

The global distribution of plants used by humans

Article

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28
29 **Abstract:**

30 Plants sustain human life. Understanding geographic patterns of the diversity of species used by
31 people is thus essential for the sustainable management of plant resources. Here, we investigate
32 the global distribution of 35,687 utilised plant species spanning ten use categories (e.g., food,
33 medicine, material). Our findings indicate general concordance between utilised and total plant
34 diversity, supporting the potential for simultaneously conserving species diversity and its
35 contributions to people. While Indigenous lands across Mesoamerica, the Horn of Africa, and

36 Southern Asia harbor a disproportionate diversity of utilised plants, the incidence of protected
37 areas is negatively correlated with utilised species richness. Finding mechanisms to preserve
38 areas containing concentrations of utilised plants and traditional knowledge must become a
39 priority for the implementation of the Kunming-Montreal Global Biodiversity Framework.

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42 **One Sentence Summary:**

43 Plants and their contributions to people are insufficiently protected globally, pointing to the need
44 to preserve biocultural hotspots.

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46

47 **Main Text:**

48 Biodiversity provides essential goods and services sustaining human life and well-being
49 (e.g., food, medicines, materials, fuel) (1, 2). The balance between humanity's needs and the
50 protection of the natural environment is nevertheless fragile, as increased consumption of
51 resources, global trade, land and sea use change, and socio-economic inequalities are having a
52 dramatic influence on biodiversity (3, 4). To minimise biodiversity loss, conservation biologists
53 have focused on identifying and prioritising regions of high species richness, endemism, and threat
54 (5). The "biodiversity hotspot" concept (6) assumes that species diversity is spatially congruent
55 with the contributions that it provides to people and therefore, protecting areas with the largest
56 concentrations of threatened species will also protect humanity indirectly (5). Moreover, as
57 biodiversity is most concentrated where human cultural diversity is highest, it is assumed that high
58 biocultural diversity is associated with high concentrations of species used by humans (7). Yet,
59 these assumptions lack empirical support, leading to growing calls for better integration of human-
60 nature interactions into conservation planning and implementation (3, 8–10), as highlighted by the
61 recently adopted Kunming-Montreal Global Biodiversity Framework (GBF) and the 2022
62 assessment report on the sustainable use of wild species of the Intergovernmental Science-Policy
63 Platform on Biodiversity and Ecosystem Services (IPBES) (2).

64 Plants are essential structuring components of ecosystems and human livelihoods (9, 11).
65 Although the geography of terrestrial plant diversity has been extensively investigated globally (6,
66 12, 13), our understanding of the distribution of ecosystem services and societal benefits provided
67 by plants is incipient, despite the importance of this information for decision-makers and local
68 stakeholders in supporting the sustainable development agenda (14, 15). Recent modelling efforts
69 have been dedicated to the global distribution of nature's contributions to people, including water
70 quality, crop pollination and carbon stocks (16, 17). However, the extent to which these
71 contributions relate to species diversity remains largely unknown, hampering progress towards a
72 more sustainable management of biodiversity. Assessing the global diversity and distribution of
73 plant species used by people is thus critical to better understand, manage and preserve both the
74 intrinsic and instrumental values of biodiversity (18).

75 **The global distribution of utilised plant species richness and endemism**

76 Most plant species can potentially be useful to people, but only a fraction of plant diversity
77 is currently known to be used. Here we consider utilised plants as vascular terrestrial species for

78 which material and non-material benefits to humans have been reported and made publicly
79 accessible (19, 20). By extracting information from twelve databases containing plant uses (Table
80 S1; (21)), we identified 35,687 utilised species and assembled >11 million georeferenced
81 occurrence records to map their global distribution (i.e., native and introduced ranges) (Figs. S1-
82 2; (19)). We built species distribution models for each utilised species and stacked the resulting
83 maps to assess the global distribution of their potential richness (Figs. S3-6; (19)). We find the
84 highest concentrations of utilised plant species in the tropics (Fig. 1), but several temperate areas
85 also contain high native (e.g., China, the Himalayas; Fig. S7) and introduced richness (e.g.,
86 Western Europe, Eastern USA; Fig. 1). Despite large discrepancies in the sampling of species
87 geographic records (Fig. S1; (22)), these results match our estimates using coarser but more
88 complete independent distribution data from the World Checklist of Vascular Plants (WCVP (23);
89 Fig. S8), which provides additional support for our predictions.

90 Distribution patterns in species richness do not systematically match those of other
91 biodiversity indices considered important for conservation such as rarity or threat (5, 6). Therefore,
92 we also estimated the distribution of utilised plant species richness weighted by each species' range
93 size (i.e., weighted endemism) to identify areas with high concentrations of rare and potentially
94 irreplaceable species. We find that many areas with high richness of utilised plant species also
95 exhibit high endemism (e.g., Mesoamerica, Gulf of Guinea, Southern Africa, the Himalayas,
96 South-East Asia; Fig. 1, Fig. S8). Other areas also to emerge as exceptional centers of endemic
97 utilised plant species include California, Macaronesia, Madagascar, the Eastern Mediterranean,
98 the Western Ghats, Sri Lanka, Eastern Australia, and the Pacific islands. Conversely,
99 concentrations of endemic utilised species are relatively low across temperate areas. This confirms
100 that the high species richness observed in some temperate regions is due to a high concentration
101 of well-surveyed, widely distributed, and often introduced plant species of economic importance
102 (22, 24). Overall, the distribution of utilised plant endemism mirrors patterns observed across all
103 vascular plants, with higher endemism in areas with insularity and high topographic and
104 environmental heterogeneity (25, 26).

105 **The latitudinal distribution of utilised plant species and their different uses**

106 To refine our understanding of the geographic patterns underpinning the diversity of
107 utilised plant species, we disaggregated plant use reports into ten use categories, adapted from the
108 Economic Botany Data Collection Standards: human food (including beverages and additives),
109 vertebrate food (forage and fodder), invertebrate food (e.g., plants feeding honey bees or
110 silkworms), materials (e.g., wood, fiber), fuels (e.g., charcoal, alcohol), social uses (e.g., narcotics,
111 ritual, religious uses), poisons (for both vertebrates and invertebrates), medicines (for both human
112 and veterinary use), environmental uses (e.g., intercrops, windbreaks, ornamentals), and gene
113 sources (e.g., crop wild relatives) (19, 21). We find that latitudinal variation of utilised plant
114 species richness is broadly consistent for all ten use categories, with higher values in the tropics
115 gradually declining towards high latitudes (Fig. 2, Figs. S9-10). Therefore, areas with high
116 concentrations of utilised plant species also contain large numbers of species for each use type.
117 Despite the overall similarity in latitudinal patterns among use categories, there are striking
118 differences amongst temperate regions that are proportionally richer in plant species associated
119 with vertebrate food, social uses and poisons, compared to species-rich tropical environments that
120 contain proportionally more species associated with the most essential uses for human subsistence
121 (i.e., human food, material, and medicine). Concentrations of species used as gene sources are
122 exceptionally high around the equator and thus diverge from domestication centers originally

123 proposed by Vavilov (27). This is due to our consideration of a larger set of both domesticated
124 species and wild relatives of potential interest for contemporary breeding programmes (15).
125 Utilised plant weighted endemism also follows a latitudinal gradient with larger relative
126 concentrations of species at higher latitudes and consistent latitudinal variation among uses (Figs.
127 S11-13).

128 **Spatial concordance between utilised plant species, total plant species and human cultures**

129 Although quantitative evidence is scarce, areas of high plant diversity are expected to
130 contain more beneficial species to human populations (5). Our global analyses at (sub-)country
131 resolution indicate that utilised plant species richness is strongly positively associated with total
132 plant species richness (t-value=20.703, $P < 0.001$; Fig. S14-15; Table S2), and that this relationship
133 holds for all categories of uses and for endemism (Fig. S15; Table S3). It also highlights that large
134 proportions of the flora of relatively low diversity regions have documented uses (e.g.,
135 Scandinavia, Canada, Sahel), while smaller proportions of utilised species are reported across
136 megadiverse regions (e.g., Madagascar, Brazil, tropical Andes; Fig. S14). Future investigation will
137 be required to identify whether this pattern is due to sampling gaps in our database or in the wider
138 literature for these regions, or because the areas have reached a maximum capacity of utilised plant
139 species richness. Overall, our findings substantiates the combined importance of preserving
140 hotspots of plant diversity, which not only contain many unique species but also a considerable
141 diversity of potential services for humanity (1). Nevertheless, while the spatial concordance
142 between total plant diversity and utilised plant diversity is evident at a global scale, it is now crucial
143 to assess whether this pattern holds at smaller scale, where political decisions are taken and
144 management strategies implemented (28, 29).

145 Biodiversity and cultural diversity have been shown to be highly intertwined spatially,
146 giving rise to the notion of biocultural diversity (7, 30). Our data suggest that cultural diversity not
147 only correlates with total plant richness, but also covaries with utilised plant species richness (t-
148 value=5.743, $P < 0.001$; Fig. S14-15; Table S2) and inconsistently with endemism indices (Fig.
149 S15; Table S3). This finding supports previous hypotheses that geographic similarities between
150 biodiversity and cultural diversity could be due to increased competition or reduced necessity for
151 collaboration among human populations when biological resources (including plants) are widely
152 available, ultimately causing social separation and generating greater linguistic diversity (7, 30,
153 31). However, other historical, evolutionary, and environmental factors may also be involved, and
154 the identity and directionality of causal links for these correlations remain elusive and deserve
155 future investigation (7, 30).

156 **Indigenous Peoples and protected areas: preserving plant diversity and its contributions to** 157 **people**

158 Indigenous Peoples are particularly dependent on wild species for subsistence and well-
159 being, in addition to being critical custodians of both plant diversity and traditional knowledge
160 (32). Surprisingly, at a large spatial scale, we find that the lands over which the estimated >370
161 million Indigenous Peoples of the world exert traditional rights do not contain higher
162 concentrations of plant species with globally documented uses compared to neighbouring non-
163 Indigenous regions (Fig. 3, Fig. S17). This finding may reflect the fact that many Indigenous
164 Peoples have been dispossessed of their lands throughout history (33), including biologically
165 diverse areas, and that the largest remaining Indigenous territories are located in remote areas of
166 low primary productivity (e.g., Greenland, Siberia, the Tibetan plateau, the Sahara, Sahel, Central

167 Australia) (34). Exceptions include Indigenous lands located in multiple biocultural hotspots that
168 harbor higher utilised plant species richness/endemism than surrounding non-Indigenous regions:
169 Central America, the Horn of Africa, South and South-East Asia. While Indigenous areas
170 containing exceptionally high utilised plant diversity should be considered priorities for the joint
171 conservation of nature and traditional knowledge (34, 35), Indigenous lands containing fewer
172 species should not be overlooked given local populations may be particularly vulnerable to
173 changing environmental conditions and species losses (36). Fostering the engagement of
174 Indigenous, local and scientific knowledge systems will be essential for enhancing ethics and
175 actions towards protection at multiple scales (37).

176 Protected areas are at the forefront of global actions to preserve biodiversity and drive
177 sustainable development (38). However, despite currently covering ca. 17% of the Earth's
178 terrestrial surface, the protected area network contributes to the conservation of a small fraction of
179 plant diversity and ecosystem services (16). Spatial correlations between the proportion of land
180 that is protected, and utilised plant species richness and endemism indicate that regions with large
181 protected area networks do not contain higher numbers or more unique utilised plant species than
182 their non-protected counterparts (Fig. 3, Fig. S17). Indeed, although protected areas in Europe, the
183 Mediterranean, West Africa and the Horn of Africa contain more and more unique utilised plant
184 species than non-protected neighbouring regions, several regions exhibit higher relative richness
185 and endemism of utilised plant species outside of protected areas (e.g., Americas, Southern Africa,
186 Southeast Asia, Australia). Our results point to the urgent need of considering plant diversity and
187 its contributions to people in future area-based conservation planning (10, 29, 39), especially under
188 the ambitious Target 4 of the GBF, which aims to conserve biodiversity across 30% of global land
189 areas by 2030 (40). The latter also acknowledges the importance of “recognizing and respecting
190 the rights of Indigenous Peoples and local communities” and “ensuring that any sustainable use
191 [...] is fully consistent with conservation outcomes”. In this context, it is essential to strike an
192 appropriate balance between strictly protected areas that limit access to humans, and protected
193 areas that accommodate the sustainable use of natural resources by local populations while
194 preserving of their well-being and cultural heritage (41, 42).

195 Halting the overexploitation of species and ensuring their sustainable use has also been
196 highlighted as a key priority by the GBF, notably in Target 5. The sustainable management of a
197 few animal and plant species has proved to be an efficient tool for conservation (43, 44). However,
198 the sustainability of species use remains unknown across most plant diversity. Out of 2,800 utilised
199 plant species previously assessed by the International Union for Conservation of Nature (IUCN),
200 over one in three is considered to be at risk of global extinction (43). More than one in ten plant
201 species with a documented human food use in our study is also considered globally threatened
202 (45). While our findings show that utilised plant diversity remains largely under-protected in the
203 wild, most species (and their genetic diversity) additionally lack representation in ex-situ
204 collections such as seed banks and botanical gardens (46). Documenting and understanding the
205 diversity and distribution of plant species used by humans is thus crucial to implement
206 conservation strategies and develop plant-based solutions to address global societal challenges
207 such as hunger, diseases, and climate change (47–49). Our study aims to pave the way for efforts
208 towards reconciling human needs and biodiversity protection for a more sustainable future.

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596 Original conception: SP, IO, MD, TU, KJW

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598 Data acquisition: SP, IO, MD, RA, ACB, RC-L, CC, ZD-C, RG, SH, AJH, RL, WM, MN, KP,
599 GS, RMT, TRA, TU, KJW

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602 Data visualization: IO in close collaboration with SP

603 Data interpretation: SP, IO, AA, KJW

604 Writing – original draft: SP, IO

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606 MN, KP, GS, RMT, TRA, TU, AA, KJW

607

608 **Competing interests:**

609 Authors declare that they have no competing interests.

610

611 **Data and materials availability:**

612 Occurrence records, environmental layers, lists of names and uses of the 35,687 modelled utilised
613 plant species, and outputs from species distribution models are available on Zenodo and Github
614 (50, 51) alongside the R package *UsefulPlants*, which gathers codes, functions, tutorials, and
615 documentation allowing to reproduce analyses of the paper.

