surface to that potentially available, ranging from 0 to 1) value had a significant effect on tree surface temperatures [18][32]. It is presumed that the mean radiant temperature, which is linked to global temperature, air temperature and wind speed [19], is closely related to the urban thermal comfort [22]. Park *et al.* [33] proposed a multilayer mean radiant temperature model for pedestrians in a street canyon with trees. Physiological equivalent temperature is also an important assessing index of outdoor thermal comfort [18][34]. Zölch *et al.* [35] showed that planting trees had the strongest impact with an average physiological equivalent temperature reduction of 13% compared with existing vegetation. Zheng *et al.* [31] measured the influence of trees on the outdoor thermal environment in subtropical areas through field tests of relevant physiological indices and microclimatic parameters. They found that the widely planted fig tree *Ficus microcarpa* had the best cooling performance among four tree species studied, with the maximum reduction of physiological equivalent temperature due to the highest leaf area

index (LAI).

As to the determinants of tree cooling effects in terms of tree physiological indices, Morakinyo *et al.* [36] revealed that LAI was the main driver of tree cooling for outdoor temperature regulation, followed by trunk height, tree height and crown diameter. LAI was also highlighted by Armson *et al.* [37] and Rahman *et al.* [38]. Zhang *et al.* [39] stated that tall trees with a large LAI and canopy diameter should be a priority to improve the comfort of outdoor environments. Furthermore, different tree species may differ in microclimate benefits. Sanusi *et al.* [40] declared that the microclimatic benefits in streets with *Ulmus procera* and *Platanus x acerifolia* trees were significantly greater than the street with *Eucalyptus scoparia* trees, in terms of air temperature, relative humidity, solar radiation, mean radiant temperature, wind speed. In assessing a tree's cooling capacity via transpiration, it is presumed that different tree species have significant difference in evaporative cooling [24][41][42][43]. For example, as reported in [44], *Tilia cordata* trees with higher LAI and sap-wood area provided three times more transpiration than *Robinia pseudoacacia*. Konarska *et al.* [45] observed that night-time transpiration in all the seven species they studied amounted to 7 and 20 % of midday transpiration of sunlit and shaded leaves, respectively, in a high latitude city in Gothenburg, Sweden. There are some other ways of quantifying the cooling effects of urban trees through transpiration. For instance, Wang *et al.* [46] quantified the cooling capacity of urban trees as the surface cooling rate (the negative ratio of land surface temperature changes to fractional tree cover changes). They found that the surface

cooling rate was dominated by plant transpiration, up to 1.336 °C per percentage of fractional tree cover in heat waves in cities of the contiguous United States. Additionally, tree shade provides a good outdoor thermal comfort for pedestrians and enables energy savings. Rahman *et al.* [44][47] investigated vertical air temperature gradients under tree shades during summer days. Tree radiative shading effect was simulated by Upreti *et al.* [8] in a regional built environment, who predicted the capacity of urban trees in reducing urban surface and air temperature by about 2–9 °C and 1–5 °C, respectively. It is argued that shade trees have more prominent energy saving potential than urban lawns in a desert city [12]. Urban lawns tend to be hugely wasteful of water resources, so well chosen species of shade trees or xerophytes may be far more energy efficient than lawns in arid or semi-arid environments [48][49].

Apart from research into urban tree cooling effect, reasonable tree planting strategies were widely considered. Reasonable arrangement of spacing and size of street trees was proven to be beneficial for decreasing pedestrian mean radiant temperature [50]. Different tree planting strategies are available in [39][51][52][53][54][55][56] for improving outdoor thermal comfort in different scenarios and climates.

Generally, previous research into the cooling effect of trees in regulating urban microclimates simplified the physical characterization of radiative performance of trees [8][39][57][58][59]. Some studies have excluded the radiative shading effect of trees in urban microclimatic modelling [60][61]. Where foliage albedo values for different tree

species have been incorporated in microclimate modelling to illustrate their influences on thermal environment they are commonly fixed [59], yet the foliage albedo of trees varies temporally and spatially throughout a day (as presented in sections 4.2.3 and 4.3 of the present study). To better understand the cooling effect and capacity of urban trees, it is essential to characterise the radiative performance of different tree species from the perspective of physical characteristics, especially in the near infrared (NIR) region. In an earlier study, we established a novel methodology of characterising infrared (IR) radiative performance of urban trees using tree crown spectroscopy [62]. Experimental tests on *Tilia cordata* (aka small-leaved lime or little-leaf linden), a commonly planted tree species in the UK and Europe, were carried out to demonstrate the impact factors of IR radiative performance qualitatively in terms of transfectance at the tree crown level. However, the way in which IR radiative performance varies spatially across a tree crown and interspecific differences remain unanswered. Hence, the present study aims to explore spatial distribution of IR radiative performance across a tree crown, as well as to ascertain differences of IR radiative performance levels in tree species that are often planted in the UK. Since the radiative performance of tree species in terms of the tree crown transfectance varies with solar time throughout a day in a clear sky, as stated in our previous work [62], the way in which transfectance varies with solar time was explored as well, helping to make a benchmark for comparing interspecific performance differences. Ten commonly planted tree species in the UK were chosen for extensive in-situ tests of tree crown transfectance on sunny days using the established spectroscopy measurement method, in order to assess interspecific

differences of IR radiative performance levels on the same benchmark.

## 2 Test site, method and conditions

### 2.1 Test site and selection of tree species

Given the convenience and accessibility of trees for extensive in-situ tests, the test site was chosen at the Whiteknights campus, University of Reading (51.44° N, 0.94° W), UK. The campus covers an area of 123 hectares and features high plant diversity, with an estimated number of 150 different tree species. Ten commonly planted tree species in urban spaces throughout the UK were chosen for tests, as shown in Figure 1. Five of these are native British trees [63]: *Carpinus betulus* (hornbeam), *Acer campestre* (field maple), *Quercus robur* (English oak), *Tilia platyphyllos* (large-leaved lime), *Betula pendula* (silver birch).



**Figure 1.** Ten tree species selected in field tests, top row left to right: *Sequoiadendron*

*giganteum* (giant sequoia), *Carpinus betulus* ('Fastigiata' cultivar, hornbeam), *Acer campestre* (field maple), *Quercus robur* (English oak) and *Platanus x acerifolia* (London plane); bottom row left to right: *Tilia platyphyllos* (large-leaved lime), *Acer x freemanii* (autumn blaze maple), *Betula pendula* (silver birch), *Acer platanoides* ('Schwedleri' cultivar, copper Norway maple) and *Aesculus hippocastanum* (horse chestnut).

## 2.2 Test methods, instruments and facilities

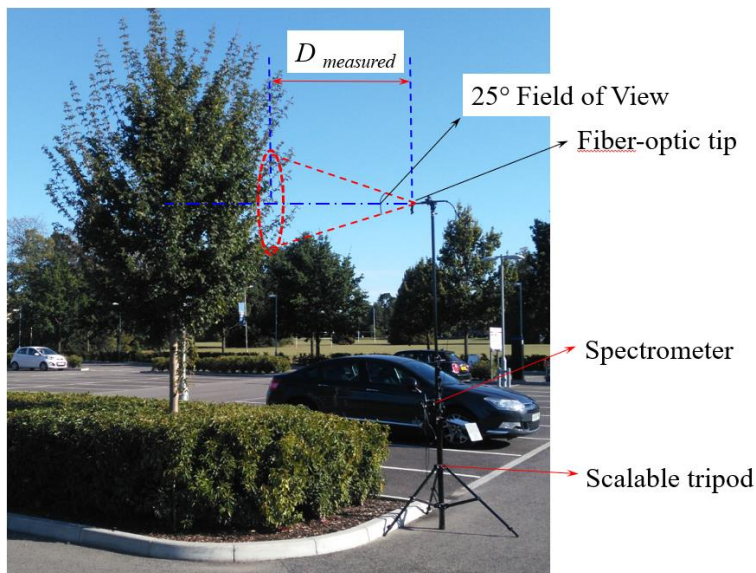
We previously established a methodology for characterising IR radiative performance of urban trees using tree crown spectroscopy [62]. The term tree crown **transflectance** ( $\tau R$ ) or **transflection**, which represents comprehensive radiative performance of trees at the crown level, is introduced in contrast to reflectance or transmittance at tree leaf levels. This is because when an optical sensor (i.e. spectrometer fiber-optic cable) is positioned at one side of trees to measure the radiative performance of 'a patch of tree crown surfaces' (abbreviated as 'measuring patch' hereafter), light received by the spectrometer usually comprises single-reflected, multi-reflected, multi-transmitted and transmitted-reflected rays through leaves.

A Black-Comet-SR model CXR-SR (StellarNET Inc., Tampa, Florida, USA) concave grating miniature spectrometer with a wavelength range of 350–1000 nm and a spectroradiometer mode was used for in-situ tests of tree crown transflectance. It covers the full visible (VIS) region (350–700 nm) and an important spectrum transition from

VIS to NIR around 700 nm where the leaf reflectance and crown transreflectance jump sharply (see testing results in section 4). Its spectral resolution is 0.5 nm. In the spectroradiometer mode, the fiber-optic tip can be fitted with a cosine receptor which allows measurement of solar spectra irradiance in a 180° field of view in the 350–1000 nm wavelength range. The spectroradiometer mode was only used for solar spectral irradiance measurements in conditions of different viewing angles of the fiber-optic tip.

Regarding the test facilities, a scalable tripod with a full height of 8 m was used to hold and position the fiber-optic cable of the miniature spectrometer in the field tests, as shown in Figure 2(a). The fiber-optic cable was mounted onto the top of the tripod at one end and connected to a data acquisition laptop at the other. A USB camera was fixed close to the fiber-optic tip in order to capture viewing vision of the latter in field tests. The miniature spectrometer was powered through a USB cable connected to the laptop. The battery when fully charged usually lasted for about five hours powering both the laptop and the spectrometer. The viewing angle of the optical sensor can be adjusted in all directions. The fiber-optic cable was usually used without any cosine receptor and had a field of view of 25°. A reflectance standard RS50 shown in Figure 2(b) was adopted to measure reference spectra. The reference spectrum was always sampled in a vertical plane in the transient solar azimuth (SAz) direction, as justified in our previous work [62]. Particularly, for a test sequence of transreflectance spatial distribution, a single vertical reference plane was employed, and the test sets were performed within several minutes. Appropriate measuring distance from the trees should be chosen to make the measuring results robust. It was found that the sampling

distance ( $D_{measured}$  in Figure 2(a)) from the fiber-optic tip to the measuring patches on the tree crown surfaces could not be too close (e.g. less than 1.5 m), as it would lead to higher transfectance levels than a sampling distance beyond 2.0 m where the measuring result was robust. This measurement issue was greatest for tree species with large-sized leaves such as *Platanus x acerifolia* (London plane), *Aesculus hippocastanum* (horse chestnut), in which the IR transfectance would be overestimated by 25% at a closer sampling distance (e.g. 1.0 m) due to a specular reflectance effect from partial leaf surfaces. In view of this, the sampling distance was kept between 2.0–5.0 m for most of the tested trees, depending on the tree crown diameters and the viewing vision of the USB camera. In some small trees, e.g. *Acer campestre* (field maple) a distance of 1.5–2.0 m was chosen because the height of the trees was less than 5 m and the diameter of the tree crowns was less than 3 m.



(a)





(b)

**Figure 2.** (a) A scalable tripod holding the optical fiber spectrometer for field tests; (b) Optic fiber tip of the spectrometer and a reflectance standard RS50.

A model SM2500 spectrometer (Spectral Evolution, Haverhill, Massachusetts, USA) with spectral resolution of 3.5–22 nm in the full range of UV (ultraviolet), VIS, NIR (wavelength range: 350–2500 nm and wavelength reproducibility of 0.1 nm at an accuracy of 0.5 bandwidth) was used in the laboratory to measure leaf reflectance of different tree species, in contrast to the tree crown transreflectance levels. The spectrometer was deployed together with a leaf clamp supplied by the manufacturer for leaf reflectance measurements. This spectrometer with a broad spectral range was bulky and expensive, thus not suitable for mounting on a tripod for field tests.

### 2.3 Test conditions

To explore spatial distribution rules of the radiative performance of individual trees, one fastigate hornbeam tree of 7.0 m height and one *Acer campestre* tree of 4.4 m height were primarily measured at multiple times, with different viewing angles of the spectrometer fiber-optic tip pointing at different spatial locations of their crowns. Wider field tests of the ten tree species were implemented to explore interspecific differences

of the radiative performance in terms of the transreflectance in the sunlit area of trees. At least five individual trees were sampled for each species in order to assess both intraspecific (within-species) and interspecific (between-species) variability in transreflectance. All field tests of tree crown transreflectance were carried out on sunny days or in sunny time slots without clouds shading the sun, in order to maintain a constant solar irradiance for the reference spectrum and the sampling transreflectance spectrum. Measuring heights of the transreflectance were usually chosen around the center height of the tree crown within  $\pm 1.0$  m deviation. Background noise signal was measured for several individual trees to assess its impact on measurements, which was found to be negligible as shown in section 4.2.1.

In addition to meteorological microclimate conditions (outdoor air temperature, incoming and outgoing shortwave radiation, incoming and outgoing longwave radiation), the soil moisture contents and physiological conditions (leaf temperature) of a single fastigiate hornbeam (*Carpinus betulus*) tree were monitored during the testing period of June to September 2019, providing information reference for the test conditions. Chlorophyll fluorescence and heat stress of five tree species were measured to inform their physiological stress status. Reflectance spectra of multiple individual leaves of various tree species were also measured in the lab for contrast.

### **3 Data processing and error analysis**

Repeated measurements of leaf reflectance or tree crown transreflectance spectra for

various tree species were implemented to obtain the statistical mean. Notate  $M$  as the number of spectra samples of a specific tree species. For each spectrum sample, assume that  $N$  rows of spectral values are recorded at different wavelength intervals. The statistical mean spectrum is calculated based on the sample component values in each row, as given in Equation (1).

$$\bar{x}_i = \sum_{j=1}^M x_{i,j} \quad (1)$$

where  $\bar{x}_i$  is the  $i$ -th average component value of the statistical mean spectrum, while  $x_{i,j}$  ( $i = 1, 2, 3, \dots, N$ ;  $j = 1, 2, 3, \dots, M$ ) denotes the  $i$ -th measured component value in the  $j$ -th sample spectrum.

With respect to error analysis of the statistical mean spectrum of the tree crown transfectance or the reflectance of individual leaves, A-type standard error (uncertainty) is usually used to estimate the statistical mean errors [64], as described in Equation (2).

$$\sigma_{A,x} = \sqrt{\frac{1}{M(M-1)} \sum_{j=1}^M (x_j - \bar{x})^2} \quad (2)$$

where  $\sigma_{A,x}$  is the A type standard error of a variable  $x$ ,  $M$  is the number of samples,  $\bar{x}$  is the statistical mean value of variable  $x$ , and  $x_j$  represents the  $j$ -th sample value.