616

617 **Supplementary Materials:**

618 Materials and Methods

619 Figs. S1 to S17

620 Table S1 to S3

621 References (50-119)

622

623 **Figure legends:**

624 **Figure 1. Global species richness and endemism of plants with known uses by humans.** (A)
625 Utilised plant species richness corresponds to the sum of species occurrence probabilities predicted
626 in each ten arc-minutes (~20 km) pixel found across their native and introduced ranges. (B)
627 Utilised plant species endemism corresponds to the sum of species occurrence probabilities
628 predicted in each pixel weighted by the inverse of their range size, calculated as the sum of the
629 predicted probabilities within their study region (i.e., weighted endemism). High values are thus
630 associated with areas containing high concentrations of species with small geographic ranges.

631

632 **Figure 2. Latitudinal distribution of utilised plant species richness across t**

633 **en categories of plant uses.** The black curve on the left represents the latitudinal distribution of
634 all utilised plant species richness. The dendrogram on the top orders the ten use categories
635 according to the (dis-)similarity of their species richness latitudinal profiles. Black curves
636 underneath the dendrogram correspond to the species richness latitudinal profile for each use
637 category. The heatmap describes the latitudinal variation in the deviation of utilised plant species
638 richness for the ten plant use categories from total utilised plant species richness. Colors indicate
639 higher (green) or lower (purple) proportions in utilised plant species richness of a given use relative
640 to the total utilised plant species richness pattern. The bar chart underneath the heatmap shows the
641 number of species considered in each use category.

642

643 **Figure 3. Spatial correlations between utilised plant species richness and proportions of both**
644 **Indigenous lands and terrestrial protected areas.** Pearson's correlation coefficients were
645 computed across all values contained in 71 cells (~3,550km)-wide windows built around each
646 pixel. Pixel color indicates regions where utilised plant species richness is positively (green) or
647 negatively (pink) correlated with the proportion of Indigenous lands and terrestrial protected areas.
648 Regions crossed in beige indicate pixels containing more than 50% of Indigenous lands and
649 protected areas. All Indigenous lands and protected areas are thus not represented on the maps,
650 although they are all accounted for in the analyses. Frequencies of Pearson's correlation
651 coefficients found across the world are given in histograms. The median correlation across the
652 world is indicated by the black vertical line, while zero correlation is indicated by the red dashed
653 line. One-sample Wilcoxon signed rank tests were performed to assess whether median
654 correlations are significantly different from zero.

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Supplementary Materials for

The global distribution of plants used by humans

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The PDF file includes:

Materials and Methods
Figs. S1 to S17
Tables S1 to S3
References (50-119)

26 **Materials and Methods**

27 List of plant species used by humans

28 Analyses in this paper are based upon the “World Checklist of Useful Plant Species” (21). We use the term “utilised
29 plant” to refer to vascular terrestrial plant species for which material and non-material benefits to humans have been
30 documented and made publicly accessible. Uses may represent direct (e.g., food, medicine, material) or indirect
31 benefits (e.g., contributions to environmental services such as water, soil or air quality protection). Utilised plants may
32 be wild, introduced, cultivated or weeds. Their uses may have been reported at different time periods (from prehistory
33 to contemporary times), scales (from local to global, by individuals or societies), and economic levels (from local
34 personal use to commercial enterprise). Although commonly referred to as “useful” plants in the literature, we used
35 the term “utilised” here to avoid relying on the subjective concept of “usefulness” and rather focus on the simple fact
36 of a plant being known to be used. In fact, it is likely that all plants have an effective or potential use, but utilised
37 plants are those for which the use by humans is documented in the scientific literature. This definition, derived from
38 (52, 53), was employed in the 2020 State of the World’s Plants and Fungi report (11, 21).

39 The checklist was compiled using 12 available datasets, of which five have a global coverage: *Crop Wild*
40 *Relatives* ((54), <https://www.cwrdiversity.org>), Royal Botanic Gardens (RBG), Kew’s *Economic Botany Collection*
41 (<http://apps.kew.org/ecbot/search>), *Germplasm Resources Information Network* (GRIN) from the United States
42 Department of Agriculture (USDA; <https://www.ars-grin.gov>), *Medicinal Plant Names Services* version 8.2
43 (<http://www.kew.org/mpns> (55, 56)), and *Palms of the World Online* (<http://www.palmweb.org>); two cover primarily
44 South-East Asia: *Plant Resources of South-East Asia* (PROSEA (57)) and *the indigenous knowledge of New Guinea’s*
45 *useful plants* (58, 59); three span primarily Africa: *Plant Resources for Tropical Africa* (PROTA;
46 <https://www.prota4u.org/database>), *Useful Plants of West Tropical Africa* (UPWTA; (60)) and *Survey of Economic*
47 *Plants for Arid and Semi-Arid Lands* (<http://apps.kew.org/sepasalweb/sepaweb>); one has an American coverage:
48 *Plants for Malaria, Plants for Fever: medicinal species in Latin America* (61); and one project (Project MGU – the
49 Useful Plants Project (UPP; (62)) covers plants from both the Americas and Africa. A more comprehensive description
50 of the data sources is provided in Table S1 and (21). We acknowledge that these 12 datasets do not constitute a
51 complete list of all plant species used by humans across the world and knowledge gaps remain; nevertheless, this
52 selection of such a range of large-scale databases covering different continents, disciplines, and taxonomic groups
53 constitutes the most comprehensive source of utilised plant data. Because potential gaps, biases and uncertainties
54 inevitably remain in plant use data, and those have not been assessed previously at global scale contrary to plant
55 distribution data (e.g., (22)), we then adopted a cautious approach by mainly focusing our analyses on regional to
56 continental differences and latitudinal gradients, and by avoiding over-interpretation. Our study thus emphasises
57 the need for additional collection and collaboration efforts to document and understand the sampling distribution of plant
58 use data at multiple spatial and temporal scales.

59 Describing plant uses requires a standardisation of use categories and terminologies. Our study used an adapted
60 version of level 1 of the Economic Botany data standards (20) with the ten following plant use categories:

- 61 1. FOOD: Food for humans only, including beverages and food additives (45).
- 62 2. ANIMAL FOOD: Forage and fodder for vertebrate animals only.
- 63 3. INVERTEBRATE FOOD: Plants consumed by invertebrates used by humans, such as bees, silkworms, lac
64 insects and edible grubs.
- 65 4. MEDICINES: Both human and veterinary.
- 66 5. POISONS: Plants which are poisonous to both vertebrates and invertebrates, both accidentally and
67 intentionally, e.g., for hunting and fishing, molluscicides, herbicides, and insecticides.
- 68 6. MATERIALS: Woods, fibers, cork, cane, tannins, latex, resins, gums, waxes, oils, lipids, etc. and their
69 derived products.
- 70 7. FUELS: charcoal, petroleum substitutes, fuel alcohols, etc. Given the importance of energy plants for people
71 (61), those were distinguished from MATERIALS.
- 72 8. ENVIRONMENTAL USES: Examples include intercrops and nurse crops, ornamentals, barrier hedges,
73 shade plants, windbreaks, soil improvers, plants for revegetation and erosion control, wastewater purifiers,
74 indicators of the presence of metals, pollution, or underground water.
- 75 9. SOCIAL USES: Plants used for social purposes, which cannot be defined as food or medicine, for instance,
76 masticatories, smoking materials, narcotics, hallucinogens and psychoactive drugs, and plants with ritual or
77 religious significance.
- 78 10. GENE SOURCES: Wild relatives of major crops which may possess traits associated with biotic or abiotic
79 resistance and may be valuable for breeding programs.

80 While original sources of data provide information about the identity of species uses, they do not consistently inform
81 us about the parts of the plants that are used, the intensity of use, the identity and origin of the users, or the time at
82 which the use has been observed; therefore, questions associated with this information have not been addressed in this
83 study and will require further data collection at a smaller scale.

84 We standardised plant nomenclature and taxonomic following the World Checklist of Vascular Plants (WCVP)
85 (23), which contains data for >1,400,000 scientific names found in the literature and >360,000 accepted names from
86 >450 vascular plant families. To verify and correct species names, RBG Kew's semi-automated taxonomic names-
87 reconciliation procedure was used (<http://data1.kew.org/reconciliation/>) to match names of each species of each source
88 dataset against RBG Kew's taxonomic backbone on 26/11/2019. The backbone is built on the taxonomy from the
89 WCVP employing names drawn from the International Plant Names Index (<http://www.ipni.org>). Once synonyms and
90 non-accepted names were removed from the concatenated species lists, a total of 39,957 unique utilised plant species
91 was retrieved (21).

92 Species distribution modelling

93 Our modelling framework follows standards and best practices recently highlighted in the literature (64, 65), by
94 pursuing the five following steps: 1. gathering and processing occurrence and environmental data, 2. selecting the
95 modelling technique, 3. fitting models, 4. evaluating models, and 5. projecting species distribution.

96 1.a. Utilised plant occurrence data

97 We compiled occurrence records for each utilised plant species from the following seven databases: (i) Global
98 Biodiversity Information Facility (GBIF, <https://www.gbif.org>; accessed on January-March 2020) using the function
99 *get_gbifid* from the R package *taxize* for species name matching and then the function *occ_download* from the R
100 package *rgbif* for downloading records; (ii) Botanical information and Ecology Network (BIEN version 4.1,
101 <https://bien.nceas.ucsb.edu/bien>; accessed on January 2020) using the function *BIEN_occurrence_species* from the R
102 package *BIEN*; (iii) RAINBIO, which provides direct access to wild and native species occurrence records from Sub-
103 Saharan Africa, but from which we also obtained introduced and cultivated records directly from the authors (66); (iv)
104 speciesLink, which provides occurrences for species found in Brazil (spLink, <http://www.splink.org.br/>); (v) BioTIME
105 that provides local species assemblage data globally (67); (vi) Genesys from the global portal to information about
106 Plant Genetic Resources for Food and Agriculture (PGRFA), discarding records from markets and stores
107 (<https://www.genesys-pgr.org>); and (vii) the Crop Wild Relatives global occurrence database accessed via GBIF
108 (<http://www.cwrdiversity.org>).

109 Because georeferencing errors, imprecisions and biases are common in occurrence databases and can
110 significantly impact species distribution modelling (68), we first discarded possibly erroneous points found outside of
111 each species' known geographic range according to the WCVP. Known ranges correspond to both species' native and
112 introduced areas at the level-2 (regional or sub-continental scale) of the World Geographical Scheme for Recording
113 Plant Distribution (WGSRPD) developed by the International Working Group on Taxonomic Databases for Plant
114 Sciences (69). Here we used level-2 because level-1 (continental scale) was too coarse, and levels-3 and 4
115 (national/sub-national scales) were less reliable given a few countries have not been assessed by the WCVP and
116 probabilities of assigning false presences and absences are significantly higher at local scale. We then used metadata
117 information kept from primary data sources to only retain (i) records collected from 1945 onwards (70), and remove
118 those with (ii) latitude or longitude coordinates out of their ranges ([-90, 90] and [-180, 180] respectively); (iii)
119 coordinates uncertainty above 20 km; (iv) no decimal (i.e. rounded coordinates); (v) a count of individuals set to zero;
120 (vi) both coordinates set to zero; (vii) coordinates equal to each other (i.e. latitude equals longitude); (viii) coordinates
121 located within a buffer distance of ten kilometers around the centroids of countries, provinces, capitals, and botanical
122 institutions, and (ix) within a buffer distance of 1km around the GBIF headquarters. We used the *cc_** functions from
123 the R package *CoordinateCleaner* (71) to clean the data according to the nine points given above. To limit spatial
124 autocorrelation and avoid redundant information (72), we finally kept only one record per ten arc-minutes (~20 km)
125 grid cell for each species.

126 After cleaning, we did not retrieve any occurrence point for 4,270 utilised plant species. We thus performed
127 analyses for 35,687 species (89% of the original list), of which 6,461 (18%) have reported uses for human food, 4,087
128 (11%) for animal food, 971 (3%) for invertebrate food, 23,842 (67%) for medicines, 2,816 (8%) for poisons, 12,418
129 (35%) for materials, 2,348 (7%) for fuels, 8,314 (23%) for environmental uses, 2,385 (7%) for social uses, and 4,713

130 (13%) for gene sources (considering that species can have more than one use). The final occurrence dataset contained
131 >11 million records (Fig. S2).

132 1.b. Environmental data

133 We assembled an original set of 28 environmental variables. We used the 19 bioclimatic variables (temperature and
134 precipitation averages over the 1979–2013 period) provided by the Climatologies at High resolution for the Earth’s
135 land Surface Areas (CHELSA v.1.2). CHELSA has recently shown to outperform other climate data products for
136 predicting species distribution (73). We collected four terrain variables calculated from a Digital Elevation Model
137 (DEM) provided by the Global Multi-resolution Terrain Elevation Database (GMTED) (74): terrain roughness (R)
138 and Topographic Ruggedness Index (TRI), two measures of local terrain heterogeneity as the maximum difference
139 and the square root of averaged differences in elevation found within 3-by-3 cell size windows around each grid cell,
140 calculated using the *focal* function from the R package *raster*; Topographic Wetness Index (TWI), calculated with the
141 GRASS GIS (version 7.6.0) (75) *r.topidx* module, that combines local land slope and the upslope area converging to
142 a grid cell to describe hydrological processes contributing to soil wetness spatial patterns; and the land slope,
143 calculated with the function *Slope* from the *Spatial Analyst toolbox* of ARC/INFO GIS ESRI (version 10.5). We used
144 three soil related variables (averages across a range of 0-1m depth) obtained from the International Soil Reference and
145 Information Centre (ISRIC; <http://www.data.isric.org>): soil organic carbon stock (SOC), soil pH, and soil water
146 capacity (76). We also considered the Enhanced Vegetation Index (EVI) derived from MODIS time-series data
147 spanning the 2000–2013 period, extracted from the LEFT tool (77), and the Human Footprint Index representing the
148 anthropogenic impacts on the environment for the 1995–2004 period based on human population pressure, land use
149 and infrastructure, and access, and produced by the Wildlife Conservation Society (WCS) and the Columbia
150 University Center for International Earth Science Information Network (CIESIN) (78). The 28 variables were
151 available for time slices that do not always match with each other’s and occurrence data; however, we believe that
152 considering a wide set of climatic, vegetation, edaphic, topographic, and human-related variables would better help
153 unravel the distribution of utilised plants for the broad period going from mid-20th to early 21st century. For
154 homogeneity and to match our occurrence records, all layers were resampled at a ten arc-min resolution and masked
155 non-terrestrial lands based on the revised map of Terrestrial Ecoregions of the World (79), using the *Resample* and
156 *ExtractbyMask* modules from the geoprocessing toolboxes *Data management* and *Spatial Analyst* of ArcGIS.

157 From this pool of 28 variables, we then made a sub-selection intending to: (i) capture most of the variation
158 in environmental conditions across the world; (ii) limit multi-collinearity, and (iii) make the most sense ecologically
159 (80–82). We performed a Principal Component Analysis (PCA) using the function *dudi.pca* from the R package *ade4*
160 to assess how much of the variance is explained by the variables and correlations among them. Then, we carried out
161 a stepwise selection and a pairwise correlation analysis to exclude (multi-)collinear variables. The former consists in
162 iteratively computing the Variance Inflation Factor (VIF) (83) of each variable and excluding one variable at a time if
163 its VIF value exceeds a threshold of ten, until all remaining variables have VIF values below ten. The latter consists
164 in excluding variables with the highest VIF values out of pairs of highly correlated variables (i.e., Pearson’s correlation
165 coefficient $r > 0.7$). VIF analyses were conducted using functions *vifstep* and *vifcor* from the R package *usdm*. Based
166 on the combination of these analyses and criteria, we finally selected nine predictors: precipitation seasonality, mean
167 temperature of the coldest quarter, precipitation of the driest quarter, terrain roughness (R), Topographic Wetness
168 Index (TWI), Enhanced Vegetation Index (EVI), Soil Organic Carbon stock (SOC), mean soil pH and the human
169 footprint index. This diverse set of variables is well known to influence plant species distribution (84–86).

170 2. Algorithm selection

171 Occurrence records alone are unreliable to retrieve species richness patterns and macroecological transitions due to
172 the numerous geographic and taxonomic biases that they contain (22, 87). However, they have been shown to perform
173 better when modelled along environmental gradients (87). To retrieve species composition and richness across space,
174 we therefore decided to use a “stacked-species distribution” approach, which consists in stacking modelled species
175 distribution maps (88). Species Distribution Models (SDMs) are probabilistic models relating species occurrence to
176 environmental variables to project their distribution across space and/or time through an estimate of environmental
177 similarity (89). Many species distribution modelling techniques are available, but we decided to use MaxEnt version
178 3.4.1. as it is a commonly used model that (i) relies on presence-only data and (ii) has shown to be one of the methods
179 performing best in many different ecological contexts (81, 90). Because SDMs are sensitive to sample size (91),
180 simpler models are usually recommended when very few occurrence points are available for each species (92).

181 Geographic models (GMs) can provide a suitable alternative to SDMs since they do not rely on an environmental
182 niche approach but rather consider the spatial structure of occurrence locations. They can recover endogenous spatial
183 determinants such as dispersal limitation, which are particularly important factors shaping the distribution of small
184 range species (93). Therefore, we decided to fit a GM for species having less than ten occurrence records. The
185 minimum number of ten points to fit SDMs may prevent from overfitting, which often occurs when the number of
186 explanatory variables (nine in our case) exceeds the number of predictors (low degree of freedom). Because GMs
187 cannot be used for extremely low sample size (e.g., for very rare species), we simply rasterised occurrences for species
188 having less than three points using the function *fasterize* from the R package *fasterize*.

189 3. Model fitting

190 MaxEnt contrasts environmental conditions where species have been observed with conditions available or accessible
191 in their surroundings (i.e., background region). Background selection is such a critical step of the modelling process
192 (64) that we decided to individualise it for each species, as recommended in (94). We built an alpha-hull around
193 occurrence points, which is a generalization of the convex-hull particularly useful for estimating species ranges whose
194 habitat is irregularly shaped (95). Alpha-hulls were buffered using the “1/10th maximum” method to better account for
195 spatially structured populations (96). We then selected the biomes (79) that intersect with the alpha-hull. To avoid
196 extreme cases where a small alpha-hull intersects with biomes spreading over vast territories and provides an oversized
197 background, we implemented two conditions: (i) the number of occurrence points lying inside each intersected biome
198 must be higher or equal to ten, and (ii) the proportion of the biomes’ area covered by the alpha-hull must be higher or
199 equal to ten percent. If the conditions are not met for a particular biome, we then selected the ecoregion (79). Since
200 the alpha-hull algorithm was set up to exclude up to 5% of the records (i.e., geographic outliers), we also searched for
201 the ecoregion polygons occupied by these points and merged them back to the background. Thus, the background
202 region extends to the boundaries of the biomes or ecoregions where species were found and where other populations
203 could potentially occur.

204 Occurrence records are highly unevenly distributed in space and biased towards developed countries and
205 sampling priorities of major botanical institutions (22, 97). As this uneven geographic coverage may significantly
206 impact species distribution modelling, we accounted for this bias by (i) generating a biased prior map giving a non-
207 uniform weighting to background points (90, 98, 99), and (ii) integrating this map into the model building through the
208 MaxEnt argument *biasfile*. This bias map representing the effort in sampling utilised plants was generated in two
209 steps: (i) we used the *rasterize* function of the *raster* package in R to count the total number of utilised plant occurrence
210 records found in each ten arc-min grid cell (based on the original dataset, before removing points occurring in the
211 same cells), and (ii) we used the *kde2d* function from the R package *MASS* to interpolate these counts using a two-
212 dimensional kernel density estimator with a two degrees bandwidth value (Fig. S1).

213 We ran MaxEnt models for 28,235 (79%) utilised plant species using the *biasfile*, a maximum number of 500
214 iterations and 50,000 samples from the background area. We specified an automatic selection of response shapes
215 (hereafter called feature classes) and allowed MaxEnt to fit linear, quadratic, product (interaction) and hinge
216 (threshold-like response) feature classes to the data. We used the complementary log-log output (cloglog), which is a
217 more appropriate estimate of species probability of presence that avoids assumptions about species prevalence (100).

218 We fitted GMs for 5,464 (15%) utilised plant species using the Inverse Weighting Distance (IDW) method
219 (101). IDW is a spatial interpolation model that computes the probability of occurrence as a weighted average of
220 neighboring occurrences, with weights inversely proportional to the distance between locations. Distance-based GMs
221 thus assume that species are more likely to occur close to known locations than further away (93). Since IDW requires
222 absence data, we generated pseudo-absences to achieve a minimum prevalence of ten percent (i.e., 1:10 ratio between
223 the number of cells occupied by species records and the number of non-occupied cells in the study ecoregion). IDW
224 also requires a local neighborhood that we defined as the seven nearest locations using the function *geoIDW* from the
225 R package *dismo*. Occurrence records were directly rasterised for the remaining 1,988 species (6%) that had less than
226 three cleaned occurrence points.

227 4. Model evaluation

228 We used a species-specific tuning of the MaxEnt β regularization coefficient (hereafter called β multiplier) to assess
229 the performance of our SDMs, as recommended in the literature (98, 102). MaxEnt uses the β multiplier to prevent
230 the model from overfitting, notably by penalizing the use of overly complex feature classes. We use the masked
231 geographically structured approach (102) to explore a range of β multipliers based on (103, 104): 1(default), 2, 6, 10.

232 This method is a variant of the *k-fold* cross-validation that provides a better ability to detect overfitting and more
233 realistic estimates of model performance than other cross-validation approaches (102). We spatially segregated the
234 occurrence records into $k=3$ geographical bins with approximately the same number of points in each bin using a
235 customised algorithm using the *kmeans* function from the R package *stats*. Models were then trained iteratively using
236 $k-1$ ($= 2$) bins and tested on the third. At each iteration, we calculated for each β multipliers: (i) the corrected Akaike
237 Information criterion (AIC_c) (103, 105) using the function *ic* from the R package *rmaxent*, (ii) the tenth percentile of
238 the training omission rate (OR10) and (iii) the Area Under the Curve of the receiver operating characteristic on the
239 testing dataset (AUC) as exported automatically from MaxEnt (106), and (iv) the maximum of the True Skill Statistics
240 (TSS) (107) using the functions *prediction* and *performance* of the R package *ROCR*. We averaged these values over
241 the iterations and ranked as *best model* the one with the regularization multiplier giving the lowest AIC_c . When the
242 difference between the AIC_c values of the top models was not substantial (i.e., $\Delta AIC_c < 2$), we ranked them
243 successively by the lowest $OR10_{test}$, highest TSS_{test} and highest AUC_{test} (where the subscript test indicates that the
244 metric is calculated on the testing dataset). We therefore ranked the models according to their capacity to minimise
245 overfitting before accounting for their discriminatory ability, as recommended in the literature (102, 108).

246 GMs' predictions were assessed using the *leave-one-out* (LOO) method, a form of *k-fold* cross-validation
247 suitable when sample size is small (109). LOO consists in iteratively training the model using $n-1$ data points (where
248 n is the total number of points) and testing on the withheld point. The AUC was computed and averaged across the
249 iterations.

250 5. Projection of species distribution, richness, and endemism

251 Individual SDMs performed well according to AUC (mean AUC: 0.809 ± 0.121 ; Fig. S3), OR10 (mean OR10: 0.239
252 ± 0.137 ; Fig. S4), and TSS (mean TSS: 0.618 ± 0.160 ; Fig. S5) values, which indicate good discriminatory ability of
253 the models and minimised overfitting. GMs indicate fair overall performance (mean AUC: 0.741 ± 0.177 ; Fig. S6).
254 We therefore projected the probability of occurrence of each species from our SDMs and GMs onto the geographical
255 area used for model training (background region) based on species biomes and/or ecoregions and known total range
256 (i.e., native and introduced) or native range only according to the WCVP. Probabilities of one were assigned to the
257 pixels occupied by the 1,988 (6%) utilised plant species represented by less than three records.

258 To retrieve species richness across space, we used a "Stacked-Species Distribution Models" (S-SDMs) approach,
259 which consists in stacking individual species distribution maps obtained from SDMs (88). We summed species
260 occurrence probabilities instead of using thresholded (binary) maps, as this has been shown to provide better estimates
261 of species richness (110). We repeated this process considering species native and introduced ranges, and native ranges
262 only to retrieve two global maps of species richness. Areas containing a large proportion of species with restricted
263 ranges are potentially highly irreplaceable and do not match areas with high species richness, so they represent major
264 targets for conservation (5, 6, 111). For this reason, we also estimated endemism as the sum of each species'
265 occurrence probabilities weighted by the inverse of their range size calculated as the sum of the predicted probabilities
266 within their study region (112, 113). Finally, given that the location of the different uses of the 35,687 species remains
267 largely unknown, we stress that our study does not intend to map the uses of plants but rather the distribution of species
268 documented to be used somewhere in the world. Therefore, our results rather refer to a potential usage of the species
269 occurring in different regions of the world rather than their actual use in those regions. Similarly, temporal variation
270 in plant uses has not been quantified at such large taxonomic and spatial scales. Our results thus report a rather static
271 view of the global distribution of plants used by humans.

272 Although we used extensive information and accounted for sampling biases in our modelling framework, SDM
273 predictions can still suffer from data incompleteness and uncertainty (22). For this reason, we also mapped utilised
274 plant species richness and weighted endemism using independent data from the WCVP at the level-3 of the WGSRPD
275 (69). Level-3 provides plant species distribution data at a country scale for most of the world, except in a few large
276 countries for which information is available at a sub-country scale. While species richness is simply the count of
277 species found in each level-3 region, weighted endemism weighs each species by the total area of level-3 regions that
278 it occupies. To account for differences in surface area between regions, we estimated a scaling exponent that describes
279 the species-area relationship (SAR) between the counts of species (weighted or not) and the level-3 region areas (z -
280 values range between 0.101 (social uses) and 0.141 (all uses)). Then, we used this exponent to rescale species counts
281 for a standard area of 10,000 square kilometers as in (114). We executed this procedure for all utilised plant species
282 together and for each individual use category.

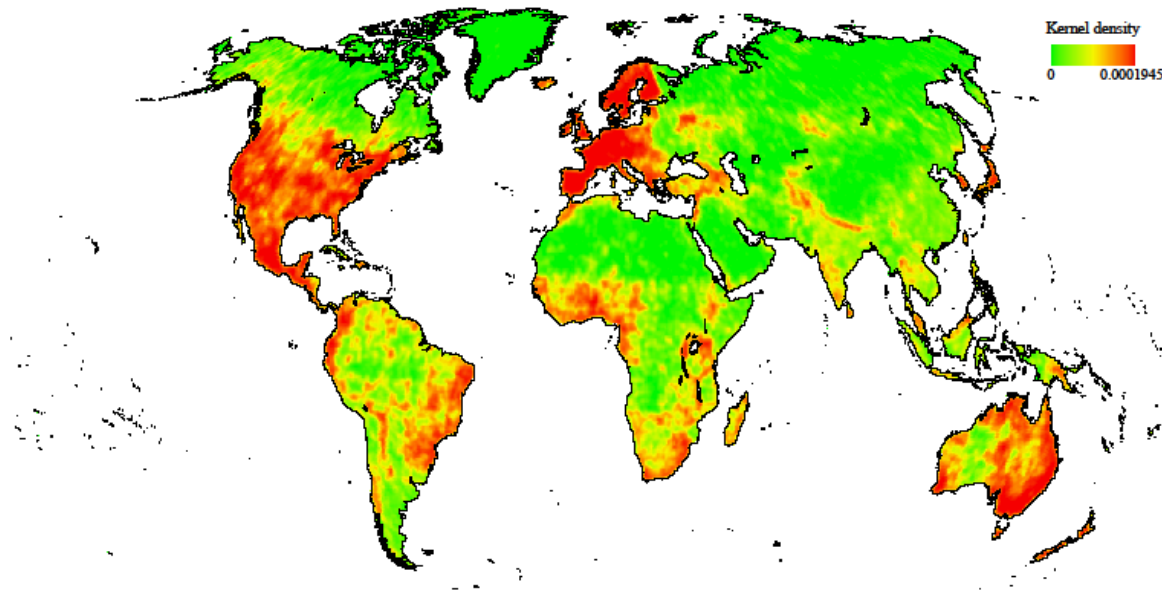
283 Analyses

284 We investigated the variation of utilised plant species richness and weighted endemism across latitude, a well-known
285 biogeographical gradient (115). To achieve this, we fitted Generalised Additive Models (GAM) between latitude and
286 species richness/weighted endemism values across all pixels of our global maps following the approach from (116).
287 Model fitting was performed using the function *bam* from the R package *mgcv* with our predicted estimates of species
288 richness and weighted endemism per grid cell as response variable and a penalised cubic regression spline on the
289 latitude of the grid cells as smooth term. Then, we predicted utilised plant species richness/endemism for each unique
290 latitude value and divided each prediction by the total sum of the predictions to obtain a relative measure of utilised
291 plant species richness/endemism along the latitudinal gradient. To better visualise differences among plant use
292 categories, we considered species richness and endemism across all utilised plants grouped together as a baseline and
293 then represented the deviation of the latitudinal patterns of each use category from this baseline. We used the function
294 *hclust* from the *stats* package to assess (dis-)similarities and clustering among latitudinal profiles of each use by
295 building dendrograms.