The standard error of the statistical mean tree crown transfectance or leaf reflectance spectrum at each wavelength band is therefore given by:

$$\sigma_{A,\tau R}(i) = \sqrt{\frac{1}{M(M-1)} \sum_{j=1}^M (x_{i,j} - \bar{x}_i)^2} \quad (3)$$

where  $\sigma_{A,\tau R}(i)$  is the  $i$ -th component of the A-type standard error for the transreflectance ( $\tau R$ ) spectrum.

When calculating the statistical mean, if a spectrum in the samples was found to be outside of the 99.8% confidence interval, i.e.  $[-3\sigma_{A,TR}, +3\sigma_{A,TR}]$ , it was identified as a spectrum outlier. The statistical mean of the targeted spectrum was then recalculated excluding any outliers to minimize the contribution of measurement errors or non-target biological processes such as damaged or discolored leaves.

### 3.1 Statistical analyses using PCA (principal component analysis)

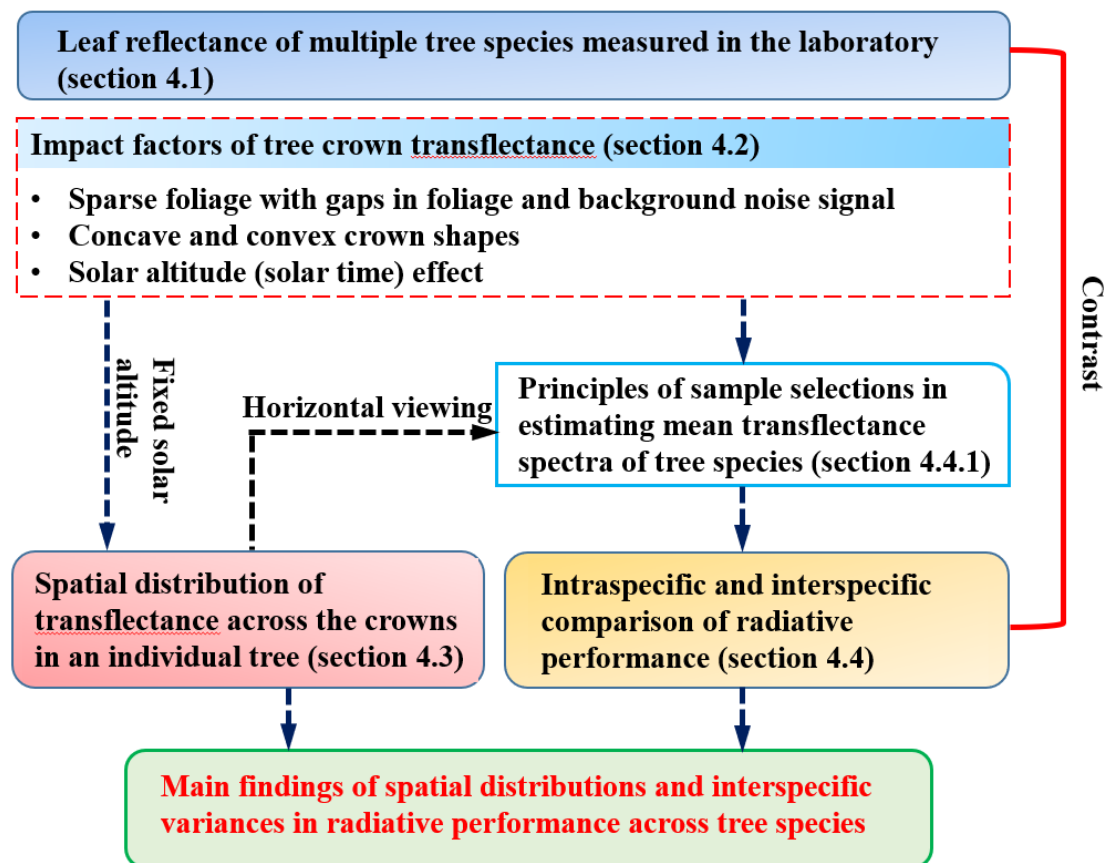
Unless otherwise noted all statistical analyses using PCA were carried out in R version 3.6.0 [65]. Packages “vegan” [66], “factoextra” [67] and “FactoMineR” [68] were used to analyse spectral data. Due to the nature of this data, ordination using Principal Component Analysis (PCA) and Principal Coordinates Analysis (PCoA or MDS - Multidimensional Scaling) generate the same outcome; for consistency these ordinations are referred to as PCA. Briefly, PCA takes a multivariate dataset of potentially correlated variables and transforms them into fewer, uncorrelated variables (principal components). This approach is commonly applied in ecology (see Legendre and Legendre [69]) and has previously been applied to leaf spectral data in remote sensing studies (e.g. Cavender-Bares et al. [70]). Spectral data (transflectance measured across the range 350-1000 nm) were interpolated for each of 67 individual trees from

0.5 nm bands into 5 nm bands and then scaled prior to analysis. The 10-fold reduction in data points following interpolation (1300 to 130 bands per tree) led to a loss of only 0.1% of overall explained variance. Significant correlations between PCA axes and spectra were used to assess the specific differences represented by the axes, and species was used as a grouping variable.

Further examination of inter- and intra-specific differences in transreflectance profiles for 350-1000 nm was conducted by converting PCA eigenvalues into pairwise euclidean distances and then applying permutational multivariate analysis of variance (PERMANOVA) on the resulting distance matrices. The multivariate homogeneity of dispersion (variance) was then assessed to examine the extent to which observed differences i) could be attributed to interspecific differences in transreflectance values, or ii) may have been confounded by differences in intraspecific variance.

#### **4 Results and discussion**

Figure 3 illustrate the flow chart of research framework in the discussion, in order to determine i), how radiative performance of individual trees varies spatially within a species, and ii), how infrared radiative performance differs between tree species.

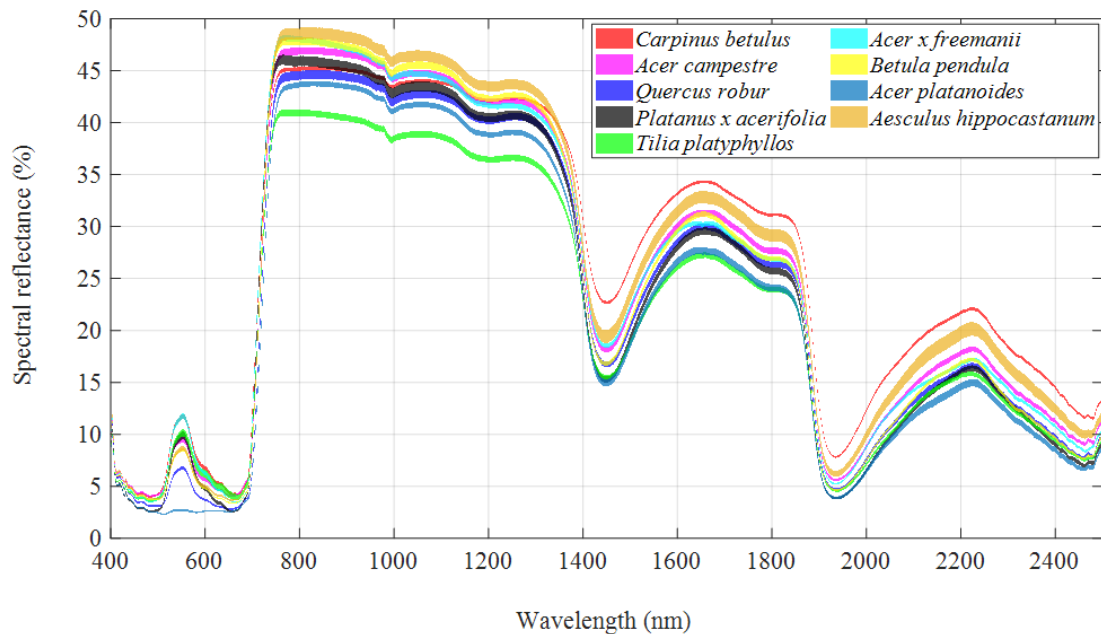


**Figure 3.** Flow chart of the research framework in discussion.

#### 4.1 Leaf reflectance measured in the laboratory

Leaf reflectance spectra of 9 targeted tree species (except *Sequoiadendron giganteum*) were measured at multiple times in the laboratory using the SM2500 spectrometer (Spectral Evolution) with a leaf clamp, in order to estimate statistical mean leaf reflectance spectra. Leaf reflectance of the species *Sequoiadendron giganteum* was not measured, as the spectrometer could not be used to measure individual needle leaves. For each tree species, 10 leaves were collected randomly from individual trees and 5–10 trees were sampled. Reflectance spectra of the collected leaves were measured within 1 hour after leaf collections to guarantee that the leaves would not lose moisture.

Each selected leaf was measured once, and the measuring point was chosen randomly with consideration of avoiding main leaf veins. Figure 4 gives the statistical mean leaf reflectance spectra of 9 tree species, indicating there is a minor difference of mean leaf reflectance spectra among various tree species (within  $\pm 4\%$  of spectral reflectance values). The standard mean errors of spectral reflectance of various tree species in the wavelength range of 400–2500 nm were within  $\pm 0.5\%$  with a large number of samples (50–100 samples for each species). Data on statistical mean reflectance spectra of leaves across species can be accessed in *Appendix A*.



**Figure 4.** Standard mean error bands of leaf reflectance spectra of nine tree species measured in the laboratory and sampled from 5–10 trees for each species.

## 4.2 Impact factors of tree crown transreflectance

Through a multitude of in-situ tests, it was found that visibly non-uniform foliage

distribution in the measuring patches of tree crown contours, such as sparse foliage with gaps in crown foliage and/or concave crown shapes, would affect tree crown transfectance ( $\tau R$ ) to different extents. On the other aspect, our previous work indicated that solar time was one of the most important factors impacting  $\tau R$  levels [62]. It is therefore necessary to elucidate the impacts and determine the principles of sample selections for estimating the radiative performance levels of multiple tree species.

#### **4.2.1 Impact of sparse foliage with gaps in crown foliage and background noise signal**

Sparse foliage is commonly visible at some locations of tree crowns. To estimate measurement errors of the measuring patches with gaps in the sparse foliage and examine the impact of background noise signal penetrating the gaps, it is essential to measure the transfectance of the measuring patch along with the background noise signal. Void ratio of the gaps in the measuring patch should be also determined, in order to correct the net transfectance of the sparse foliage by excluding background noise signal. Total area of a measuring patch with gaps ( $A_{patch}$ ) is sum of the net foliage area ( $A_{foliage}$ ) and the void area ( $A_{void}$ ) in equation (5). Void ratio of the gaps ( $VR$ ) is calculated in equation (6). Relation between the measured transfectance ( $\tau R_{meas\_patch}$ ), the net transfectance of foliage ( $\tau R_{foliage}$ ) and the background noise signal ( $\tau R_{meas\_bgd}$ ) is described by equation (7) based on radiative energy conservation. Thus, the net transfectance of foliage ( $\tau R_{foliage}$ ) is rearranged in equation (8) and the relative error is estimated by equation (9).



$$A_{patch} = A_{foliage} + A_{void} \quad (5)$$

$$VR = A_{void}/A_{patch} = A_{void}/(A_{foliage} + A_{void}) \quad (6)$$

$$\begin{aligned} \tau R_{meas\_patch} &= \frac{\tau R_{foliage} \cdot IRR_{ref} \cdot A_{foliage} + \tau R_{meas\_bgd} \cdot IRR_{ref} \cdot A_{void}}{A_{patch} \cdot IRR_{ref}} \\ &= \tau R_{foliage} \cdot VR + \tau R_{meas\_bgd} \cdot (1 - VR) \end{aligned} \quad (7)$$

$$\tau R_{foliage} = \tau R_{meas\_patch} + \frac{VR}{1 - VR} \cdot (\tau R_{meas\_patch} - \tau R_{meas\_bgd}) \quad (8)$$

$$Error_{sparsefoliage} = \frac{VR}{1 - VR} \cdot \left(1 - \frac{\tau R_{meas\_bgd}}{\tau R_{meas\_patch}}\right) \times 100\% \quad (9)$$

417

418 Take the  $\tau R$  measurement of a *Tilia platyphyllos* tree with sparse foliage as an example  
 419 to illustrate the impact of background noise. Figure 5(a) shows the viewing vision of  
 420 the tree crown  $\tau R$  measurement in the sunlit area and SAz (solar azimuth) direction  
 421 with solar altitude  $\alpha = 34^\circ$ . The void ratio of gaps (VR) in the vision was estimated by  
 422 the ImageJ software [71], resulting in  $VR = 23.3\%$  (see Figure 5(b)). The net  $\tau R$  of  
 423 the foliage ( $\tau R_{foliage}$ ) was calculated in equation (8). Figure 5(c) shows the corrected  
 424 net foliage  $\tau R$  spectrum excluding gaps in foliage compared to the measured  $\tau R$   
 425 spectra. It suggested that the measurement error by the background noise was very  
 426 small (within 3% deviation). Furthermore, the net  $\tau R$  of the foliage was lower (-4%  
 427 deviation of IR  $\tau R$  in this case) than the  $\tau R$  of a dense measuring patch in another  
 428 *Tilia platyphyllos* tree, indicating that a sparse foliage (sparse leaf density) degraded the  
 429  $\tau R$  levels.

430

431 Distributions of sparse foliage on individual trees are amorphous and heterogeneous,  
 432 resulting in different values of void ratio. When choosing measuring patches with

visibly dense foliage (no obvious gaps,  $VR < 10\%$ ), it is easy to control the measuring errors. For most of the trees in tests,  $VR \approx 5\%$ , while the observed spectral transfectance between the measuring patches and the background noise signal is usually less than 30% ( $\tau R_{meas\_patch} - \tau R_{meas\_bgd} \leq 30\%$ ), it implies that the absolute measurement error of the transfectance is below 3% using equation (9). In this sense, background noise in the transfectance measurement with visibly dense foliage can be disregarded.