296 We assessed potential associations between distribution patterns in the diversity of all plant species, utilised
297 plant species, and human cultures. To our knowledge, no continuous map is available for both overall plant richness
298 and cultural diversity, thus we worked at a country scale. Estimates of human cultural diversity were compiled by
299 (117) for 256 countries. Estimates of total vascular plant species richness/endemism and utilised plant species
300 richness/endemism were compiled from the WCVP (23). Level-3 geographic data were aggregated at a country scale
301 for each individual species using a global map of countries provided by the Database of Global Administrative Areas
302 (GADM; <http://www.gadm.org>) version 3.6, before computing area-corrected richness and weighted endemism
303 indices as described previously (Figs. S15-16). Due to geographic mismatches between certain botanical regions and
304 countries, and the absence of cultural diversity data for certain territories, we only kept 163 countries for analyses.
305 Cultural diversity is a composite index made of three estimates: total number of languages, religions, and ethnic groups
306 per country (117–119). The cultural diversity index uses SAR to account for variable country sizes and ranges from 0
307 to 1, with values close to 0 indicating the least diverse countries and values close to 1 the most diverse ones. We
308 assessed relationships between human cultural diversity, total plant, and utilised plant species richness/endemism
309 using Generalised Least Squares (GLS) models with the *gls* function of the R package *nlme*. Unlike Ordinary Least
310 Squares (OLS) models, GLSs allow to accommodate for potential spatial (auto-)correlation in the data. We used the
311 latitude and longitude coordinates of the centroids of each country as spatial covariates, and tested three different
312 correlation functions: exponential, gaussian and spherical. We fitted our GLS models using the restricted maximum
313 likelihood method (REML) and selected the correlation structure with the lowest AIC. Analyses were repeated for all
314 utilised plants together and for each individual use category separately.

315 We assessed spatial associations between utilised plant species richness, endemism, Indigenous lands, and
316 protected areas at a finer resolution given the availability of raster maps for all variables. Indigenous lands represent
317 areas managed and/or controlled by Indigenous Peoples (as defined by the International Labour Organization (120)).
318 The global map was built based on 127 source documents and provides a percentage of each 50 km grid cell covered
319 by Indigenous lands across the world (34). The completeness of this data cannot be guaranteed; it represents known
320 Indigenous land areas based on publicly available geospatial data only. We used the World Database on Protected
321 Areas (WDPA) (121) to extract the percentage of all protected areas contained in each 50 km cell of the Indigenous
322 land grid. The WDPA is the most comprehensive global database of marine and terrestrial protected areas, comprising
323 both spatial data with associated attribute data. It includes protected areas that meet definitions set by the International
324 Union for Conservation of Nature (IUCN) and the Convention on Biological Diversity (CBD). Protected area coverage
325 was calculated using ArcGIS and ESRI's Modelbuilder. Utilised plant species richness and endemism maps obtained
326 from SDMs were aggregated to the same 50 km resolution using the *Resample* tool from the *Data Management*
327 toolbox in ArcGIS. Spatial correlations between utilised plant species richness/endemism and both Indigenous lands
328 and protected areas were assessed by computing Pearson's correlation coefficients across all values contained in 71
329 cells (~3,550km)-wide windows built around each pixel. This resulted in global maps indicating the strength and
330 direction of regional correlations between the different variables. The size of the moving window was selected so that
331 all pixels of the world map have a correlation value (i.e., all windows contain at least one Indigenous land and one
332 protected area) and our analysis provides a broad estimate of the geographical variation of the correlations at large
333 spatial scale. For this analysis we used the *focal* and *cor* functions from the *raster* and *stats* R packages, respectively.
334 All analyses and data processing were based on the R statistical software version 3.6.1, ArcGIS version 10.5 and
335 GRASS GIS version 7.6.0.

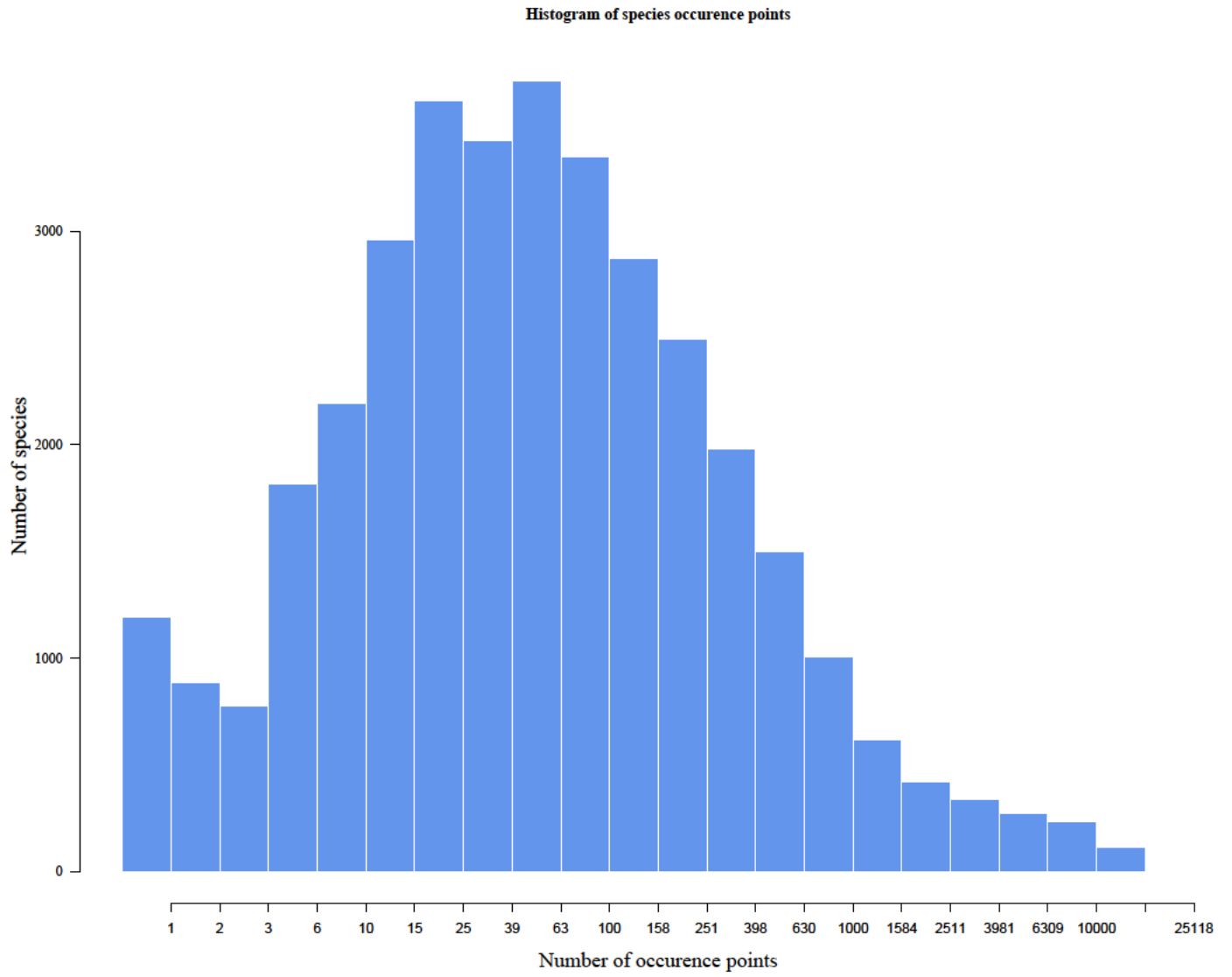
336



337

338 **Figure S1. Sampling intensity in occurrence records for all utilised plant species.** Red indicates high sampling
339 density, whereas green indicates low sampling density based on counting the number of occurrence records available
340 for all utilised plant species in each 400 km² grid cell and a kernel density probability approach providing a unitless
341 relative index of density.

342



343

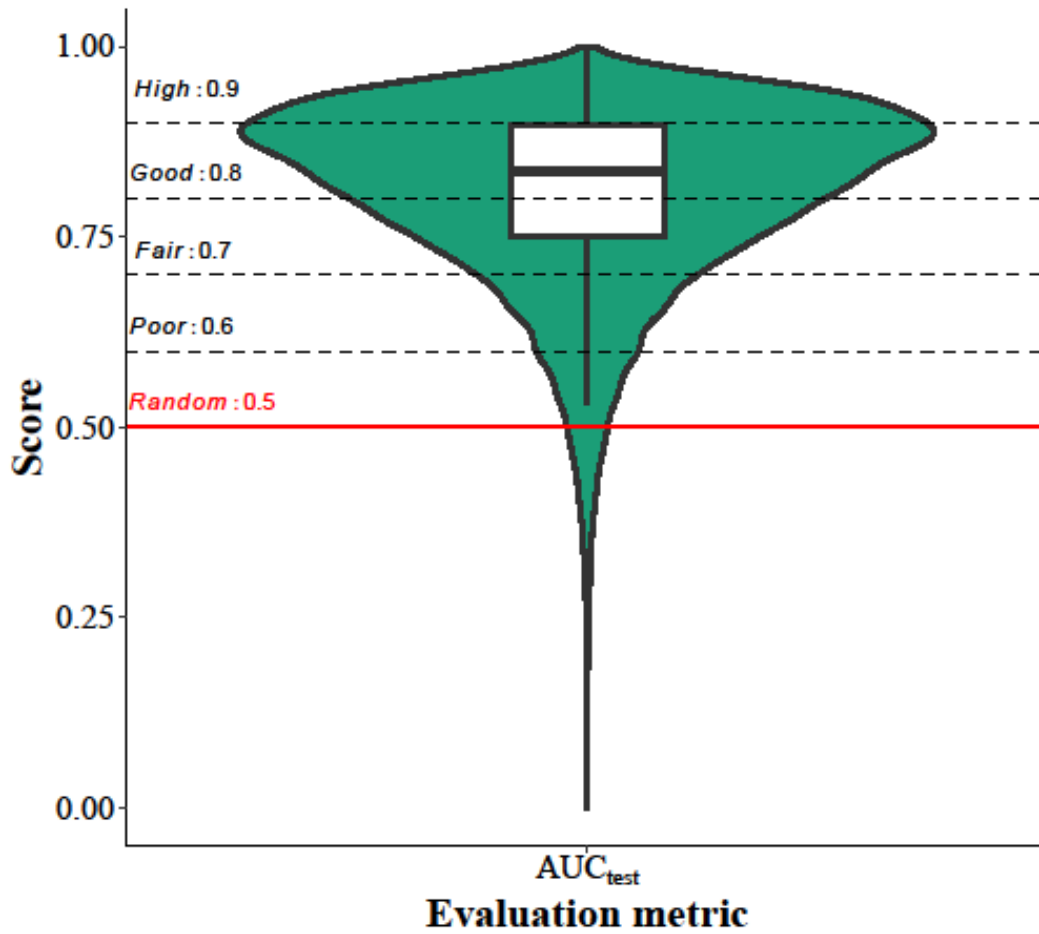
344

Figure S2. Number of occurrence records used for modelling the distribution of each utilised plant species.

345

Performance of Maxent model predictions

Based on testing records from different geographic blocks



Based on 28168 species

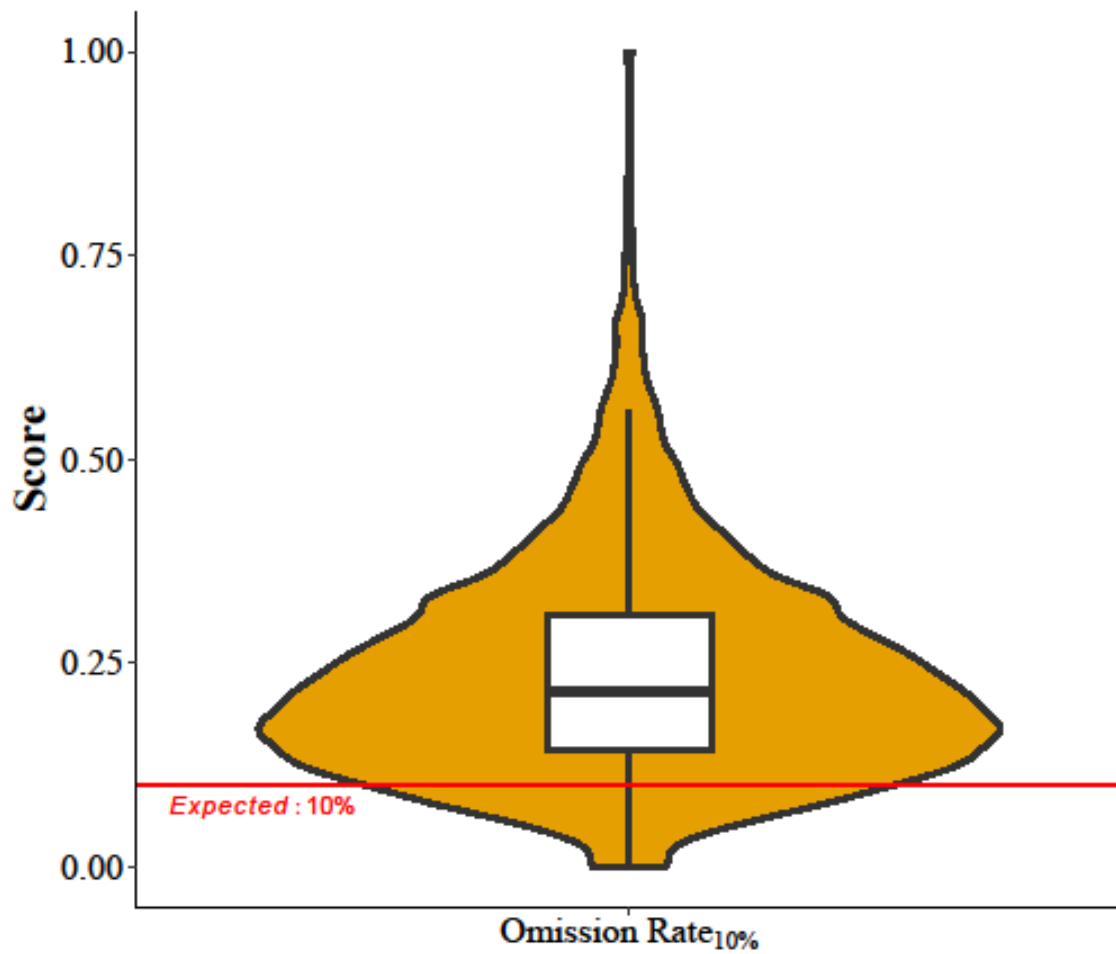
346

347 **Figure S3. AUC evaluation scores of utilised plant species distribution models.** Boxes represent upper and lower
348 extremes, upper and lower quartiles, medians and outliers of AUC indices for all utilised plant species distribution
349 models. Evaluation scores are given for MaxEnt models calibrated with more than ten occurrence records.