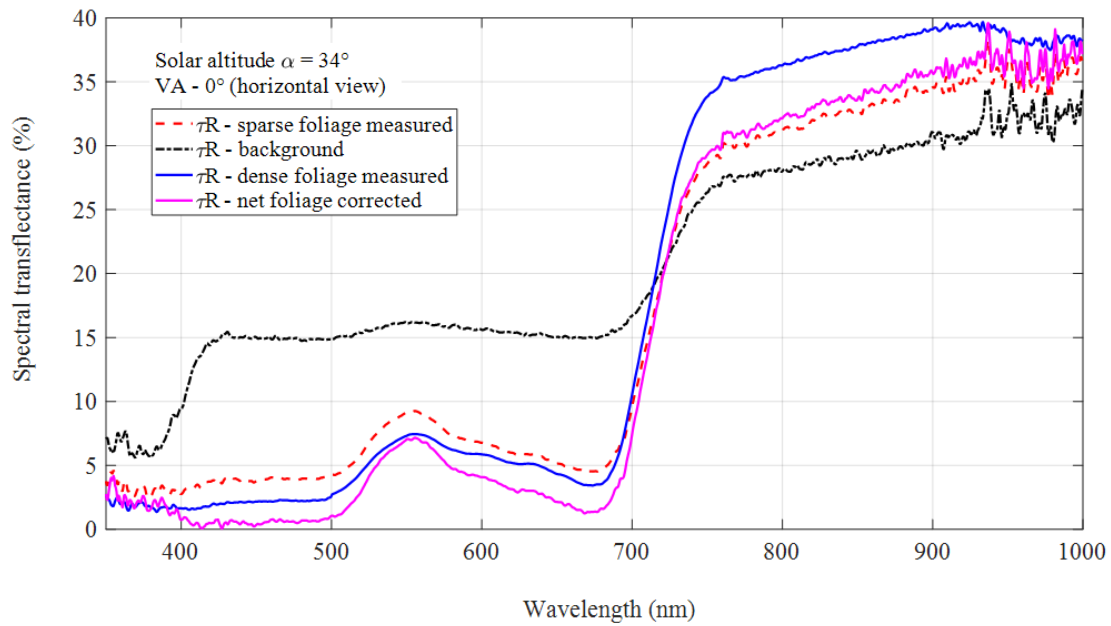
(a) Viewing vision (circle in red dashed line) of the tree crown transfectance measurement in the SAz direction



**Original image**

**Adjusted by ImageJ software**  
**Void ratio: VR = 23.3%**

(b) Void ratio of gaps in the viewing vision determined by the ImageJ software



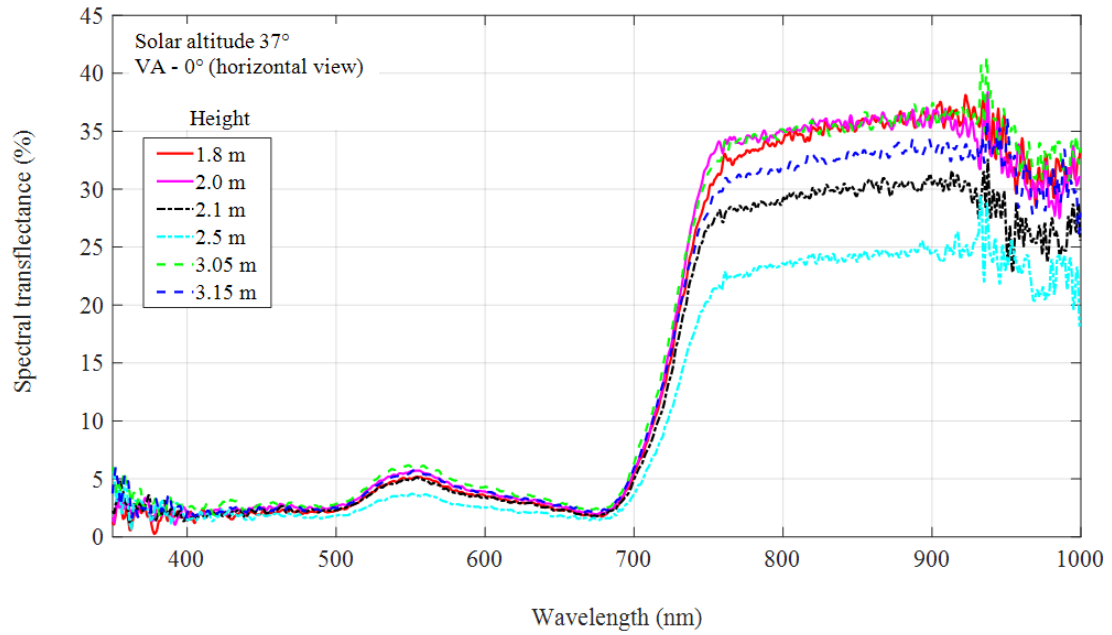
(c) Net transfectance of foliage excluding gaps compared to measured transfectance

**Figure 5.** Tree crown transfectance of a measuring patch with sparse foliage and gaps in a *Tilia platyphyllos* and estimation of the background noise impact with  $\alpha = 34^\circ$ .

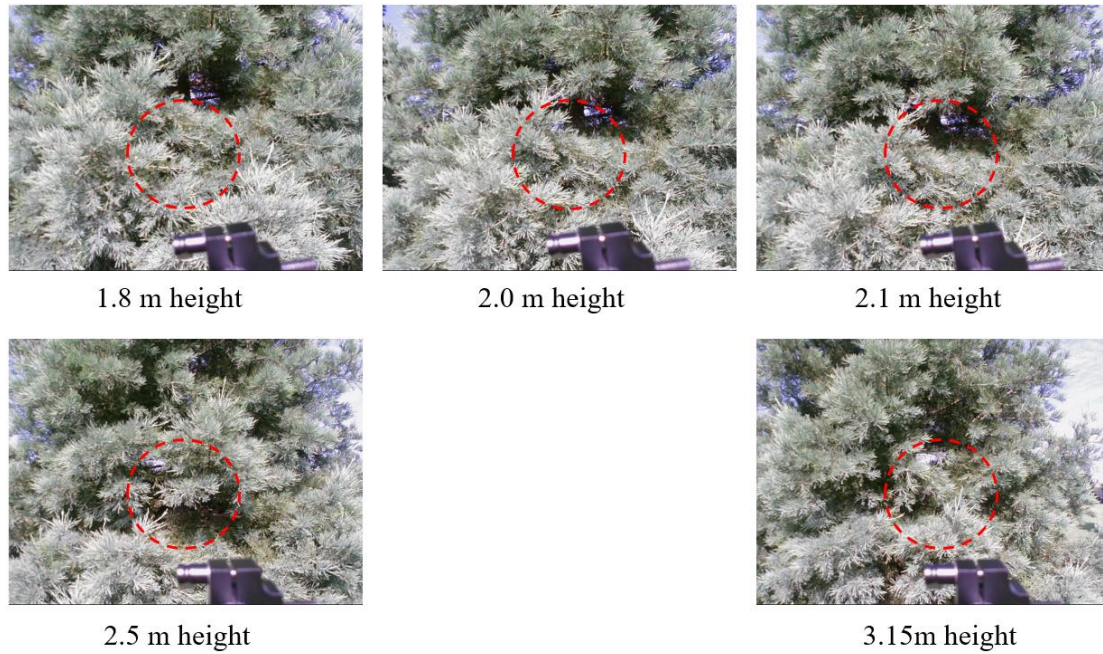
#### 4.2.2 Impact of concave and convex shapes in crown foliage

It was found that concave crown contours decreased the  $\tau R$  levels, while the convex contours maximised the  $\tau R$  levels in various tree species. Take the in-situ test of a 5.5 m *Sequoiadendron giganteum* tree as an example. Figure 6(a) shows the  $\tau R$  measurement of the tree at different heights in the SAz direction ( $\alpha = 37^\circ$ ) with or without concave shapes in the measuring patches. Viewing vision of the fiber-optic tip in tests (red circles in dashed lines represent the vision) was shown in Figure 6(b). Maximum  $\tau R$  appeared at 1.8 m height nearly without concave shapes in the measuring patch, while the minimum  $\tau R$  appeared at 2.5 m height with concave shapes accounting for approximately half of the area of the viewing vision, and the  $\tau R$  at other heights with concave shapes of different extents fell in between. Foliage gaps

were rather small compared to the area of concave shapes in these cases. As sunlight was captured by concave shapes in the measuring patches, the concave shapes degraded  $\tau R$  levels to different extents depending on the specific scenarios. When the center of the viewing vision (the circles in Figure 6(b)) deviated from the concave shapes and the ratio of concave shapes was not big, it only slightly degraded the  $\tau R$  (see  $\tau R$  spectra at heights of 2.0 m and 2.1 m in Figure 6(a)). It implies that received light of the fiber-optic tip is not evenly contributed by the viewing vision.



(a)



(b)

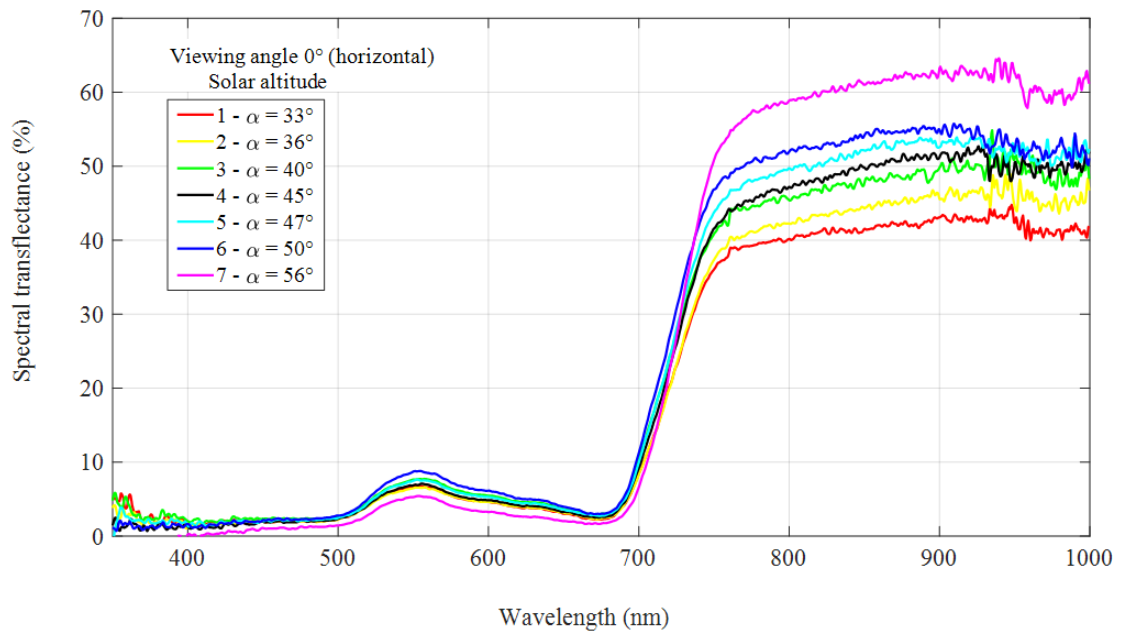
**Figure 6.** Tree crown  $\tau R$  measurement of a *Sequoiadendron giganteum* at different heights with or without concave shapes in the measuring patches (a)  $\tau R$  spectra ( $\alpha = 37^\circ$ ); (b) viewing vision of the fiber-optic tip in tests (see red circles in dashed lines).

The impact of sparse foliage with gaps and concave crown shapes on the transfectance is complicated and heterogeneous, and different trees within a species do not share a common feature of local foliage distributions. Hence, it is preferable to avoid sampling measuring patches that exhibit sparse foliage with gaps and concave contours when collecting data for statistical analysis of radiative performance levels of various tree species. Any such patches are likely to be identified as outliers in transfectance sampling. Generally, it is possible to find measuring patches on trees with relatively dense foliage and without visible concave contours. In our field tests of 10 tree species, even for the tree species with the sparsest foliage (*Betula pendula* - silver birch) it was

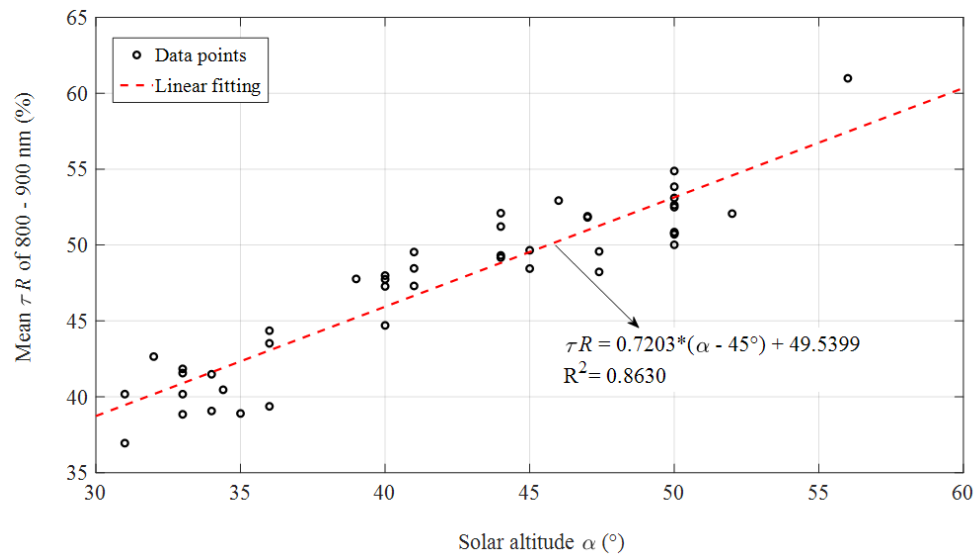
easy to find dense foliage for tests without gaps in foliage and concave contours using the naked-eye.

#### 4.2.3 Impact of solar altitude on tree crown $\tau R$

To explore the relationship between change in transreflectance and solar time, solar altitude ( $\alpha$ ) was recorded when sampling the transreflectance, referring to a website [72]. Figure 7 shows the  $\tau R$  spectra in the frontal sunlit area of a *Carpinus betulus* tree at different solar altitudes. It suggests that the tree crown transreflectance increases as solar altitude rises. In further data analysis, it is found that the  $\tau R$  in the IR region tends to be linearly correlated with  $\alpha$ , while no obvious variation is observed in the VIS region due to a low  $\tau R$  level. To dig out laws of change of  $\tau R$  versus  $\alpha$ , mean transreflectance in the wavelength range of 800–900 nm ( $\tau R_{mean,800-900}$ ) was taken as an indicator, because tree crown  $\tau R$  spectra usually tend to be flat and hold the maximum spectral transreflectance in the NIR wavelength range of 800–900 nm. Figure 8 shows the  $\tau R_{mean,800-900}$  of the *Carpinus betulus* trees in terms of the tree crown  $\tau R$  in the frontal sunlit area linearly correlated with solar altitude  $\alpha$ . The rule of change of IR transreflectance versus solar altitude allows making a benchmark for intraspecies and interspecies comparisons in section 4.4, by converting  $\tau R$  spectra at different solar altitudes into equivalent  $\tau R$  spectra at the same solar altitude. Figure 9 shows the linear fitting results of the  $\tau R$  spectra in the frontal sunlit area of the *Carpinus betulus* trees varying with solar altitude in the wavelength range of 350–1000 nm.