350

Performance of Maxent model predictions

Based on testing records from different geographic blocks



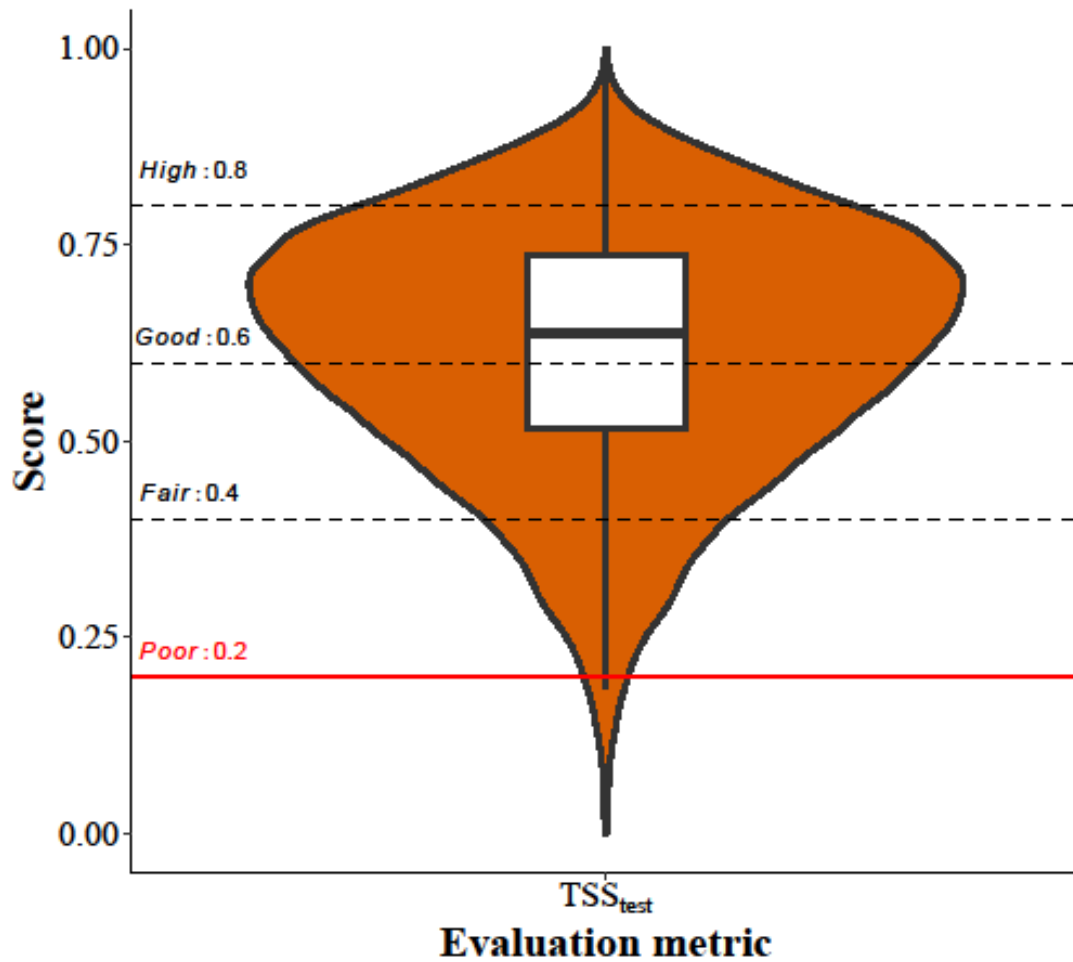
Based on 28168 species

351

352 **Figure S4. OR10 evaluation scores of utilised plant species distribution models.** Boxes represent upper and lower
353 extremes, upper and lower quartiles, medians and outliers of OR10 indices for all utilised plant species distribution
354 models. Evaluation scores are given for MaxEnt models calibrated with more than ten occurrence records.
355

Performance of Maxent model predictions

Based on testing records from different geographic blocks

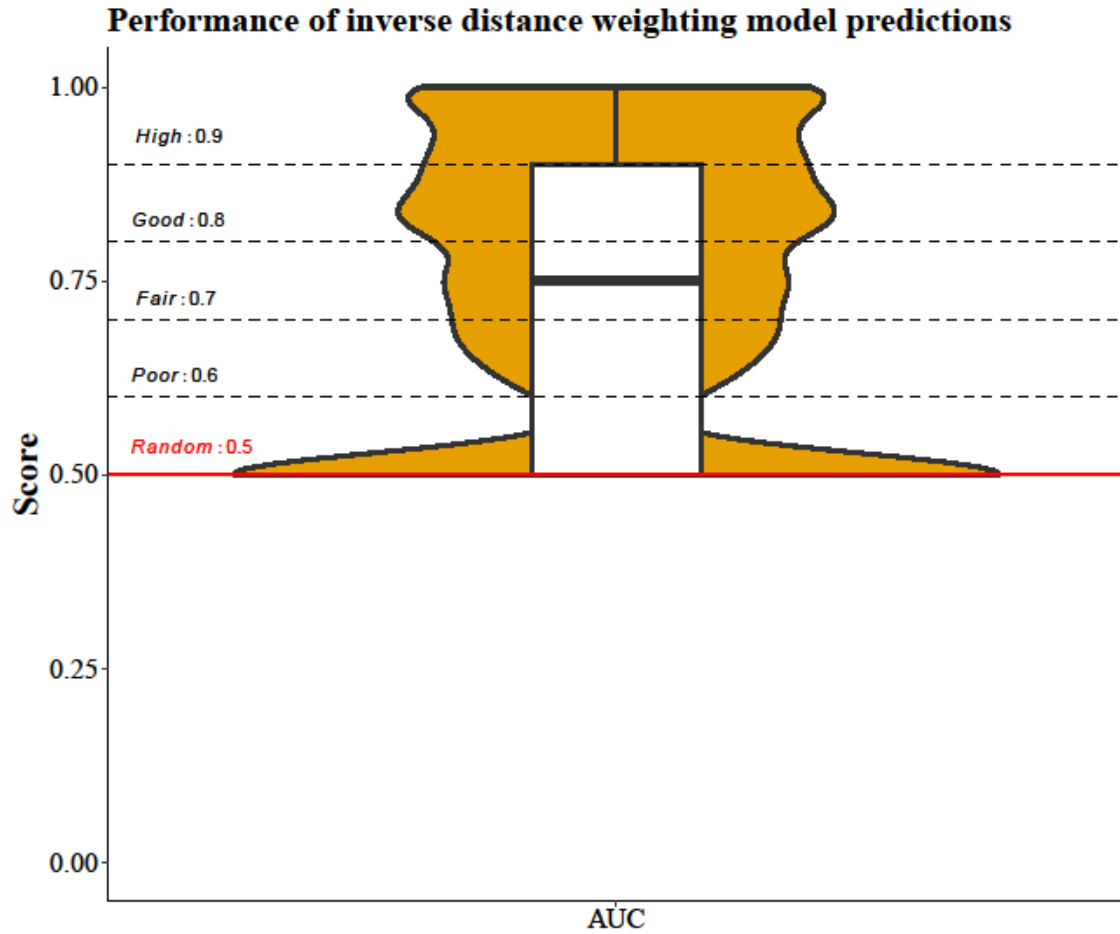


Based on 28168 species

356

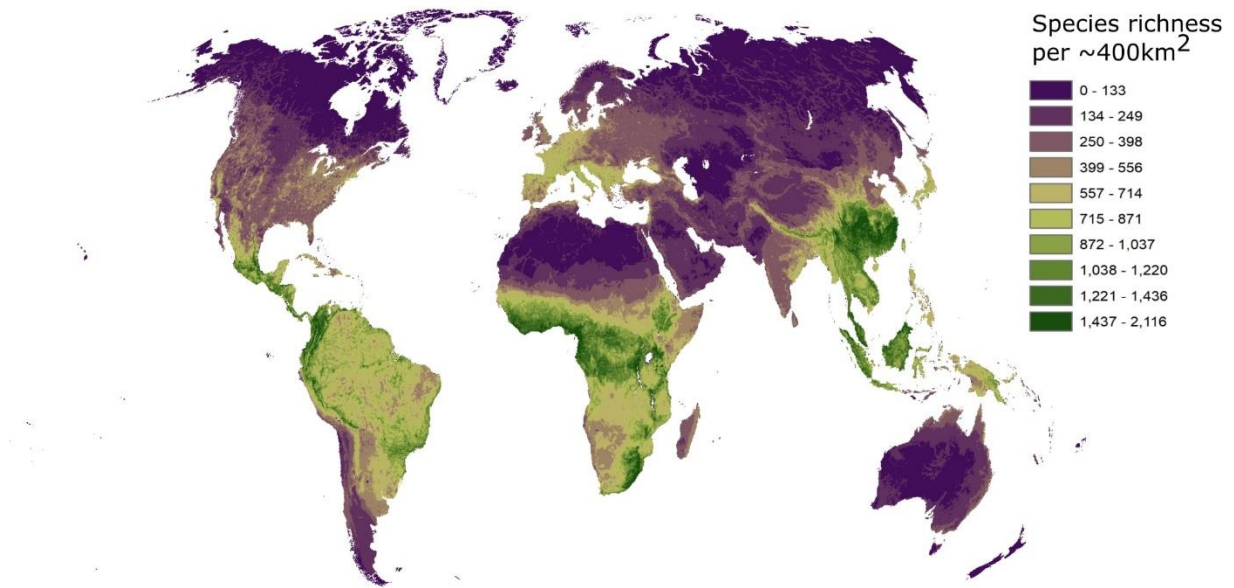
357 **Figure S5. TSS evaluation scores of utilised plant species distribution models.** Boxes represent upper and lower
358 extremes, upper and lower quartiles, medians and outliers of TSS indices for all utilised plant species distribution
359 models. Evaluation scores are given for MaxEnt models calibrated with more than ten occurrence records.

360



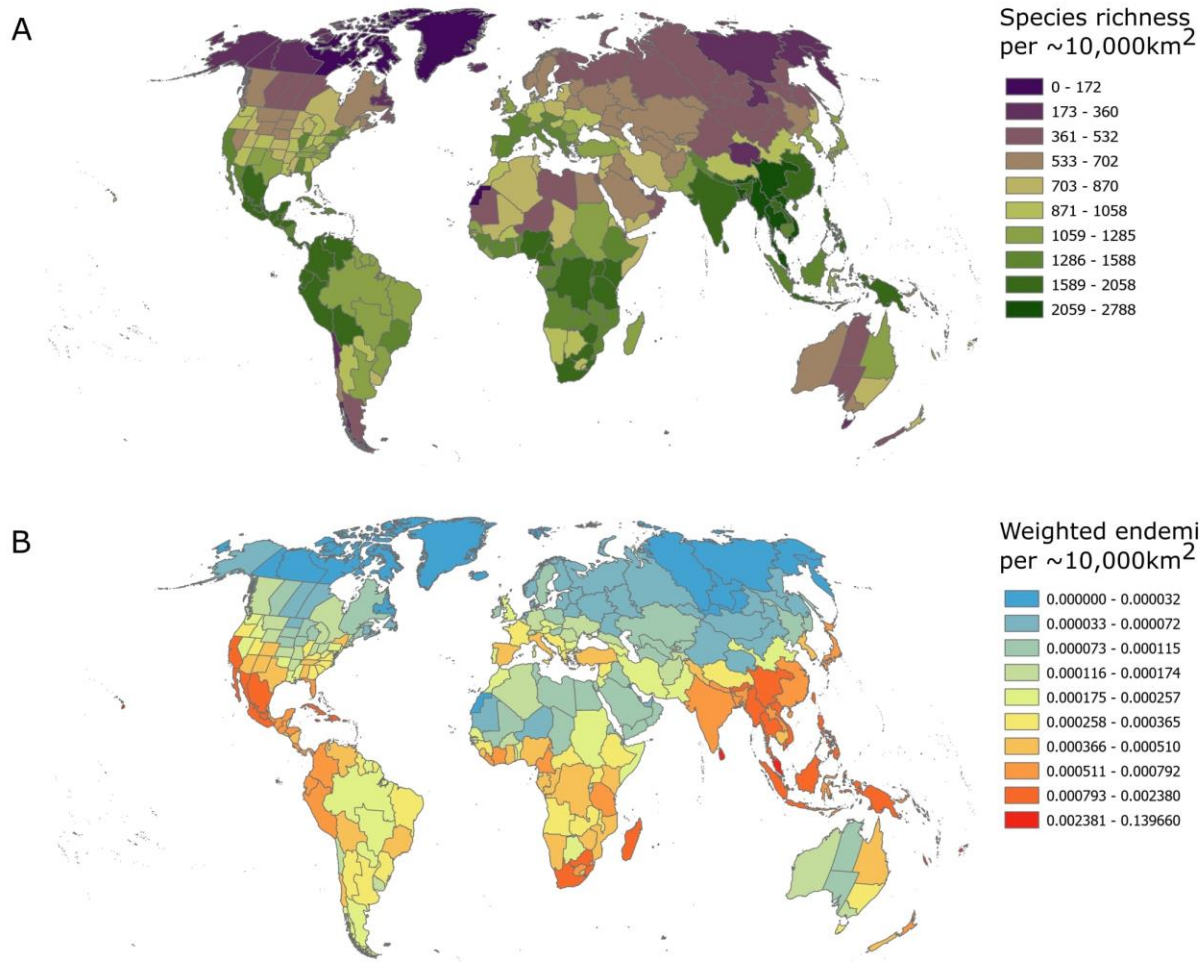
Based on 5439 species

361
 362 **Figure S6. AUC evaluation scores of utilised plant species distribution models.** Boxes represent upper and lower
 363 extremes, upper and lower quartiles, medians and outliers of AUC indices for all utilised plant species distribution
 364 models. Evaluation scores are given for geographic models (Inverse Distance Weighting) calibrated with less than ten
 365 occurrence records.
 366



367

368 **Figure S7. Global species richness of plants with known uses by humans.** Utilised plant species richness
369 corresponds to the sum of species occurrence probabilities predicted in each ten arc-minutes (~ 20 km) pixel found
370 across their native ranges.



371

372 **Figure S8. Global distribution of utilised plant species richness (A) and weighted endemism (B) at the level-3**
 373 **(sub-national) of the World Geographical Scheme for Recording Plant Distribution (WGSRPD).** Utilised plant
 374 species richness corresponds to the count of species found in each region. Utilised plant species endemism corresponds
 375 to the number of species present in each region weighted by the inverse of their range size calculated as the total area
 376 covered by the level-3 regions it occupies (i.e., weighted endemism). Geographic distribution for each species was
 377 retrieved from the World Checklist of Vascular Plants at the level-3 of the WGSRPD. To account for differences in
 378 surface area between regions, we estimated a scaling exponent that describes the species-area relationship (SAR)
 379 between the counts of species (weighted or not) and the level-3 region areas. We used this exponent to rescale species
 380 counts for a standard area of 10,000 square kilometers.

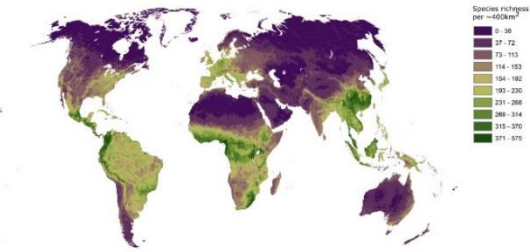
381



Animal Food



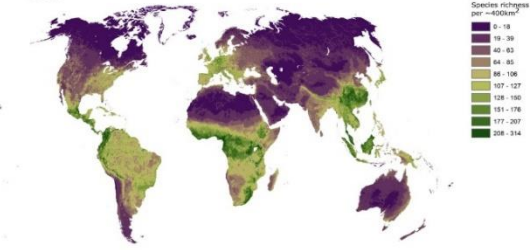
Environmental Uses



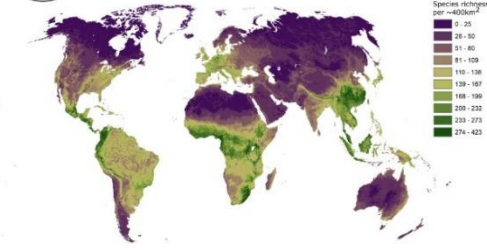
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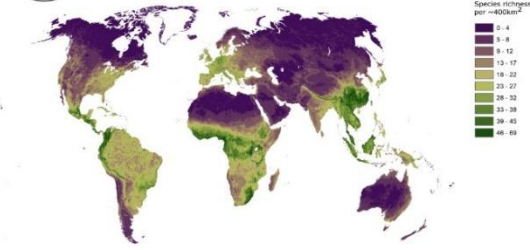
Gene Sources



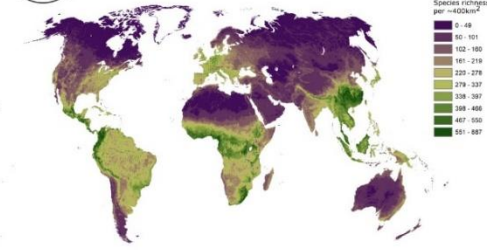
Human Food



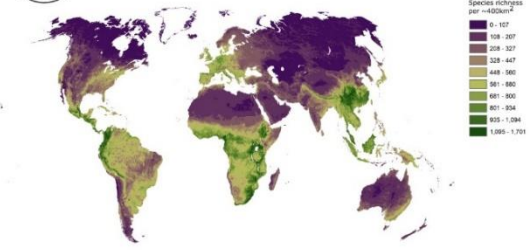
Invertebrate Food



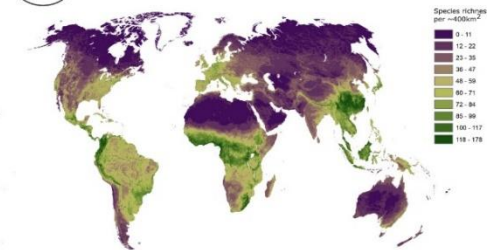
Materials



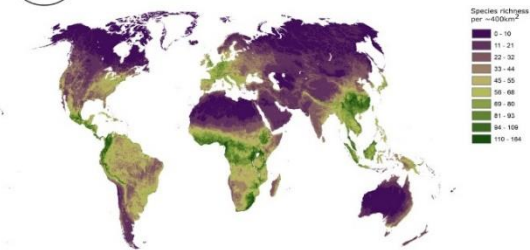
Medicines



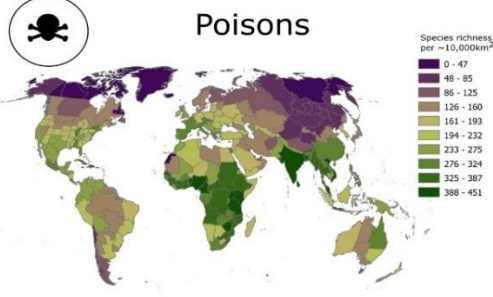
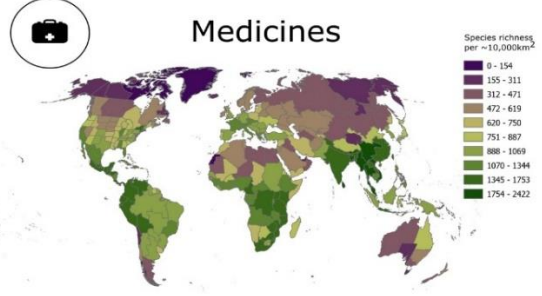
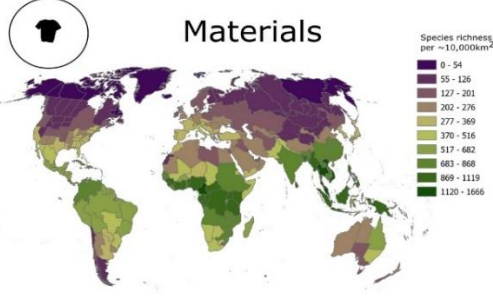
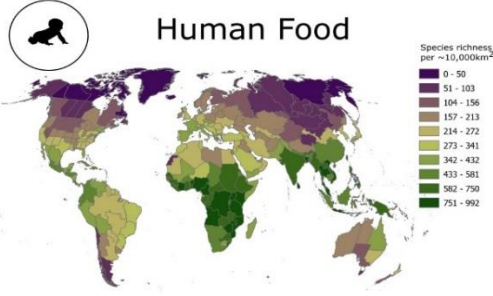
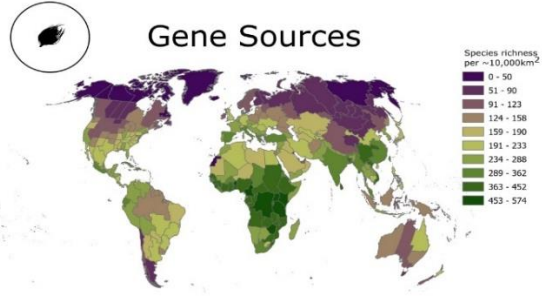
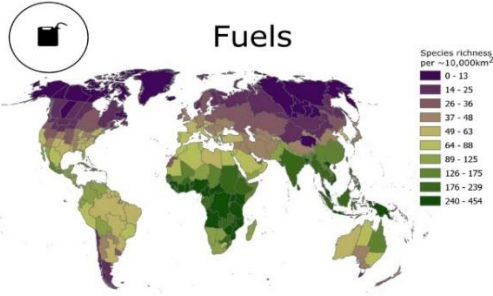
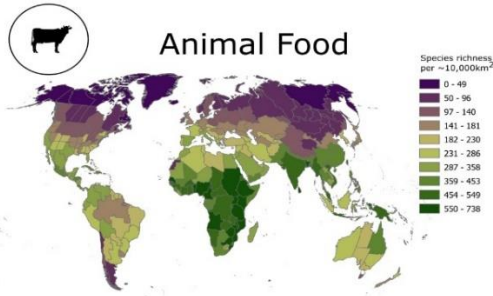
Poisons



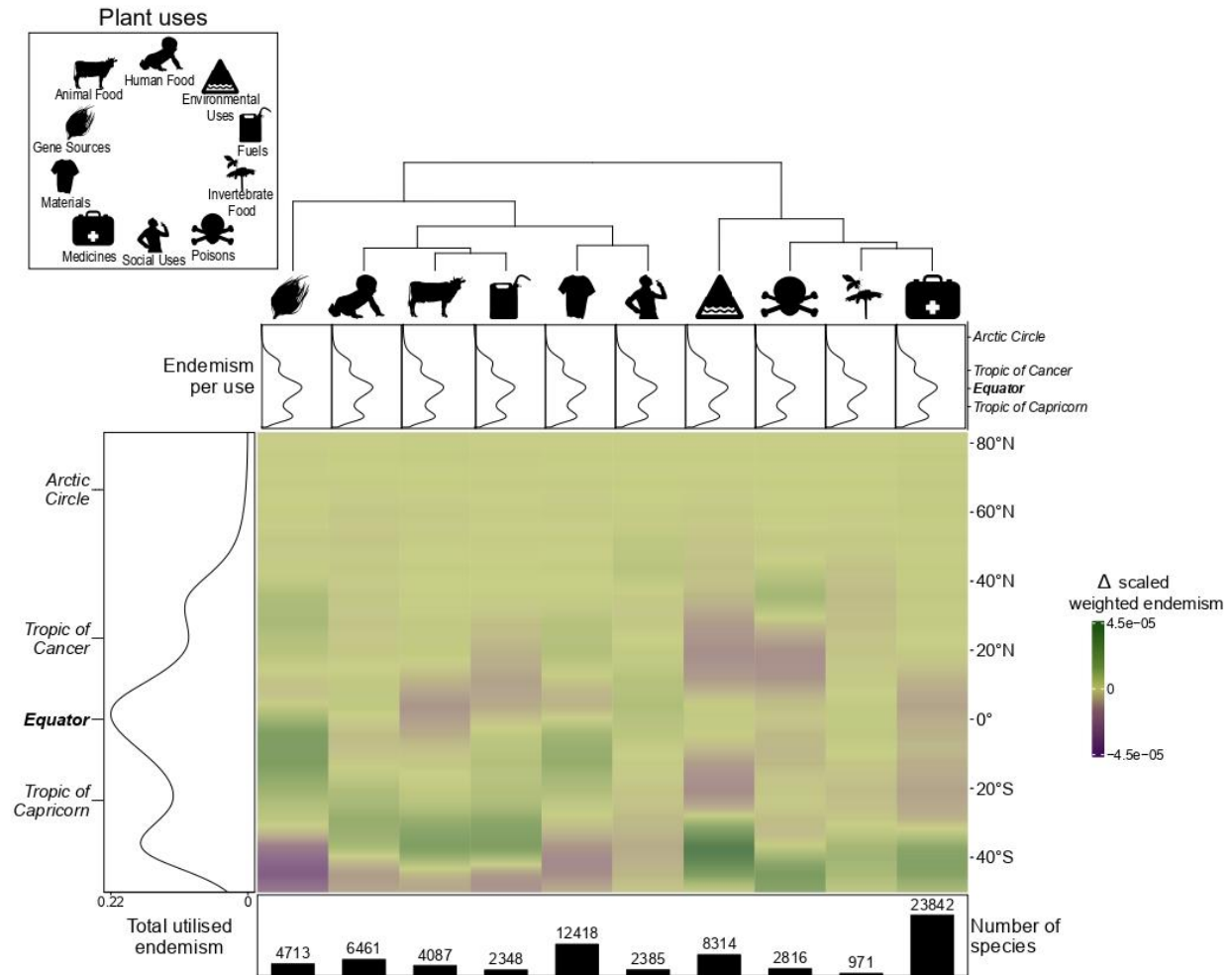
Social Uses



383 **Figure S9. Global distribution of utilised plant species richness across ten categories of uses.** Utilised plant
384 species richness corresponds to the sum of each species occurrence probabilities predicted in each pixel by our
385 “stacked-species distribution modelling” approach.



387 **Figure S10. Global distribution of utilised plant species richness at the level-3 (sub-national) of the World**
388 **Geographical Scheme for Recording Plant Distribution (WGSRPD) across ten categories of uses.** Utilised plant
389 species richness corresponds to the count of species found in each region. Geographic distribution for each species
390 was retrieved from the World Checklist of Vascular Plants at the level-3 of the WGSRPD. To account for differences
391 in surface area between regions, we estimated a scaling exponent that describes the species-area relationship (SAR)
392 between the counts of species and the level-3 region areas. We used this exponent to rescale species counts for a
393 standard area of 10,000 square kilometers.
394

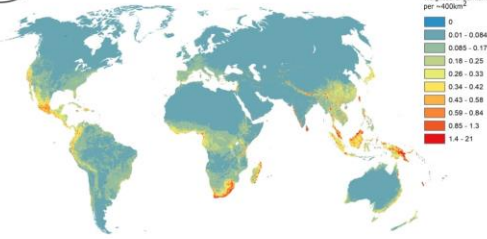


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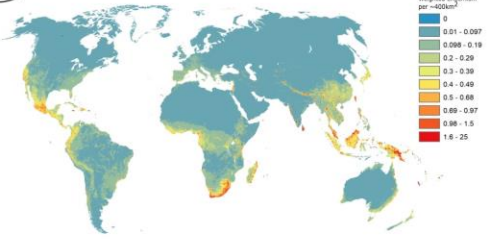
396 **Figure S11. Latitudinal distribution of utilised plant species endemism across ten categories of uses.** The black
 397 curve on the left represents the latitudinal distribution of all utilised plant species endemism. The dendrogram on the
 398 top orders the ten use categories according to the (dis-)similarity of their species endemism latitudinal profiles. Black
 399 curves underneath the dendrogram correspond to the species endemism latitudinal profile for each use category. The
 400 heatmap describes the latitudinal variation in the deviation of utilised plant species endemism for the ten plant use
 401 categories from total utilised plant species endemism. Colors indicate higher (green) or lower (purple) proportions in
 402 utilised plant species endemism of a given use relative to the total utilised plant species endemism pattern. The bar
 403 chart underneath the heatmap shows the number of species considered in each use category.
 404