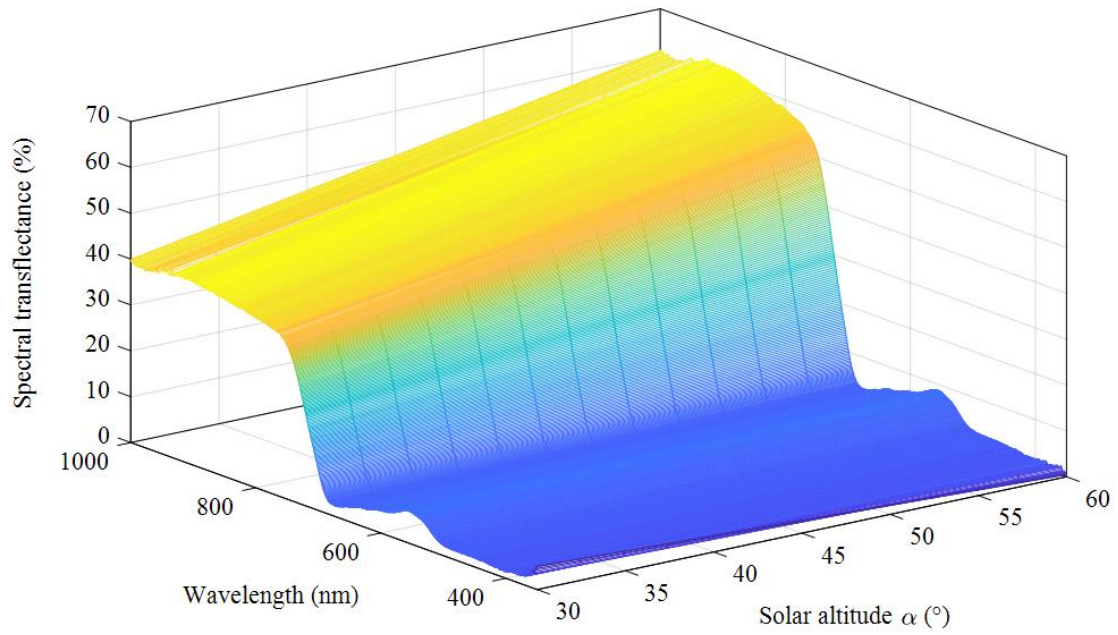


**Figure 7.** Transflectance spectra of tree crown surfaces for a *Carpinus betulus* tree in the frontal sunlit area at different solar altitudes ( $\alpha$ ).

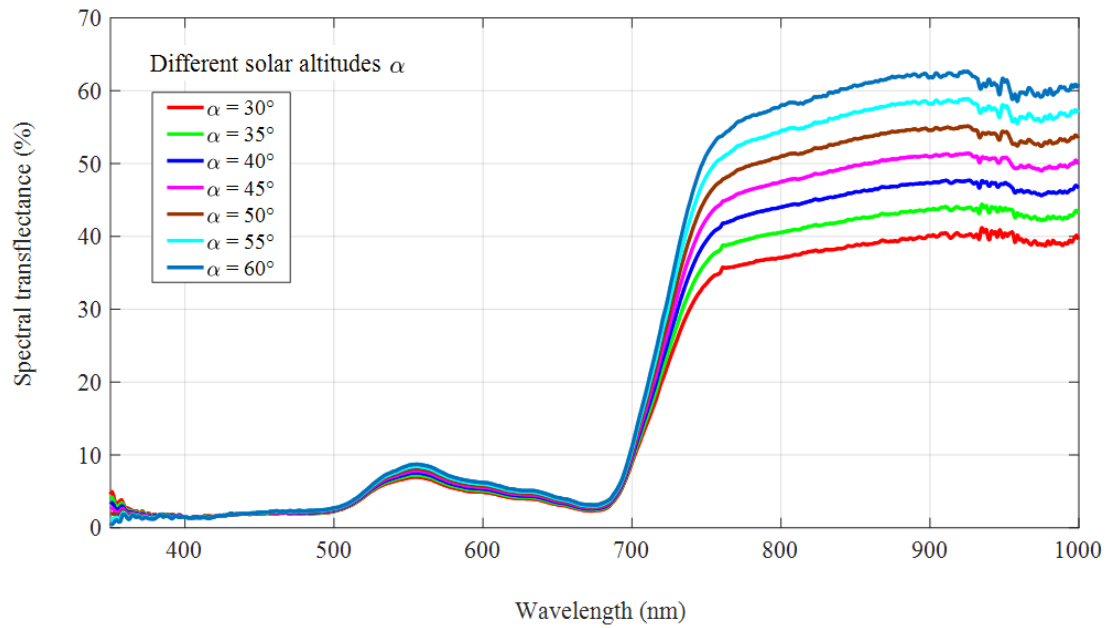


**Figure 8.** Linear fitting of the 800–900 nm mean transfectance ( $\tau R_{mean,800-900}$ ) with the solar altitude  $\alpha$  for individual *Carpinus betulus* trees using samples in the frontal sunlit area of trees.





(a) 3D plot



(b) 2D plot with discrete values of solar altitude

**Figure 9.** Linear fitting of tree crown transfectance spectra in the frontal sunlit area of *Carpinus betulus* trees varying with solar altitude in the wavelength range of 350–1000 nm.



In all the ten species measured, it was found that the tree crown IR transfectance in the frontal sunlit area of trees for a specific species was linearly correlated with solar altitude on sunny days. The coefficients of determination ( $R^2$ ) in most cases of linear fittings were above 0.76, except the tree species *Betula pendula* and *Aesculus hippocastanum*. It was presumed that the low  $R^2$  in the fitting of *Betula pendula* was mainly due to apparently sparse leaf density of the species. As to the species *Aesculus hippocastanum*, the low  $R^2$  in the fitting was attributed to the development of horse chestnut leaf-miner throughout the summer [73]. Horse chestnut leaf-miner caused brown blotch mines to develop between the leaf veins, resulting in a degradation of the transfectance spectrum especially in 750–900 nm wavelength range.

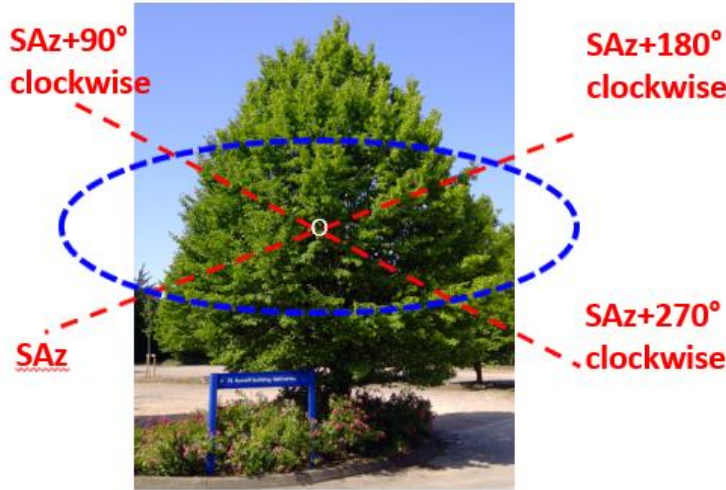
### 4.3 Spatial distribution of transfectance around tree crowns

As argued in our earlier work [62], transmission and reflection spectra at the leaf level were similar between trees (similar results also shown in section 4.1), while substantial variations were found in tree crown  $\tau R$  (transfectance) spectra due to crown structural difference and solar time. To explore spatial distribution rules of the radiative performance across a tree crown,  $\tau R$  spectra in typically horizontal and vertical loops around the tree crowns have been examined.

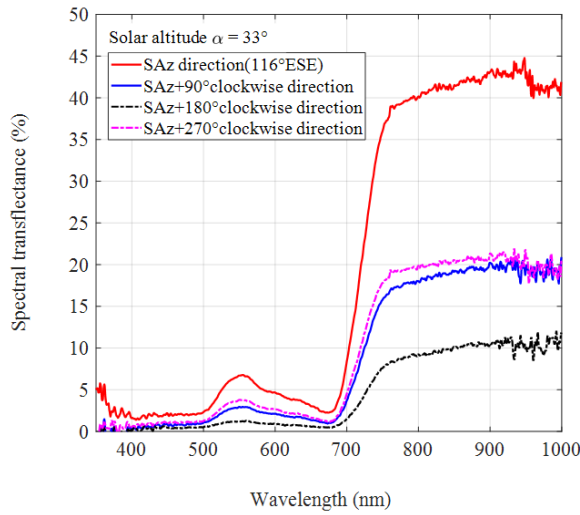
#### 4.3.1 $\tau R$ Distribution in a horizontal loop

Four orientations in a horizontal plane were concerned, as shown in Figure 10(a) with the 7.0 m height *Carpinus betulus* tree. A single vertical reference plane in the solar

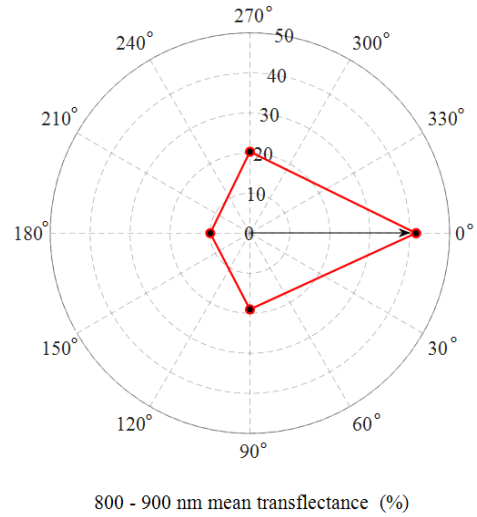
azimuth (SAz) direction was chosen for  $\tau R$  spectra measurements in the four horizontal directions (SAz; SAz +90° clockwise; SAz +180° clockwise; SAz +270° clockwise). The test sequence was implemented at 4.4 m height during 9:44–9:47 am (British Summer Time) on the sunny day of 20th August 2019 with solar altitude  $\alpha = 33^\circ$ . Figure 10(b) shows measured  $\tau R$  spectra in the four horizontal directions. To display the transfectance distribution in different directions intuitively, the  $\tau R_{mean,800-900}$  (800–900 nm mean transfectance) was taken as an indicator again. Figure 10(c) plots the distribution of  $\tau R_{mean,800-900}$  in the horizontal loop in polar coordinates, where the radius from the origin ‘0’ to the point represents the value of  $\tau R_{mean,800-900}$ . The maximum  $\tau R$  spectrum in the horizontal loop appears in the frontal sunlit area of the tree, followed by SAz+90° and SAz+270° clockwise directions. The tree crown  $\tau R$  on the shade side (SAz +180° clockwise) has the lowest  $\tau R$  level. It seems that the  $\tau R_{mean,800-900}$  in the SAz+90° and SAz+270° directions tends to be spatially symmetric. This was because no obvious concave contours were viewed on the measuring patches in both directions. We have observed scenarios of non-symmetric distribution in the two directions in an *Acer campestre* tree during 10:35–10:40 am on 27th June 2019, as shown in Figure 11. The  $\tau R_{mean,800-900}$  in the SAz+270° clockwise direction was much lower than that in the SAz+90° clockwise direction, as apparent concave contours appeared in the SAz+270° clockwise direction, resulting in a reduction of the  $\tau R$  level.



(a)

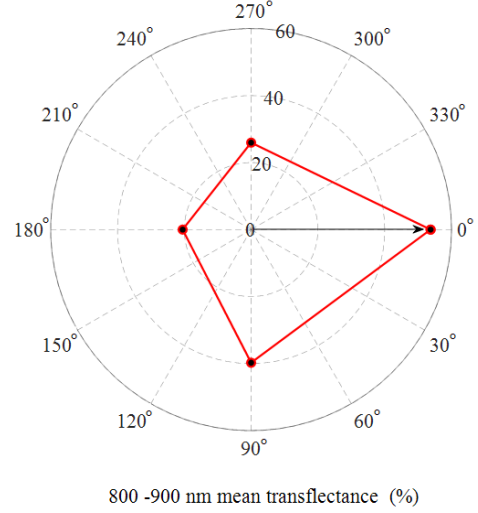
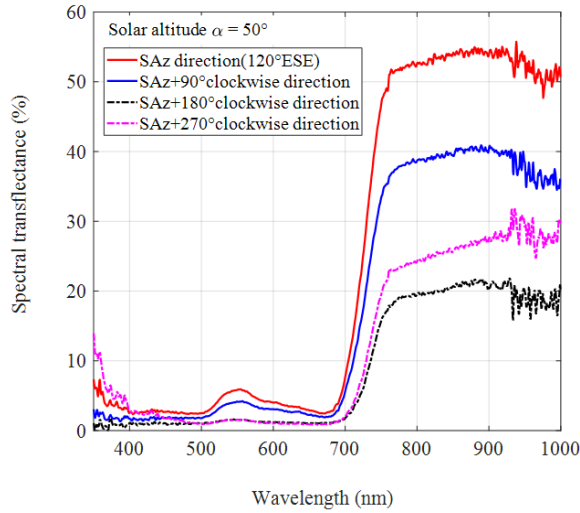


(b)



(c)

**Figure 10.** Transflectance ( $\tau R$ ) spectra distribution of a *Carpinus betulus* tree in a horizontal loop around the tree crown at  $\alpha = 33^\circ$  (a) sketch of four horizontal directions; (b) measured  $\tau R$  spectra in four directions (SAz – Solar azimuth; ESE – East South East); (c) distribution of  $\tau R_{mean,800-900}$  in polar coordinates ( $0^\circ$  – SAz direction;  $90^\circ$  – SAz +  $90^\circ$  clockwise;  $180^\circ$  – SAz +  $180^\circ$  clockwise;  $270^\circ$  – SAz +  $270^\circ$  clockwise).



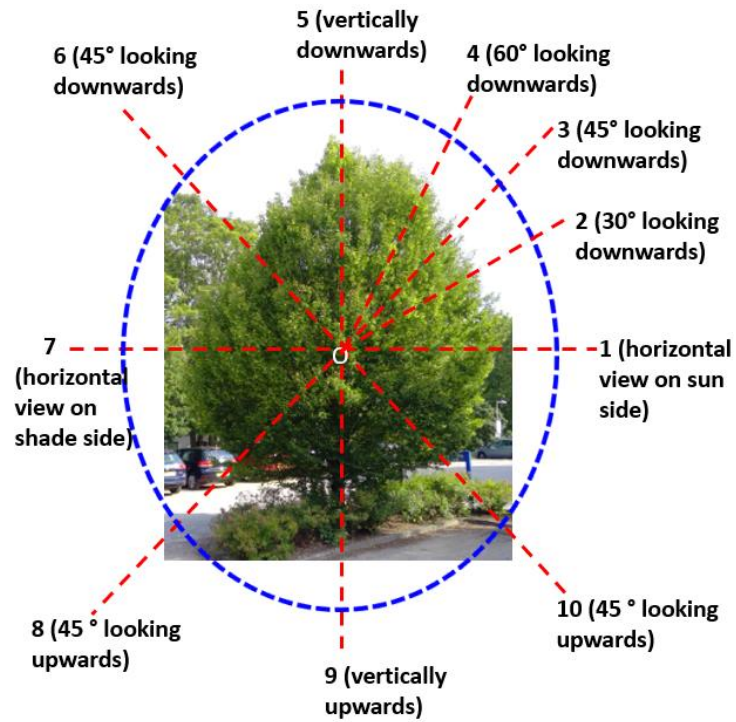
**Figure 11.** Transflectance ( $\tau R$ ) distribution of an *Acer campestre* tree in a horizontal loop around the tree crown at  $\alpha = 50^\circ$  (a) measured  $\tau R$  spectra in four directions; (b) distribution of  $\tau R_{mean,800-900}$  in polar coordinates.

#### 4.3.2 $\tau R$ Distribution in a vertical loop around the tree crown and aligned with the solar azimuth direction

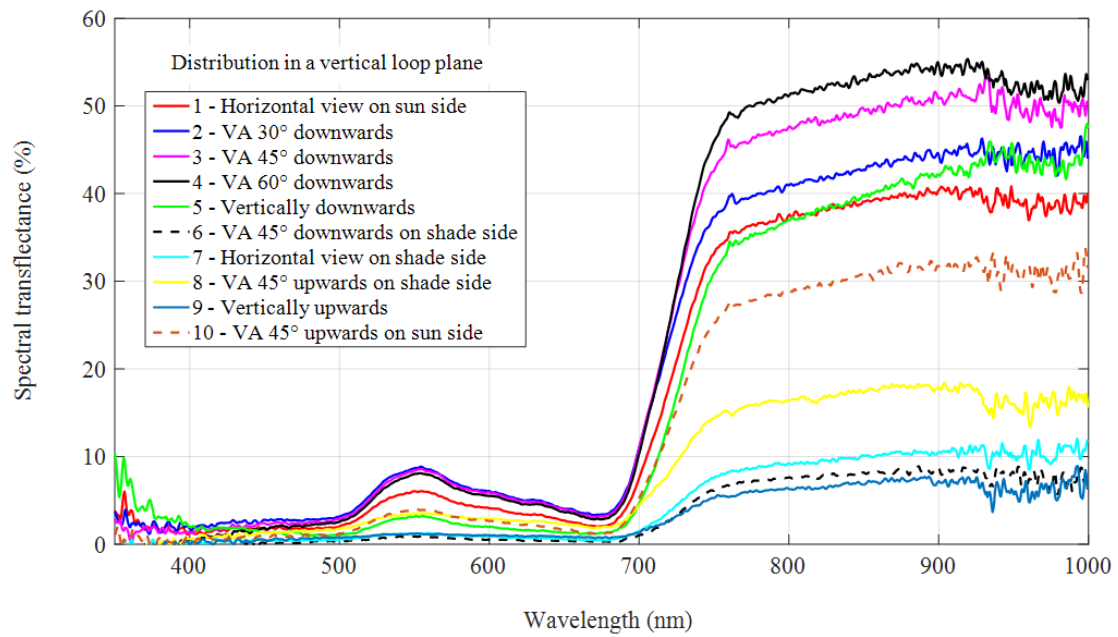
Two typical vertical loops around the tree crown were chosen in determining distribution of tree crown  $\tau R$  (transflectance) in vertical planes. One was a vertical loop aligned with the SAz (solar azimuth) direction. The other was a vertical loop perpendicular to the SAz direction (see next section 4.3.3). Figure 12(a) sketches 10 sampled directions distributed in the vertical loop aligned with the SAz in the 7.0 m height *Carpinus betulus* tree. Figure 12(b) gives measuring results of the  $\tau R$  spectra in the 10 sampled directions with the tree at  $\alpha = 35^\circ$ . The distribution of  $\tau R_{mean,800-900}$  is delineated in Figure 12(c). It indicates that points ‘1’, ‘2’, ‘3’, ‘4’, ‘5’ in the frontal sunlit area of the tree have high  $\tau R$  levels, while other points hold

relatively low  $\tau R$  levels. Especially, point ‘9’ beneath the tree has the minimum  $\tau R_{mean,800-900}$  value (only 6.9%). Additionally, the point ‘4’ representing the case of fiber-optic tip tilted  $60^\circ$  looking downwards has the maximum  $\tau R_{mean,800-900}$ . The values of  $\tau R_{mean,800-900}$  in the frontal sunlit area at different viewing angles (VA =  $30^\circ$ ,  $45^\circ$ ,  $60^\circ$  and  $90^\circ$  looking downwards) varied with solar time, as it was found that the maximum value did not maintain in the direction of VA =  $60^\circ$  downwards. Figure 13 gives the vertical loop distribution of the *Carpinus betulus* tree at the solar altitude  $\alpha = 49^\circ$ . Comparing the horizontal samples of point ‘1’ in Figures 12(c) and 13, it confirmed that the tree crown  $\tau R$  at a higher solar altitude ( $\alpha = 49^\circ$ ) is higher than that at a lower case ( $\alpha = 35^\circ$ ). Furthermore, the strongest tree crown transfection was found primarily towards sky on the sunlit side of trees rather than towards zenith. Infrared transfection towards surrounding buildings and pedestrians is substantial. The finding provides insights on understanding radiative interactions between urban trees and surrounding built environments.

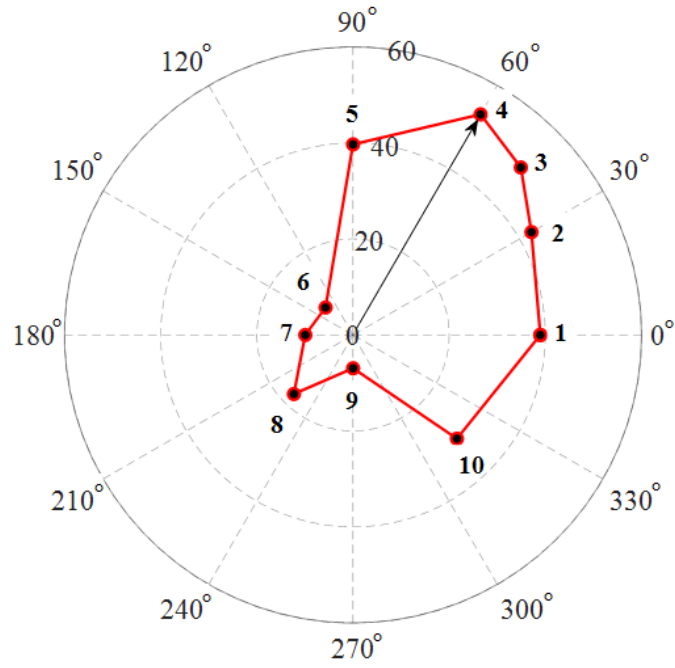
Additionally, it is noteworthy that tree crown morphology is linked to the space of the frontal sunlit area of trees, implying that tree morphology affects distributions of the tree crown  $\tau R$ .



(a)



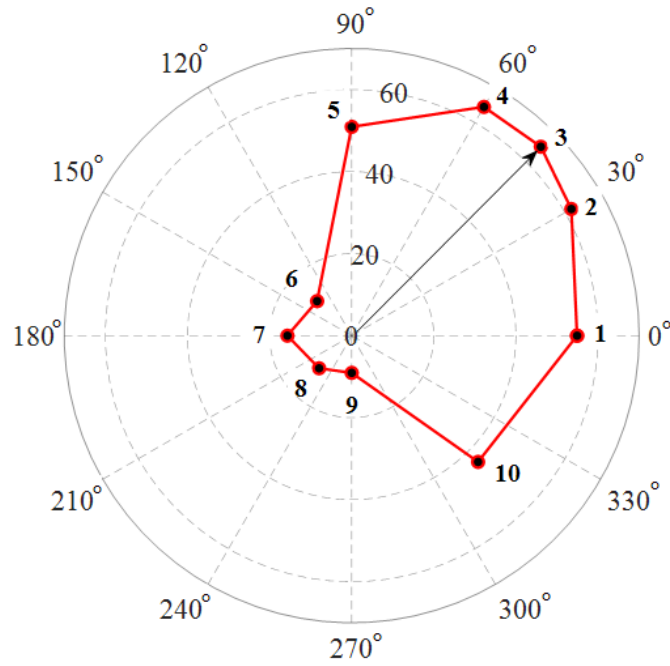
(b)



800 - 900 nm mean transfectance (%)

(c)

**Figure 12.** Transflectance ( $\tau R$ ) spectra distribution of a *Carpinus betulus* tree at  $\alpha = 35^\circ$  in a vertical loop around tree crown and aligned with the solar azimuth (SAz) direction (a) sketch of 10 sampled directions in the vertical loop; (b) measured  $\tau R$  spectra; (c) distribution of  $\tau R_{mean,800-900}$  in polar coordinates.



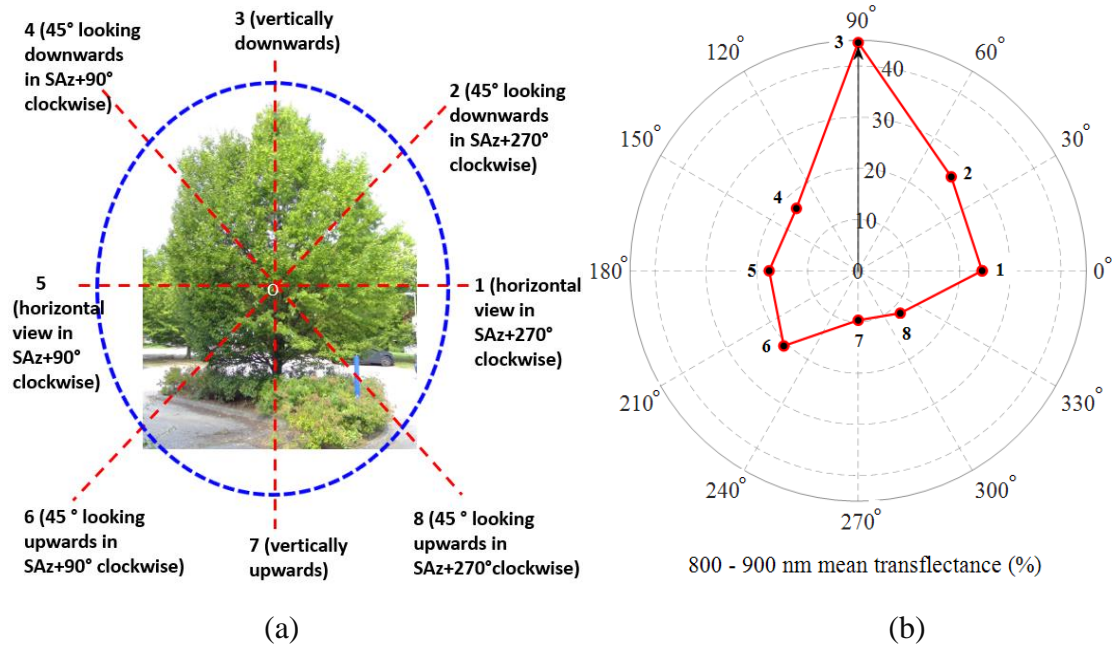
800 - 900 nm mean transfectance (%)

**Figure 13.**  $\tau R_{mean,800-900}$  distribution of a *Carpinus betulus* tree in polar coordinates at  $\alpha = 49^\circ$  in a vertical loop around the tree crown and aligned with the SAz direction

### 4.3.3 $\tau R$ Distribution in a vertical loop around the tree crown and perpendicular to the solar azimuth direction

The vertical loop perpendicular to the SAz direction was sampled in 8 directions, as sketched in Figure 14(a). Figure 14(b) gives the measured  $\tau R$  spectra of the *Carpinus betulus* tree in the 8 sampled directions at  $\alpha = 37^\circ$ , while Figure 14(b) plots distribution of  $\tau R_{mean,800-900}$  (800–900 nm mean) in polar coordinates. The maximum  $\tau R$  appears at point ‘3’ (vertically downwards), with  $\tau R$  at the other points (‘1’, ‘2’, ‘4–8’) on lower levels. Comparing Figure 14(b) with the distribution in Figure 12(c) neglecting a slight change of solar altitude, it confirms that the primary tree crown transfection is towards sky on the sunlit side of trees rather than towards zenith.





**Figure 14.** Transflectance ( $\tau R$ ) spectra distribution of a *Carpinus betulus* tree at  $\alpha = 37^\circ$  in a vertical loop around tree crown and perpendicular to the SAz direction (a) sketch of 8 sampled directions in the vertical loop;; (b) distribution of  $\tau R_{mean,800-900}$  in polar coordinates.

#### 4.4 Interspecific difference comparison of radiative performance levels

##### 4.4.1 Principles of sample selections in estimating statistical mean $\tau R$ spectra of tree species

As elucidated in sections 4.2 and 4.3, the impact factors of the tree crown  $\tau R$  are diverse, and the spatial distribution of the  $\tau R$  in tree crowns is nonuniform and varies with solar altitude. A benchmark is therefore needed to make interspecific comparison. Principles of test sample selections in estimating the statistical mean  $\tau R$  spectra of various tree species are stated below:

- Visually dense measuring patches without visible gaps in crown foliage and

659 concave crown contours in the viewing vision of the fiber-optic tip were selected.

660 All trees used in the experiment were young adult to mature adult trees. Tree  
661 ages were not accounted for – the only requirement was that dense patches of  
662 foliage could be located on the tree crowns.

663 • The  $\tau R$  spectra in the frontal sunlit area of trees with vertical reference planes  
664 normal to transient solar azimuth (SAz) directions were sampled to examine  
665 intraspecific and interspecific statistical means.

666 •  $\tau R$  spectra samples with horizontal views of the fiber-optic tip were selected  
667 for statistical mean.

668 • Solar altitude  $\alpha = 45^\circ$  was chosen as a benchmark condition to compare  
669 interspecies difference. Measured transreflectance spectra at solar altitudes  
670 different from  $45^\circ$  were converted to equivalent transreflectance spectra at  $\alpha =$   
671  $45^\circ$ , considering linear correlations of transreflectance spectra with  $\alpha$  within  
672 species (referring to section 4.2.3).

673 • Usually, at least 3 effective transreflectance spectra were sampled for an  
674 individual tree.