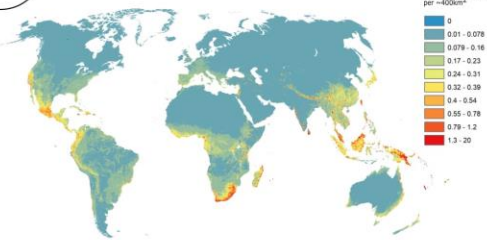
Animal Food



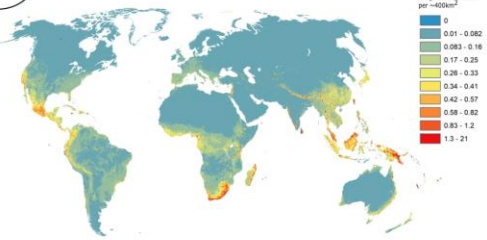
Environmental Uses



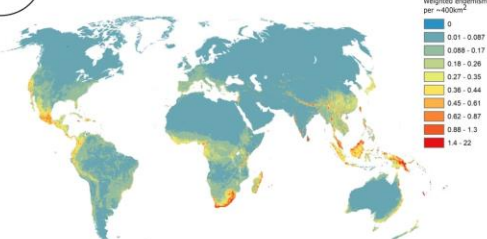
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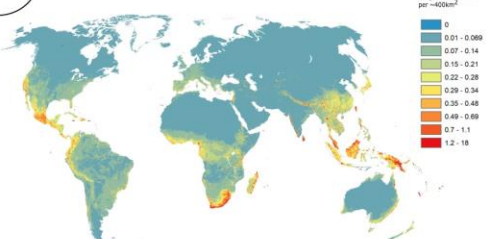
Gene Sources



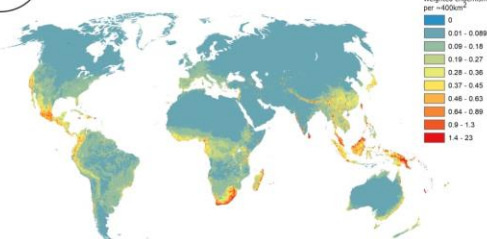
Human Food



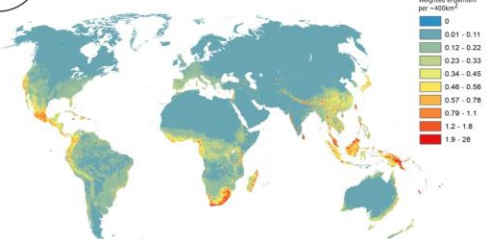
Invertebrate Food



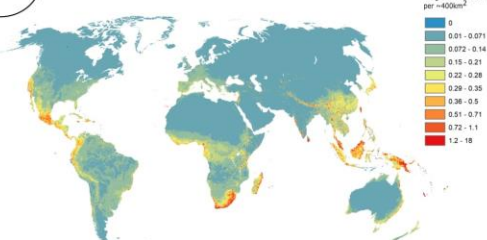
Materials



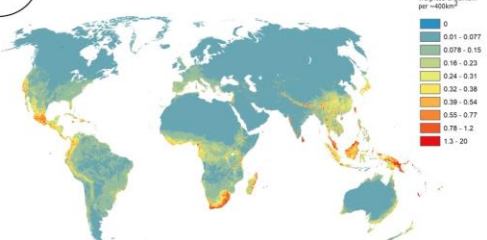
Medicines



Poisons



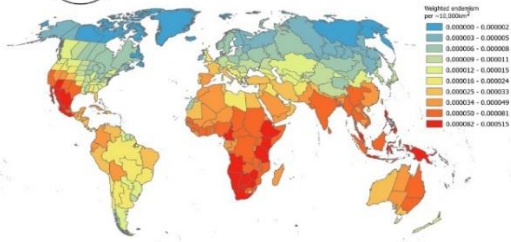
Social Uses



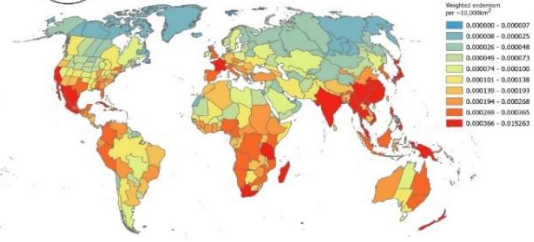
406 **Figure S12. Global distribution of utilised plant species endemism across ten categories of uses.** Utilised plant
407 species endemism corresponds to the sum of each species occurrence probabilities predicted in each pixel weighted
408 by the inverse of their range size calculated as the sum of the predicted probabilities within their study region (i.e.,
409 weighted endemism).
410



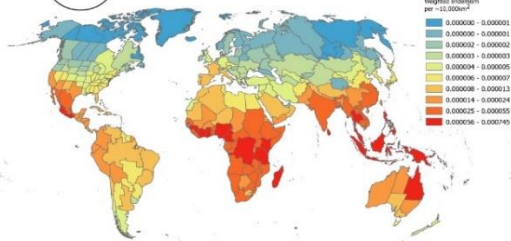
Animal Food



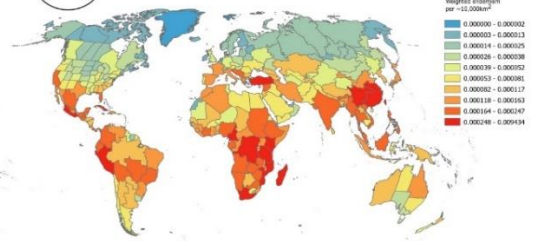
Environmental Uses



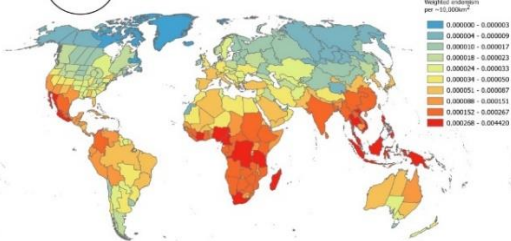
Fuels



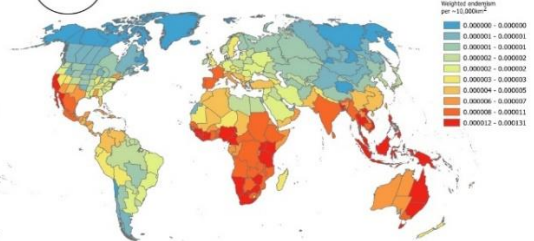
Gene Sources



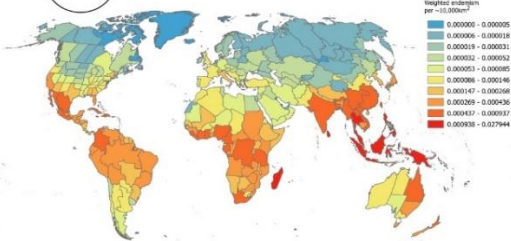
Human Food



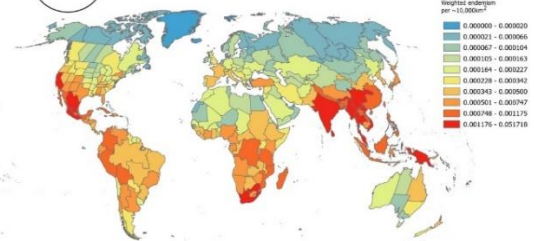
Invertebrate Food



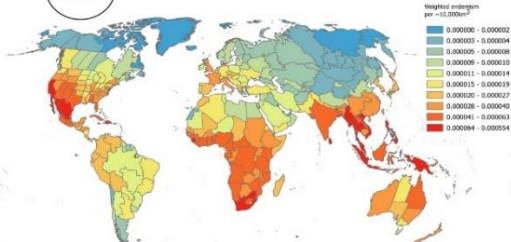
Materials



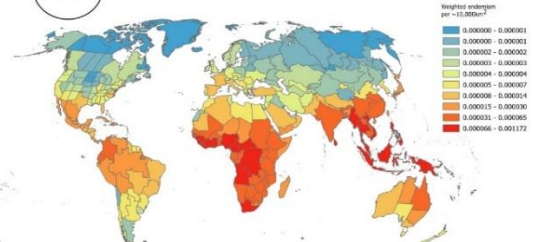
Medicines



Poisons



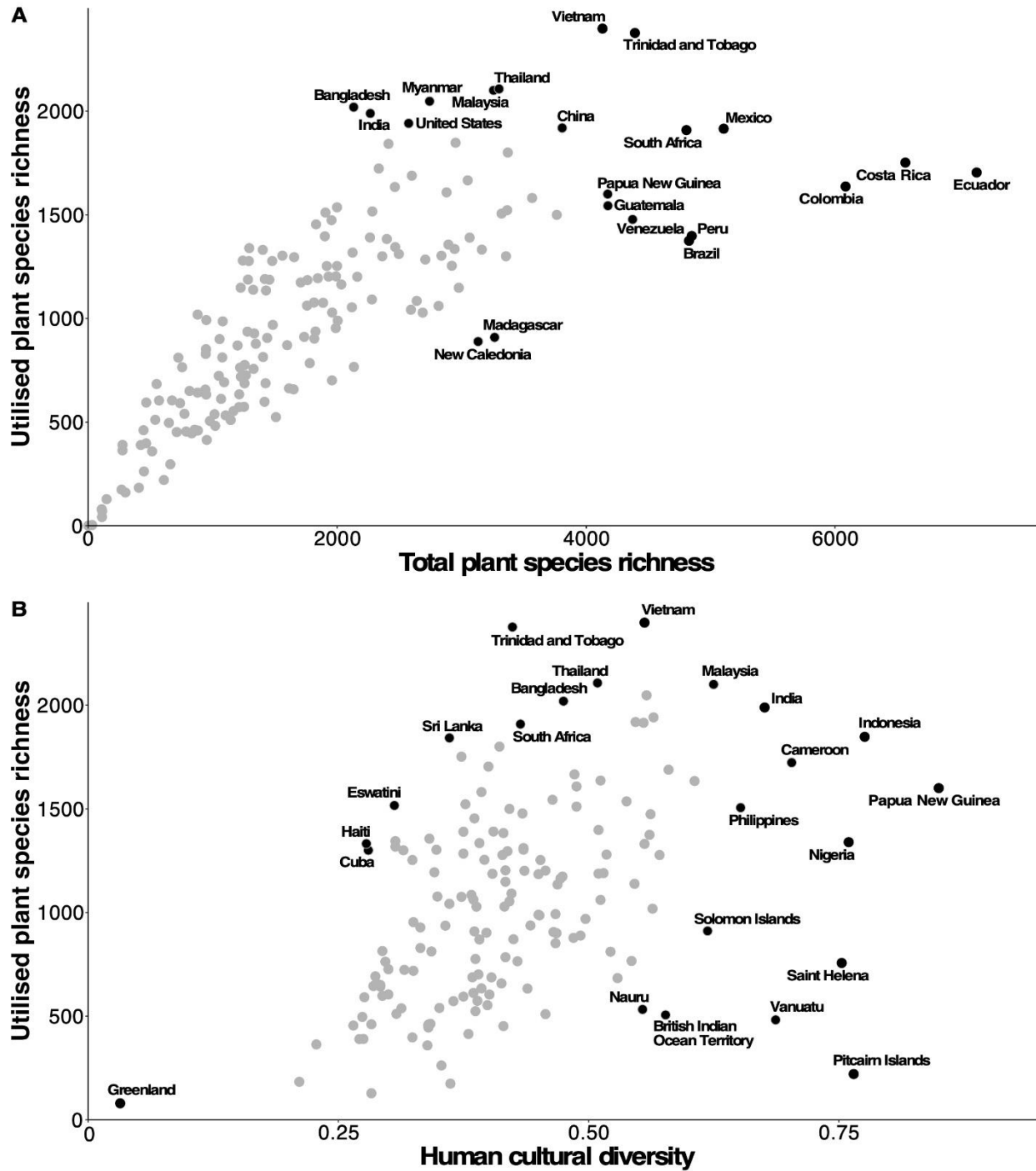
Social Uses



412 **Figure S13. Global distribution of utilised plant species endemism at the level-3 (sub-national) of the World**
413 **Geographical Scheme for Recording Plant Distribution (WGSRPD) across ten categories of uses.** Utilised plant
414 species endemism corresponds to the number of species present in each region weighted by the inverse of their range
415 size calculated as the total area covered by the level-3 regions it occupies (i.e., weighted endemism). Geographic
416 distribution for each species was retrieved from the World Checklist of Vascular Plants at the level-3 of the WGSRPD.
417 To account for differences in surface area between regions, we estimated a scaling exponent that describes the species-
418 area relationship (SAR) between the counts of species and the level-3 region areas. We used this exponent to rescale
419 species counts for a standard area of 10,000 square kilometers.