675 • At least 5 trees were sampled within a species (except the *Acer platanoides* as  
676 only four individual trees were considered) to estimate intraspecific statistical  
677 mean.

678 • Through chlorophyll fluorescence and heat stress measurements of part of trees  
679 selected in the field tests, it showed that the trees were subject to mild or  
680 moderate physiological stress and no significant effect of the tree physiological

stress on the  $\tau R$  was observed. Similarly, no significant effect of different urbanised settings of tested trees (planted on paved/sealed surfaces or on green lawns/parks) on the  $\tau R$  was found. Thus, the impact of heat stress and different urbanized settings of trees was not assessed in intraspecific statistical analysis.

#### 4.4.2 Interspecific difference of infrared radiative performance levels

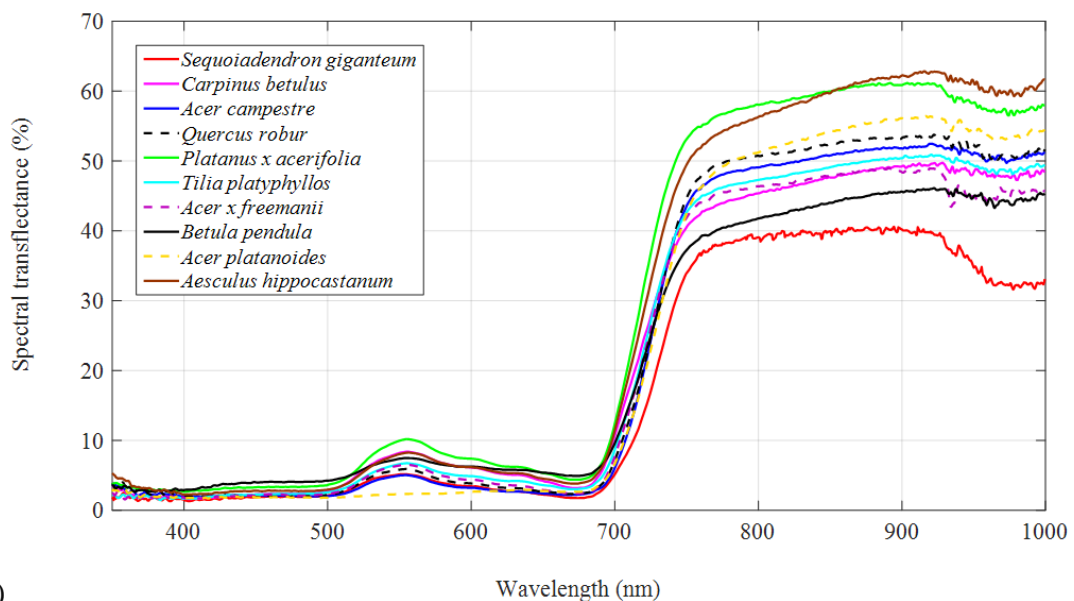
Here we use *Carpinus betulus* as an example to display the process of estimating statistical mean  $\tau R$  spectrum within a species. In total, effective  $\tau R$  spectrum samples of 9 *Carpinus betulus* trees were obtained.  $\tau R$  spectra of each individual tree were converted to the  $\tau R$  spectrum at  $\alpha = 45^\circ$ , and were then averaged to get a representative  $\tau R$  spectrum for the individual tree. The statistical mean  $\tau R$  spectrum of the *Carpinus betulus* species was estimated by using 9 representative  $\tau R$  spectra from 9 individual trees. Representative  $\tau R$  spectra of 5–9 trees were obtained for various tree species except the *Acer platanoides*, as only four individual trees were accessible for the copper Norway maple. Statistical mean  $\tau R$  spectra in the frontal sunlit area of the 10 tree species at  $\alpha = 45^\circ$  can be accessed in **Appendix A**.

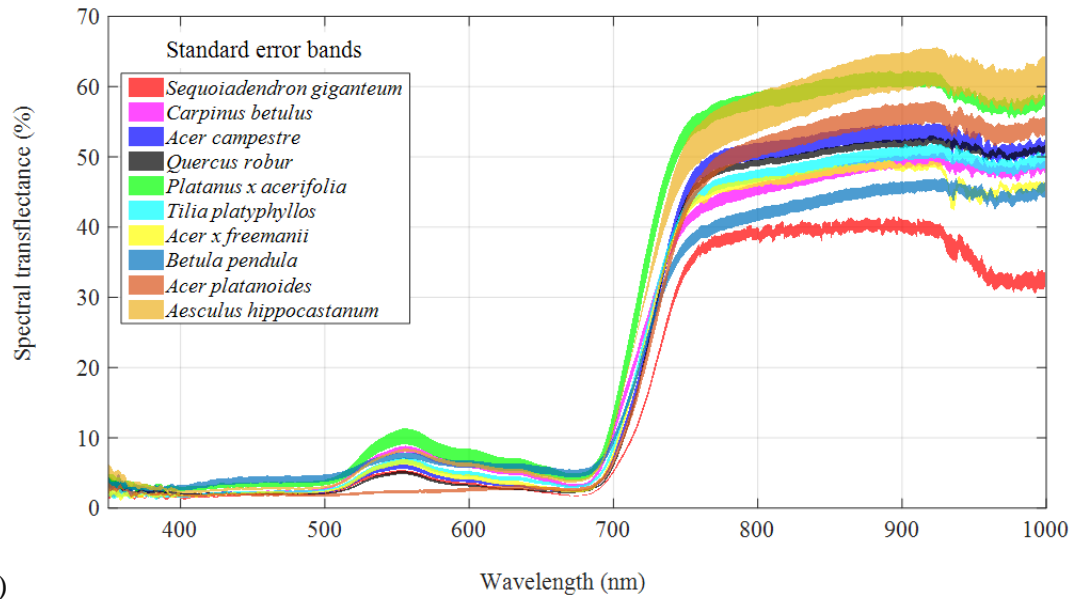
Figure 15 gathers the mean  $\tau R$  spectra and corresponding standard error bands of the 10 species, displaying interspecific performance difference. Table 1 lists statistical  $\tau R_{mean,800-900}$  (800–900 nm mean transfectance) values for the 10 tree species. It suggests that *Aesculus hippocastanum* and *Platanus x acerifolia* have the highest levels of IR radiative performance ( $\tau R_{mean,800-900}$ :  $0.597 \pm 0.027$  for *Aesculus hippocastanum*

and  $0.598 \pm 0.011$  for *Platanus x acerifolia*), followed by *Acer platanoides*, *Quercus robur*, *Acer campestre*, *Tilia platyphyllos*, *Carpinus betulus* (fastigate hornbeam) and *Acer x freemanii* successively on moderate levels ( $\tau R_{mean,800-900}$ : 0.475–0.540 and standard error within  $\pm 0.014$ ). *Betula pendula* has the second lowest levels of IR radiative performance ( $\tau R_{mean,800-900}$ :  $0.439 \pm 0.009$ ), presumably due to the fact that this native British tree species is usually observed with a small leaf size and apparently sparse leaf density. *Sequoiadendron giganteum* has the minimum levels of IR radiative performance ( $\tau R_{mean,800-900}$ :  $0.398 \pm 0.009$ ), resulting from the minimum leaf size (needle leaves) among the tested species.

Figure 16 shows images of tree leaves for the tested species corresponding to the radiative performance levels. Interesting is that, the tree species with large-sized leaves (such as *Platanus x acerifolia*, *Aesculus hippocastanum*) lead to maximum radiative performance levels, followed by the tree species with moderate-sized leaves (e.g. *Acer platanoides*, *Quercus robur*, *Acer campestre*, *Tilia platyphyllos*). *Carpinus betulus* has a leaf size smaller than other tree species on the moderate levels of radiative performance, resulting in a slightly lower performance level than the others except *Acer x freemanii*. Although the leaf size of *Acer x freemanii* seems to be close to that of *Quercus robur* and *Acer campestre*, branching structure of its leaves degrades its radiative performance level. From this perspective, each individual leaf of *Acer x freemanii* can be viewed as three leaflets which have similar leaf sizes as the *Carpinus betulus* leaves, resulting in the same radiative performance level as the latter.

*Sequoiadendron giganteum* with the smallest needle leaves among the 10 species has the minimum radiative performance level. It reveals that interspecific difference of infrared radiative performance levels strongly depends on leaf size when visibly dense foliage (no visible gaps and concave shapes in crown foliage) is observed on the tree crown contours. The finding confirms the viewpoint in [74] that species average leaf size is the most important determinant of self-shading that affected light capture. It provides insights on species selection for heat stress mitigation in urban microclimates. The impact of leaf size implies that disregarding tree ages (young or old) in sampling visually dense measuring patches (see 4.4.1) is appropriate, as mature trees of a specific-species tend to have reasonably similar leaf shape and leaf size.



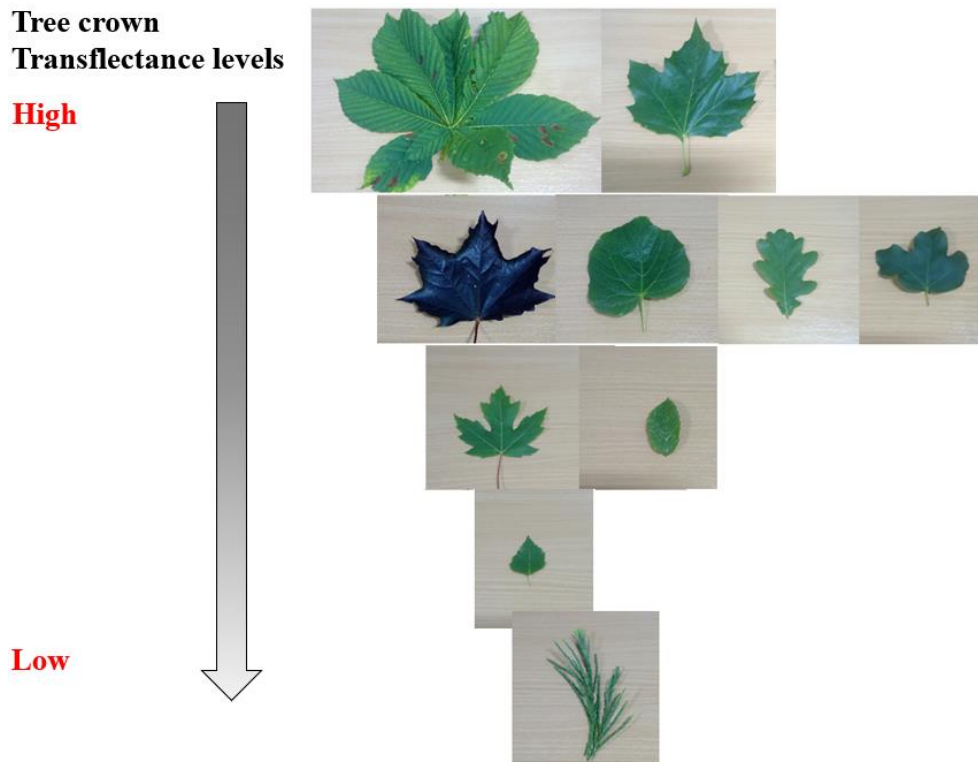


(b)

**Figure 15.** Interspecific radiative performance difference of ten tree species planted in the UK in terms of equivalent  $\tau R$  spectra at  $\alpha = 45^\circ$  (a) statistical mean  $\tau R$  spectra; (b) standard mean error bands.

**Table 1.** Statistical  $\tau R_{mean,800-900}$  for ten tree species

Tree species	$\tau R_{mean,800-900}$ (%)	Standard mean error (%)
<i>Sequoiadendron giganteum</i> (Giant sequoia)	39.8	$\pm 0.9$
<i>Carpinus betulus</i> (Fastigate hornbeam)	47.5	$\pm 0.9$
<i>Acer campestre</i> (Field maple)	52.2	$\pm 1.2$
<i>Quercus robur</i> (English oak)	50.7	$\pm 0.5$
<i>Platanus x acerifolia</i> (London plane)	59.8	$\pm 1.1$
<i>Tilia platyphyllos</i> (Large-leaved lime)	49.0	$\pm 0.9$
<i>Acer x freemanii</i> (Autumn blaze maple)	47.8	$\pm 0.6$
<i>Betula pendula</i> (Silver Birch)	43.9	$\pm 0.9$
<i>Acer platanoides</i> (Copper Norway maple)	54.0	$\pm 1.4$
<i>Aesculus hippocastanum</i> (Horse chestnut)	59.7	$\pm 2.7$



**Figure 16.** Radiative performance levels of multiple tree species classified by leaf size.

#### 4.4.3 Identification of intraspecific and interspecific differences in canopy transfectance by PCA

At the scale of crown patches significant differences in transfectance were observed between species, and these corresponded to different biological and spectral properties.

PCA of the 5 nm spectral bands ( $n = 130$  per tree crown) collected from 67 individual trees across 10 species revealed significant interspecific differences in crown transfectance for the 350-1000 nm range. The first two PCA axes explained 86.9% of the variance in spectral profiles and were significantly correlated with different canopy transfectance properties (see Table 2).

**Table 2.** Relationship between first five PCA axes and bands of canopy transfectance

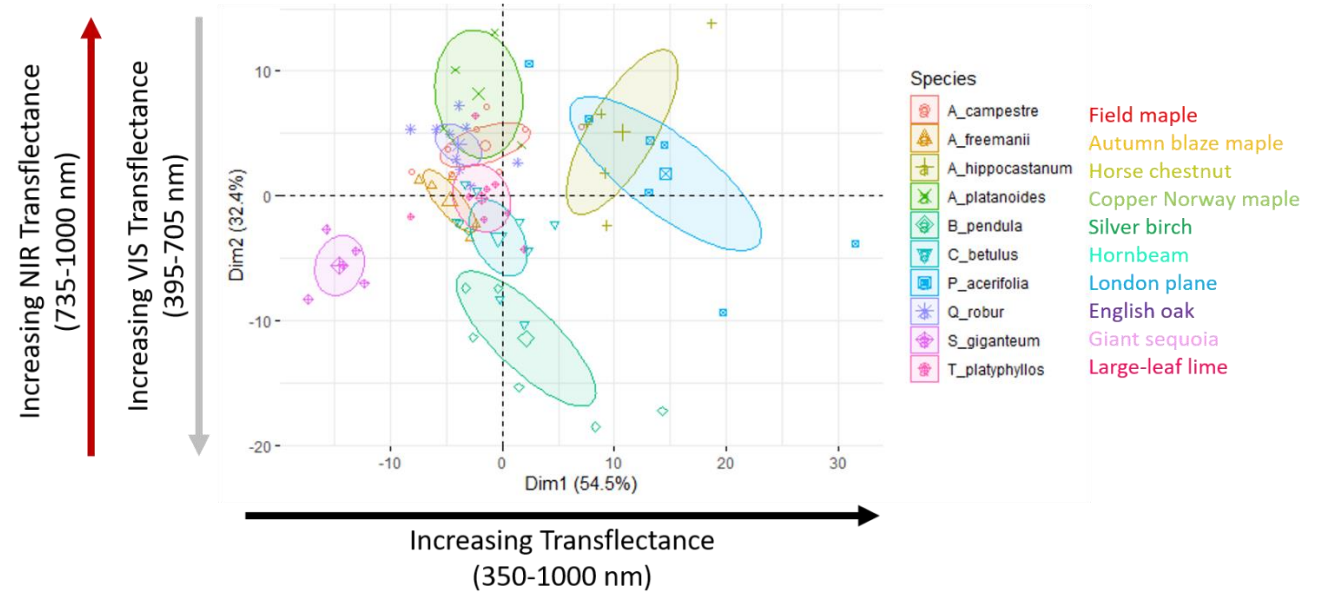
in the 350-1000 nm range across 67 individual trees drawn from 10 species

Axis	Variance explained	Cumulative variance	Positive correlation (r)	Negative correlation (r)
1	54.5%	54.5%	All except 370-375 nm	
2	32.4%	86.9%	735-1000 nm	395-705 nm
3	7.2%	94.1%	350-485 nm	525-615 & 695-735 nm
4	2.6%	96.7%	350-400 nm	
5	1.5%	98.2%		350-355 nm

Increasing transreflectance in the 350-1000 nm range corresponded to axis 1 of the PCA and was the most explanatory factor contributing to species differences (Table 2). The strongest correlation to axis 1 was found in the transition between VIS and NIR; large-leaved species (*Platanus x acerifolia* and *Aesculus hippocastanum*) were positively associated with this axis, whereas the only conifer (*Sequoiadendron giganteum*) was negatively associated (Species  $R^2 = 0.731$ ;  $P < 0.001$ ; Figure 17). Species were further sub-divided by their opposing interactions with NIR and components of VIS. Axis 2 corresponded with increasing NIR and decreasing VIS (violet, blue, yellow and orange) transreflectance (Table 2); *Acer platanoides* and *Quercus robur* were positively associated with this second axis, whereas *Sequoiadendron giganteum* and *Betula pendula* were negatively associated (Species  $R^2 = 0.663$ ;  $P < 0.001$ ; Figure 17). These results show that interspecific differences between the transreflectance profiles of commonly planted tree species are readily measurable and reveal key differences between species that can alter radiative performance in urban areas. As noted by Cavender-Barres *et al.* [70], leaf spectra can be considered as an integrated measure of phenotype; hence, further research into the impacts of environmental stress on canopy transreflectance could help in understanding the likely consequences of climate change



for urban areas, in terms of the impact on radiative performance and energy balance.



**Figure 17.** PCA of crown transreflectance spectra for axes 1 and 2 displaying differentiation between species and spectral properties correlated with each axis. Polygons represent 95% confidence interval for each species.

Furthermore, PERMANOVA using distance indices supported significant interspecific variation in crown transreflectance (Species, 999 permutations:  $Pseudo-F_{df\ 9,57} = 14.3$ ;  $R^2 = 0.694$ ;  $P = 0.001$ ). Although variability in crown transreflectance was observed within species, examination of the homogeneity of multivariate dispersion indicated that intraspecific variance (i.e. variance within species) was not significantly different between species (Species, 999 permutations:  $F_{df9,57} = 1.86$ ;  $P = 0.077$ ). Hence, crown transreflectance values were relatively similar within species, and the degree of multivariate variability within species was relatively similar between species. However, this observation was based on collecting spectral data from canopy patches with a

consistent, generally convex, coverage of leaves and avoiding gaps or concave patches. Therefore, intraspecific differences are expected to be strongly influenced by their unique life-history, including interactions with the biotic/abiotic environment and local government landscaping and management decisions.

## 5 Conclusions

Based on a tree crown spectroscopy measurement method established earlier, substantial in-situ tests of radiative performance of 10 tree species have been implemented in terms of the tree crown transfectance ( $\tau R$ ). Spatial distribution rules of the  $\tau R$  across tree crowns were identified. Infrared radiative performance difference of the 10 tree species (*Sequoiadendron giganteum*, *Carpinus betulus*, *Acer campestre*, *Quercus robur*, *Platanus x acerifolia*, *Tilia platyphyllos*, *Acer x freemanii*, *Betula pendula*, *Acer platanoides*, *Aesculus hippocastanum*) commonly planted in the UK was statistically determined in terms of  $\tau R$  spectra in frontal sunlit area of the trees, by converting  $\tau R$  spectra on the same benchmark of solar altitude  $\alpha = 45^\circ$ . Main findings are as follows:

- Mean leaf reflectance spectra of various tree species have a minor difference between each other, with spectral reflectance deviations of  $\pm 4\%$ , contrasting to substantial differences in spatial distribution of the  $\tau R$  across the tree crowns.
- Impact factors of tree crown transfectance (sparse foliage with gaps in crown foliage, concave and convex shapes in crown foliage, solar time): Visibly non-uniform foliage distribution in the measuring patches, such as sparse foliage

with gaps and concave shapes degrades  $\tau R$  levels to different extents. Impact of sparse foliage with gaps and concave shapes on  $\tau R$  is complicated and heterogeneous.  $\tau R$  in the frontal sunlit area of trees for a specific species is linearly correlated with solar altitude on sunny days, allowing this to be used as a benchmark for comparing differences in intraspecific and interspecific performance.

- Spatial distribution rules: The primary tree crown  $\tau R$  in a horizontal loop around tree crowns appears in the frontal sunlit area (SAz direction), followed by those in SAz+90° and SAz +270° clockwise directions. The  $\tau R$  on the tree shade side (SAz+180° clockwise) has the lowest levels).  $\tau R$  distributions in two typically vertical loops perpendicular to each other confirm that the strongest tree crown transfection was found primarily towards sky on the sunlit side of trees rather than towards zenith. The direction of the maximum transfection in the frontal sunlit area varies with solar time and depends on the polar angle consisting of solar azimuth and solar altitude. The  $\tau R$  beneath trees has the minimum level.
- Interspecific difference comparison of infrared radiative performance levels indicates that tree species such as *Platanus x acerifolia* and *Aesculus hippocastanum* with large-sized leaves, lead to maximum radiative performance levels, followed by the tree species with moderate-sized leaves (e.g. *Acer platanoides*, *Acer campestre*, *Quercus robur*, *Tilia platyphyllos*). *Sequoiadendron giganteum* has the minimum radiative performance level,

mainly due to its small needle leaves. It reveals that interspecific difference of the infrared radiative performance levels strongly depends on leaf size if visually dense foliage (no obvious gaps in foliage and no concave shapes) is observed on the tree crown contours.

The findings provide insights on understanding radiative interactions between urban trees and the surrounding built environments.

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## Appendix A. Supplementary materials

Supplementary data associated with this article is openly available in the University of Reading Research Data Archive, in the online version, at: <http://dx.doi.org/10.17864/1947.231>. Metadata will be available by request.

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## References

- [1] R. Garcia-Herrera, J. Díaz, R.M. Trigo, J. Luterbacher, E.M. Fischer, A review of the european summer heat wave of 2003, *Crit. Rev. Environ. Sci. Technol.* 40 (2010) 267–306. <https://doi.org/10.1080/10643380802238137>.
- [2] IPCC, Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change, 2014.
- [3] A. Gasparrini, B. Armstrong, The impact of heat waves on mortality, *Epidemiology*. (2011). <https://doi.org/10.1097/EDE.0b013e3181fdcd99>.
- [4] Y. Guo, A. Gasparrini, B.G. Armstrong, *et al.*, Heat wave and mortality: A multicountry, multicomunity study, *Environ. Health Perspect.* (2017). <https://doi.org/10.1289/EHP1026>.
- [5] E.M. Fischer, R. Knutti, Anthropogenic contribution to global occurrence of heavy-precipitation and high-temperature extremes, *Nat. Clim. Chang.* 5 (2015) 560–564. <https://doi.org/10.1038/nclimate2617>.
- [6] L. Zhao, M. Oppenheimer, Q. Zhu, J.W. Baldwin, K.L. Ebi, E. Bou-Zeid, K.

- 879 Guan, X. Liu, Interactions between urban heat islands and heat waves, *Environ.*  
880 *Res. Lett.* 13 (2018) 034003. <https://doi.org/10.1088/1748-9326/aa9f73>.
- 881 [7] C.C. Konijnendijk, K. Nilsson, T.B. Randrup, J. Schipperijn, Urban forests and  
882 trees: A reference book, 2005. <https://doi.org/10.1007/3-540-27684-X>.
- 883 [8] R. Upreti, Z.H. Wang, J. Yang, Radiative shading effect of urban trees on cooling  
884 the regional built environment, *Urban For. Urban Green.* 26 (2017) 18–24.  
885 <https://doi.org/10.1016/j.ufug.2017.05.008>.
- 886 [9] Z. Zou, Y. Yang, G.Y. Qiu, Quantifying the evapotranspiration rate and its  
887 cooling effects of urban hedges based on three-temperature model and infrared  
888 remote sensing, *Remote Sens.* 11 (2019) 1–18.  
889 <https://doi.org/10.3390/rs11020202>.
- 890 [10] M. Taleghani, Outdoor thermal comfort by different heat mitigation strategies-  
891 A review, *Renew. Sustain. Energy Rev.* 81 (2018) 2011–2018.  
892 <https://doi.org/10.1016/j.rser.2017.06.010>.
- 893 [11] L. Kong, K.K.L. Lau, C. Yuan, Y. Chen, Y. Xu, C. Ren, E. Ng, Regulation of  
894 outdoor thermal comfort by trees in Hong Kong, *Sustain. Cities Soc.* 31 (2017)  
895 12–25. <https://doi.org/10.1016/j.scs.2017.01.018>.
- 896 [12] Z.H. Wang, X. Zhao, J. Yang, J. Song, Cooling and energy saving potentials of  
897 shade trees and urban lawns in a desert city, *Appl. Energy.* 16 (2016) 437–444.  
898 <https://doi.org/10.1016/j.apenergy.2015.10.047>.
- 899 [13] S. Gillner, J. Vogt, A. Tharang, S. Dettmann, A. Roloff, Role of street trees in  
900 mitigating effects of heat and drought at highly sealed urban sites, *Landsc. Urban*

- 901 Plan. 143 (2015) 33–42. <https://doi.org/10.1016/j.landurbplan.2015.06.005>.
- 902 [14] H. Lee, H. Mayer, L. Chen, Contribution of trees and grasslands to the mitigation  
903 of human heat stress in a residential district of Freiburg, Southwest Germany,  
904 *Landsc. Urban Plan.* 148 (2016) 37–50.  
905 <https://doi.org/10.1016/j.landurbplan.2015.12.004>.
- 906 [15] A.S. Yang, Y.H. Juan, C.Y. Wen, C.J. Chang, Numerical simulation of cooling  
907 effect of vegetation enhancement in a subtropical urban park, *Appl. Energy*. 192  
908 (2017) 178–200. <https://doi.org/10.1016/j.apenergy.2017.01.079>.
- 909 [16] D.E. Bowler, L. Buyung-Ali, T.M. Knight, A.S. Pullin, Urban greening to cool  
910 towns and cities: A systematic review of the empirical evidence, *Landsc. Urban*  
911 *Plan.* 97 (2010) 147–155. <https://doi.org/10.1016/j.landurbplan.2010.05.006>.
- 912 [17] W. Zhou, J. Wang, M.L. Cadenasso, Effects of the spatial configuration of trees  
913 on urban heat mitigation: A comparative study, *Remote Sens. Environ.* 195 (2017)  
914 1–12. <https://doi.org/10.1016/j.rse.2017.03.043>.
- 915 [18] S. Sodoudi, H. Zhang, X. Chi, F. Müller, H. Li, The influence of spatial  
916 configuration of green areas on microclimate and thermal comfort, *Urban For.*  
917 *Urban Green.* 34 (2018) 85–96. <https://doi.org/10.1016/j.ufug.2018.06.002>.
- 918 [19] Z. Tan, K.K.L. Lau, E. Ng, Urban tree design approaches for mitigating daytime  
919 urban heat island effects in a high-density urban environment, *Energy Build.* 114  
920 (2016) 265–274. <https://doi.org/10.1016/j.enbuild.2015.06.031>.
- 921 [20] T. Zölch, M.A. Rahman, E. Pfleiderer, G. Wagner, S. Pauleit, Designing public  
922 squares with green infrastructure to optimize human thermal comfort, *Build.*

- 923 Environ. 149 (2019) 640–654. <https://doi.org/10.1016/j.buildenv.2018.12.051>.
- 924 [21] E. Jamei, P. Rajagopalan, M. Seyedmahmoudian, Y. Jamei, Review on the impact  
925 of urban geometry and pedestrian level greening on outdoor thermal comfort,  
926 Renew. Sustain. Energy Rev. 54 (2016) 1002–1017.  
927 <https://doi.org/10.1016/j.rser.2015.10.104>.
- 928 [22] Y. Wang, U. Berardi, H. Akbari, Comparing the effects of urban heat island  
929 mitigation strategies for Toronto, Canada, Energy Build. 114 (2016) 2–19.  
930 <https://doi.org/10.1016/j.enbuild.2015.06.046>.
- 931 [23] Y. Liu, D.J. Harris, Effects of shelterbelt trees on reducing heating-energy  
932 consumption of office buildings in Scotland, Appl. Energy. 85 (2008) 115–127.  
933 <https://doi.org/10.1016/j.apenergy.2007.06.008>.
- 934 [24] J.L. Moss, K.J. Doick, S. Smith, M. Shahrestani, Influence of evaporative  
935 cooling by urban forests on cooling demand in cities, Urban For. Urban Green.  
936 37 (2019) 65–73. <https://doi.org/10.1016/j.ufug.2018.07.023>.
- 937 [25] C.M. Hsieh, J.J. Li, L. Zhang, B. Schwegler, Effects of tree shading and  
938 transpiration on building cooling energy use, Energy Build. 159 (2018) 382–397.  
939 <https://doi.org/10.1016/j.enbuild.2017.10.045>.
- 940 [26] N. Bréda, V. Badeau, Forest tree responses to extreme drought and some biotic  
941 events: Towards a selection according to hazard tolerance?, Comptes Rendus -  
942 Geosci. 340 (2008) 651–662. <https://doi.org/10.1016/j.crte.2008.08.003>.
- 943 [27] C. Calfapietra, J. Peñuelas, Ü. Niinemets, Urban plant physiology: Adaptation-  
944 mitigation strategies under permanent stress, Trends Plant Sci. 20 (2015) 72–75.



945 <https://doi.org/10.1016/j.tplants.2014.11.001>.

946 [28] R. Teskey, T. Wertin, I. Bauweraerts, M. Ameye, M.A. McGuire, K. Steppe,  
 947 Responses of tree species to heat waves and extreme heat events, *Plant Cell*  
 948 *Environ.* 38 (2015) 1699–1712. <https://doi.org/10.1111/pce.12417>.

949 [29] I. Bauweraerts, M. Ameye, T.M. Wertin, M.A. McGuire, R.O. Teskey, K. Steppe,  
 950 Water availability is the decisive factor for the growth of two tree species in the  
 951 occurrence of consecutive heat waves, *Agric. For. Meteorol.* 189–190 (2014) 19–  
 952 29. <https://doi.org/10.1016/j.agrformet.2014.01.001>.

953 [30] S. Leuzinger, R. Vogt, C. Körner, Tree surface temperature in an urban  
 954 environment, *Agric. For. Meteorol.* 150 (2010) 56–62.  
 955 <https://doi.org/10.1016/j.agrformet.2009.08.006>.

956 [31] S. Zheng, J.M. Guldman, Z. Liu, L. Zhao, Influence of trees on the outdoor  
 957 thermal environment in subtropical areas: An experimental study in Guangzhou,  
 958 China, *Sustain. Cities Soc.* 42 (2018) 482–497.  
 959 <https://doi.org/10.1016/j.scs.2018.07.025>.

960 [32] M.A. Irmak, S. Yilmaz, E. Mutlu, H. Yilmaz, Assessment of the effects of  
 961 different tree species on urban microclimate, *Environ. Sci. Pollut. Res.* 25 (2018)  
 962 15802–15822. <https://doi.org/10.1007/s11356-018-1697-8>.

963 [33] C.Y. Park, D.K. Lee, E.S. Krayenhoff, H.K. Heo, S. Ahn, T. Asawa, A.  
 964 Murakami, H.G. Kim, A multilayer mean radiant temperature model for  
 965 pedestrians in a street canyon with trees, *Build. Environ.* 141 (2018) 298–309.  
 966 <https://doi.org/10.1016/j.buildenv.2018.05.058>.

- 967 [34] C.Y. Park, D.K. Lee, E.S. Krayenhoff, H.K. Heo, J.H. Hyun, K. Oh, T.Y. Park,  
968 Variations in pedestrian mean radiant temperature based on the spacing and size  
969 of street trees, *Sustain. Cities Soc.* 48 (2019) 1–9.  
970 <https://doi.org/10.1016/j.scs.2019.101521>.
- 971 [35] T. Zölch, J. Maderspacher, C. Wamsler, S. Pauleit, Using green infrastructure for  
972 urban climate-proofing: An evaluation of heat mitigation measures at the micro-  
973 scale, *Urban For. Urban Green.* 20 (2016) 305–316.  
974 <https://doi.org/10.1016/j.ufug.2016.09.011>.
- 975 [36] T.E. Morakinyo, K.K.L. Lau, C. Ren, E. Ng, Performance of Hong Kong's  
976 common trees species for outdoor temperature regulation, thermal comfort and  
977 energy saving, *Build. Environ.* 137 (2018) 157–170.  
978 <https://doi.org/10.1016/j.buildenv.2018.04.012>.
- 979 [37] D. Armson, M.A. Rahman, A.R. Ennos, A comparison of the shading  
980 effectiveness of five different street tree species in Manchester, UK, *Arboric.*  
981 *Urban For.* 39 (2013) 157–164.
- 982 [38] M.A. Rahman, D. Armson, A.R. Ennos, A comparison of the growth and cooling  
983 effectiveness of five commonly planted urban tree species, *Urban Ecosyst.* 18  
984 (2015) 371–389. <https://doi.org/10.1007/s11252-014-0407-7>.
- 985 [39] L. Zhang, Q. Zhan, Y. Lan, Effects of the tree distribution and species on outdoor  
986 environment conditions in a hot summer and cold winter zone: A case study in  
987 Wuhan residential quarters, *Build. Environ.* 130 (2018) 27–39.  
988 <https://doi.org/10.1016/j.buildenv.2017.12.014>.

- 989 [40] R. Sanusi, D. Johnstone, P. May, S.J. Livesley, Microclimate benefits that  
 990 different street tree species provide to sidewalk pedestrians relate to differences  
 991 in Plant Area Index, *Landsc. Urban Plan.* 157 (2017) 502–511.  
 992 <https://doi.org/10.1016/j.landurbplan.2016.08.010>.
- 993 [41] X. Chen, P. Zhao, Y. Hu, *et al.*, Canopy transpiration and its cooling effect of  
 994 three urban tree species in a subtropical city- Guangzhou, China, *Urban For.*  
 995 *Urban Green.* 43 (2019) 126368. <https://doi.org/10.1016/j.ufug.2019.126368>.
- 996 [42] M. Jiao, W. Zhou, Z. Zheng, *et al.*, Patch size of trees affects its cooling  
 997 effectiveness: A perspective from shading and transpiration processes, *Agric. For.*  
 998 *Meteorol.* 247 (2017) 293–299. <https://doi.org/10.1016/j.agrformet.2017.08.013>.
- 999 [43] E. Litvak, H.R. McCarthy, D.E. Pataki, A method for estimating transpiration of  
 1000 irrigated urban trees in California, *Landsc. Urban Plan.* 158 (2017) 48–61.  
 1001 <https://doi.org/10.1016/j.landurbplan.2016.09.021>.
- 1002 [44] M.A. Rahman, A. Moser, A. Gold, T. Rötzer, S. Pauleit, Vertical air temperature  
 1003 gradients under the shade of two contrasting urban tree species during different  
 1004 types of summer days, *Sci. Total Environ.* 633 (2018) 100–111.  
 1005 <https://doi.org/10.1016/j.scitotenv.2018.03.168>.
- 1006 [45] J. Konarska, J. Uddling, B. Holmer, M. Lutz, F. Lindberg, H. Pleijel, S. Thorsson,  
 1007 Transpiration of urban trees and its cooling effect in a high latitude city, *Int. J.*  
 1008 *Biometeorol.* 60 (2016) 159–172. <https://doi.org/10.1007/s00484-015-1014-x>.
- 1009 [46] C. Wang, Z.H. Wang, C. Wang, S.W. Myint, Environmental cooling provided by  
 1010 urban trees under extreme heat and cold waves in U.S. cities, *Remote Sens.*

1011 Environ. 227 (2019) 28–43. <https://doi.org/10.1016/j.rse.2019.03.024>.

1012 [47] M.A. Rahman, A. Moser, T. Rötzer, S. Pauleit, Comparing the transpirational  
 1013 and shading effects of two contrasting urban tree species, Urban Ecosyst. 22  
 1014 (2019) 683–697. <https://doi.org/10.1007/s11252-019-00853-x>.

1015 [48] W.T.L. Chow, A.J. Brazel, Assessing xeriscaping as a sustainable heat island  
 1016 mitigation approach for a desert city, Build. Environ. 47 (2012) 170–181.  
 1017 <https://doi.org/10.1016/j.buildenv.2011.07.027>.

1018 [49] A. Middel, N. Chhetri, R. Quay, Urban forestry and cool roofs: Assessment of  
 1019 heat mitigation strategies in Phoenix residential neighborhoods, Urban For.  
 1020 Urban Green. 14 (2015) 178–186. <https://doi.org/10.1016/j.ufug.2014.09.010>.

1021 [50] C.Y. Park, D.K. Lee, E.S. Krayenhoff, *et al.*, Variations in pedestrian mean  
 1022 radiant temperature based on the spacing and size of street trees, Sustain. Cities  
 1023 Soc. (2019). <https://doi.org/10.1016/j.scs.2019.101521>.

1024 [51] L.V. de Abreu-Harbich, L.C. Labaki, A. Matzarakis, Effect of tree planting  
 1025 design and tree species on human thermal comfort in the tropics, Landsc. Urban  
 1026 Plan. 138 (2015) 99–109. <https://doi.org/10.1016/j.landurbplan.2015.02.008>.

1027 [52] Z. Tan, K.K.L. Lau, E. Ng, Planning strategies for roadside tree planting and  
 1028 outdoor comfort enhancement in subtropical high-density urban areas, Build.  
 1029 Environ. 120 (2017) 93–109. <https://doi.org/10.1016/j.buildenv.2017.05.017>.

1030 [53] Z. Wu, L. Chen, Optimizing the spatial arrangement of trees in residential  
 1031 neighborhoods for better cooling effects: Integrating modeling with in-situ  
 1032 measurements, Landsc. Urban Plan. 167 (2017) 463–472.

- 1033 <https://doi.org/10.1016/j.landurbplan.2017.07.015>.
- 1034 [54] Z. Wu, P. Dou, L. Chen, Comparative and combinative cooling effects of  
 1035 different spatial arrangements of buildings and trees on microclimate, *Sustain.*  
 1036 *Cities Soc.* (2019). <https://doi.org/10.1016/j.scs.2019.101711>.
- 1037 [55] Q. Zhao, D.J. Sailor, E.A. Wentz, Impact of tree locations and arrangements on  
 1038 outdoor microclimates and human thermal comfort in an urban residential  
 1039 environment, *Urban For. Urban Green.* 32 (2018) 81–91.  
 1040 <https://doi.org/10.1016/j.ufug.2018.03.022>.
- 1041 [56] M. Aminipouri, D. Rayner, F. Lindberg, *et al.*, Urban tree planting to maintain  
 1042 outdoor thermal comfort under climate change: The case of Vancouver’s local  
 1043 climate zones, *Build. Environ.* 158 (2019) 226–236.  
 1044 <https://doi.org/10.1016/j.buildenv.2019.05.022>.
- 1045 [57] Z.H. Wang, Monte Carlo simulations of radiative heat exchange in a street  
 1046 canyon with trees, *Sol. Energy.* 110 (2014) 704–713.  
 1047 <https://doi.org/10.1016/j.solener.2014.10.012>.
- 1048 [58] T.E. Morakinyo, Y.F. Lam, Simulation study on the impact of tree-configuration,  
 1049 planting pattern and wind condition on street-canyon’s micro-climate and  
 1050 thermal comfort, *Build. Environ.* 103 (2016) 262–275.  
 1051 <https://doi.org/10.1016/j.buildenv.2016.04.025>.
- 1052 [59] T. Eckmann, A. Morach, M. Hamilton, *et al.*, A. McNamee, A. Haripriyan, D.  
 1053 Castillo, S. Grandy, A. Kessi, Measuring and modeling microclimate impacts of  
 1054 *Sequoiadendron giganteum*, *Sustain. Cities Soc.* 38 (2018) 509–525.

1055 <https://doi.org/10.1016/j.scs.2017.12.028>.

1056 [60] S.H. Lee, S.U. Park, A vegetated urban canopy model for meteorological and  
 1057 environmental modelling, *Boundary-Layer Meteorol.* 126 (2008) 73–102.  
 1058 <https://doi.org/10.1007/s10546-007-9221-6>.

1059 [61] H. Simon, J. Lindén, D. Hoffmann, P. Braun, M. Bruse, J. Esper, Modeling  
 1060 transpiration and leaf temperature of urban trees – A case study evaluating the  
 1061 microclimate model ENVI-met against measurement data, *Landsc. Urban Plan.*  
 1062 174 (2018) 33–40. <https://doi.org/10.1016/j.landurbplan.2018.03.003>.

1063 [62] J. Deng, B.J. Pickles, A. Kavakopoulos, T. Blanus, C.H. Halios, S.T. Smith, L.  
 1064 Shao, Concept and methodology of characterising infrared radiative  
 1065 performance of urban trees using tree crown spectroscopy, *Build. Environ.* 157  
 1066 (2019) 380–390. <https://doi.org/10.1016/j.buildenv.2019.04.056>.

1067 [63] British Trees: native and non-native trees.  
 1068 [https://www.woodlandtrust.org.uk/visiting-woods/trees-woods-and-](https://www.woodlandtrust.org.uk/visiting-woods/trees-woods-and-wildlife/british-trees/)  
 1069 [wildlife/british-trees/](https://www.woodlandtrust.org.uk/visiting-woods/trees-woods-and-wildlife/british-trees/) (accessed June 21, 2019).

1070 [64] C.F. Dietrich, Uncertainty, calibration and probability: The statistics of scientific  
 1071 and industrial measurement, 2nd ed., New York, 1991.  
 1072 <https://doi.org/10.1201/9780203734759>.

1073 [65] R Core Team, R: A language and environment for statistical computing, (2019).  
 1074 <https://www.r-project.org/>.

1075 [66] W.H. Oksanen J, Blanchet FG, Friendly M, et al., vegan: Community Ecology  
 1076 Package. R package version 2.5-6, (2019). <https://cran.r->

1077 project.org/package=vegan.

1078 [67] M.F. Kassambara A, factoextra: Extract and Visualize the Results of Multivariate  
 1079 Data Analyses, (2017). <https://cran.r-project.org/package=factoextra>.

1080 [68] S. Lê, J. Josse, F. Husson, FactoMineR: An R package for multivariate analysis,  
 1081 J. Stat. Softw. 25 (2008) 1–18. <https://doi.org/10.18637/jss.v025.i01>.

1082 [69] P. Legendre, L. Legendre, Numerical Ecology, 3rd Editio, Elsevier, Amsterdam,  
 1083 Netherlands, 2012. [https://www.elsevier.com/books/numerical-](https://www.elsevier.com/books/numerical-ecology/legendre/978-0-444-53868-0)  
 1084 ecology/legendre/978-0-444-53868-0.

1085 [70] J. Cavender-Bares, J.E. Meireles, J.J. Couture, *et al.*, Associations of leaf spectra  
 1086 with genetic and phylogenetic variation in oaks: Prospects for remote detection  
 1087 of biodiversity, Remote Sens. 8 (2016). <https://doi.org/10.3390/rs8030221>.

1088 [71] ImageJ User Guide. <https://imagej.nih.gov/ij/docs/guide/index.html> (accessed  
 1089 September 20, 2019).

1090 [72] Reading, ENG, United Kingdom — Sunrise, Sunset, and Moon Times.  
 1091 <https://www.timeanddate.com/astronomy/uk/reading> (accessed June 10, 2019).

1092 [73] Horse chestnut leaf-mining moth. <https://www.rhs.org.uk/advice/profile?pid=533>  
 1093 (accessed September 8, 2019).

1094 [74] D.S. Falster, M. Westoby, Leaf size and angle vary widely across species: What  
 1095 consequences for light interception?, New Phytol. 158 (2003) 509–525.  
 1096 <https://doi.org/10.1046/j.1469-8137.2003.00765.x>.