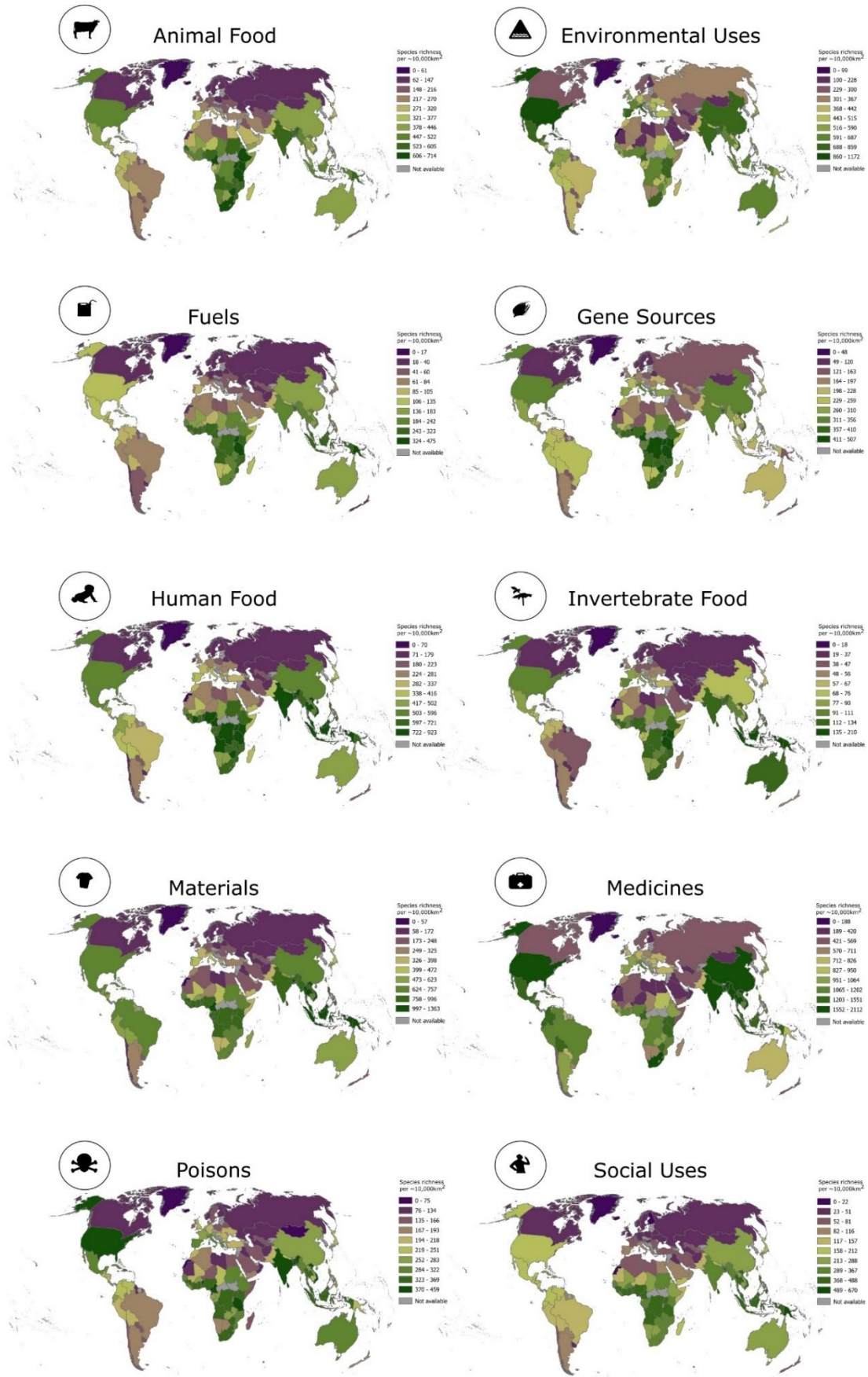
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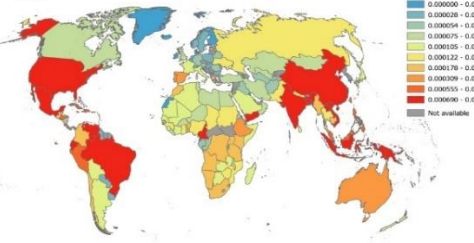
423 **Figure S14. Relationships between utilised plant species richness and total plant species richness (A), and**
 424 **human cultural diversity (B) across 163 countries/territories.** Geographic distribution for each species was derived
 425 from the World Checklist of Vascular Plants at the level-3 of the WGSRPD. To account for differences in surface area
 426 between regions, we estimated a scaling exponent that describes the species-area relationship (SAR) between the
 427 counts of species and the region areas. We used this exponent to rescale species counts for a standard area of 10,000
 428 square kilometers.



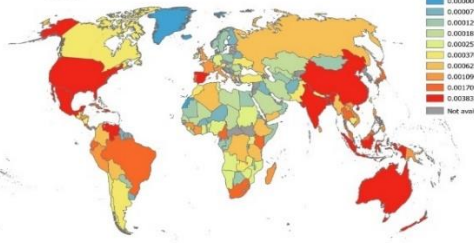
430 **Figure S15. Global distribution of utilised plant species richness at a country/territory scale across ten**
431 **categories of uses.** Utilised plant species richness corresponds to the count of species found in each region.
432 Geographic distribution for each species was derived from the World Checklist of Vascular Plants at the level-3 of
433 the WGSRPD. To account for differences in surface area between regions, we estimated a scaling exponent that
434 describes the species-area relationship (SAR) between the counts of species and the region areas. We used this
435 exponent to rescale species counts for a standard area of 10,000 square kilometers.
436



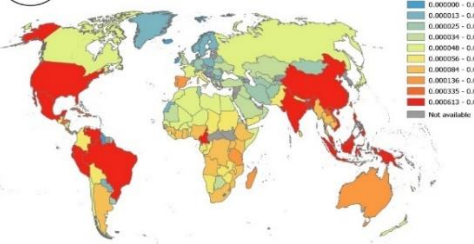
Animal Food



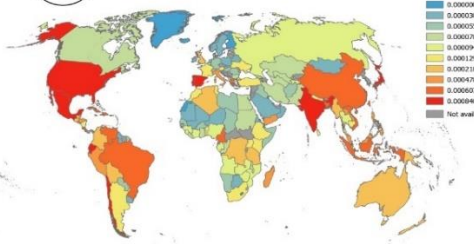
Environmental Uses



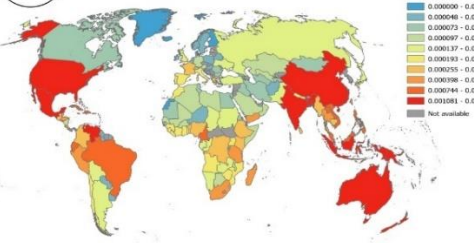
Fuels



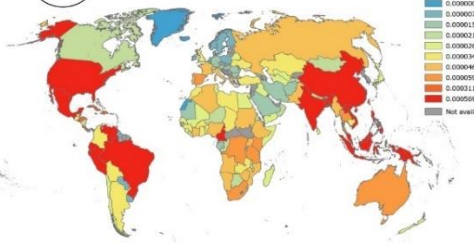
Gene Sources



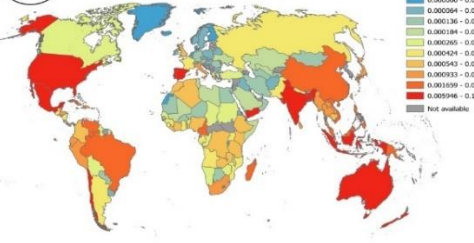
Human Food



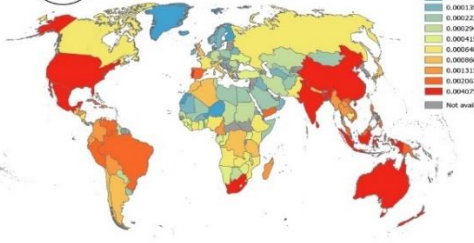
Invertebrate Food



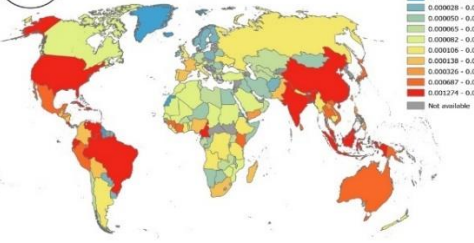
Materials



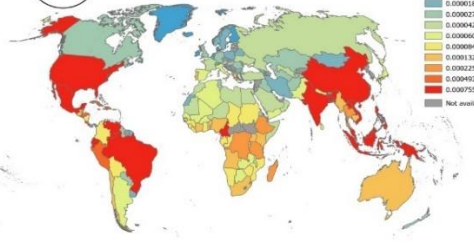
Medicines



Poisons

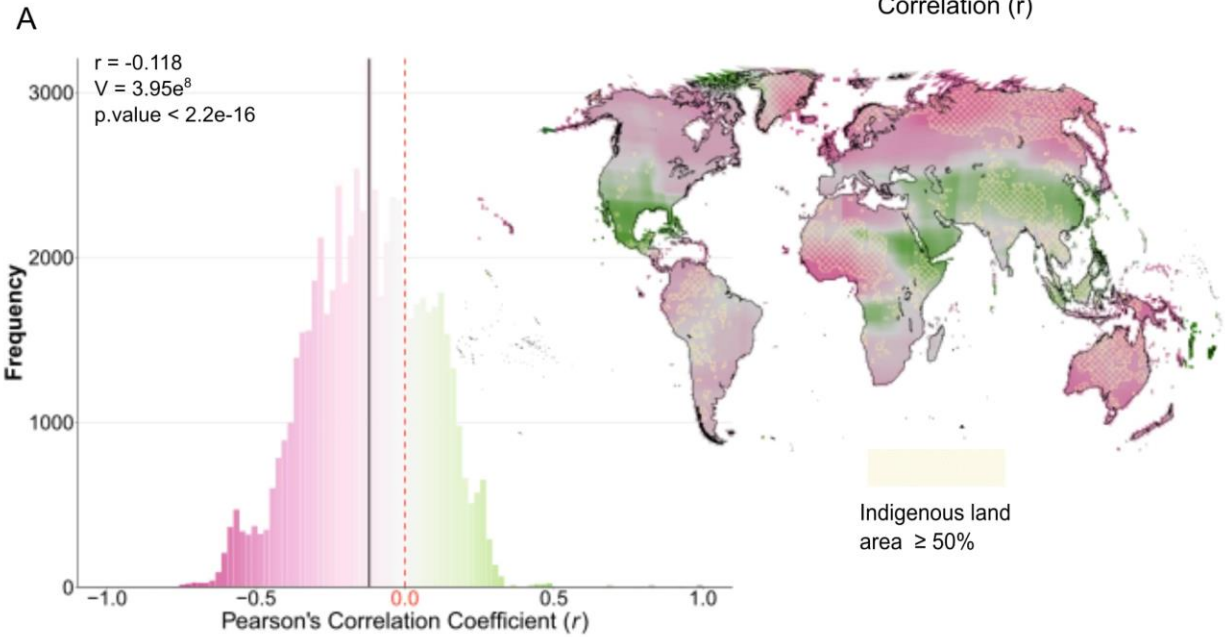


Social Uses

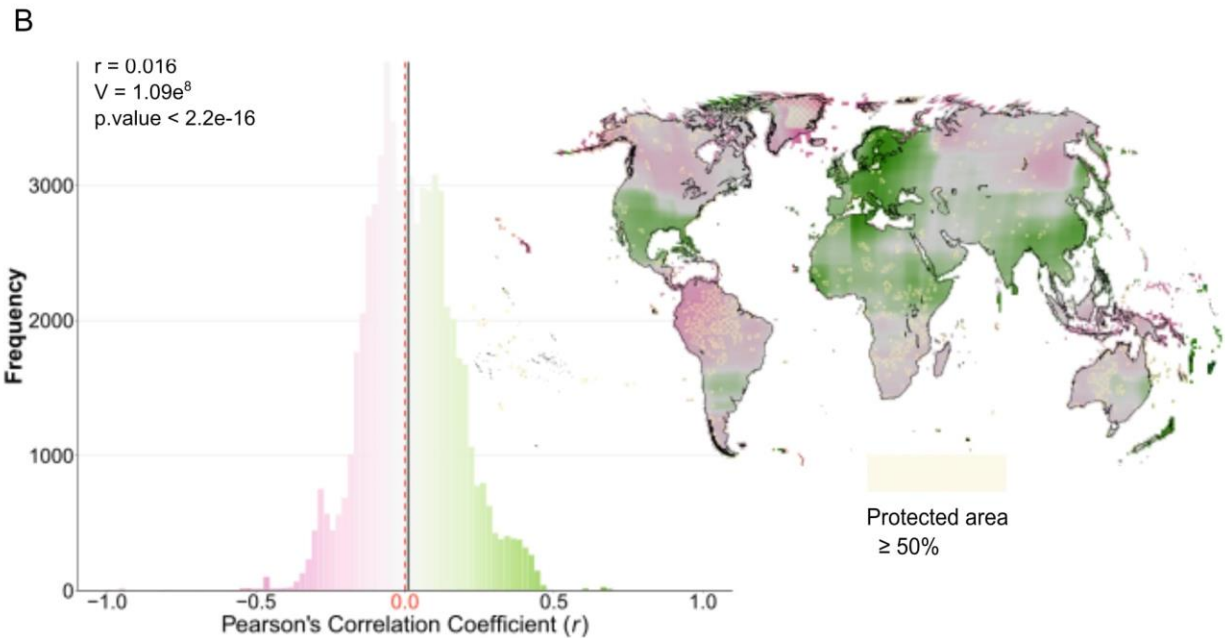


438 **Figure S16. Global distribution of utilised plant species endemism at a country/territory scale across ten**
439 **categories of uses.** Utilised plant species endemism corresponds to the number of species present in each region
440 weighted by the inverse of their range size calculated as the total area covered by the regions it occupies (i.e., weighted
441 endemism). Geographic distribution for each species was retrieved from the World Checklist of Vascular Plants at the
442 level-3 of the WGSRPD. To account for differences in surface area between regions, we estimated a scaling exponent
443 that describes the species-area relationship (SAR) between the counts of species and the region areas. We used this
444 exponent to rescale species counts for a standard area of 10,000 square kilometers.
445

Indigenous land areas



Terrestrial protected areas



448 **Figure S17. Spatial correlations between utilised plant species endemism and proportions of both Indigenous**
 449 **lands and terrestrial protected areas.** Pearson's correlation coefficients were computed across all values contained
 450 in 71 cells (~3,550km)-wide windows built around each pixel. Pixel color indicates regions where utilised plant
 451 species endemism is positively (green) or negatively (pink) correlated with the proportion of Indigenous lands and
 452 terrestrial protected areas. Regions crossed in beige indicate pixels containing more than 50% of Indigenous lands and

453 protected areas. All Indigenous lands and protected areas are thus not represented on the maps, although they are all
454 accounted for in the analyses. Frequencies of Pearson's correlation coefficients found across the world are given in
455 histograms. The median correlation across the world is indicated by the black vertical line, while zero correlation is
456 indicated by the red dashed line. One-sample Wilcoxon signed rank tests were performed to assess whether median
457 correlations are significantly different from zero.
458

Dataset	Source	Acronym	Number of species	Geographic extent	Description
Crop wild relative inventory	https://www.cwi-diversity.org Dempewolf <i>et al.</i> 2014	CWR	2,459	Global	A global priority Crop Wild Relatives inventory, based on both gene pool and taxon group concepts
Economic Botany Collection	http://apps.kew.org/ecobot/search	EBot	8,162	Global	RBG Kew's Economic Botany Collection containing over 120,000 specimens objects
Germplasm Resources Information Network, United States Department of Agriculture	https://www.ars-gm.gov	GRIN	10,988	Global	Online database of taxonomic information on cultivated plants from the USDA-ARS germplasm resources information network (GRIN)
Medicinal Plant Names Services version 8.2	http://www.kew.org/mpns	MPNS	26,690	Global	Global resource for medicinal plant names with access to information about plants and plant products
Palms of the world online	http://www.palmweb.org	PalmWeb	152	Global	An online palm encyclopaedia gleaned from taxonomic publications
Plant resources of South-East Asia	Jansen <i>et al.</i> 1991	PROSEA	4,182	South-East Asia	Book inventorying useful plants of a core area (Brunei, Indonesia, Malaysia, New Guinea, the Philippines and Singapore) occurring and used also in neighbouring areas (Burma, Cambodia, Laos, Thailand and Vietnam). Based on literature review.
Indigenous knowledge of New Guinea's useful plants	Camara-Leret <i>et al.</i> 2019	NewGuinea	3,071	New Guinea	Quantitative review of 488 references reporting indigenous knowledge and plant uses in New Guinea
Plant Resources of Tropical Africa	https://www.prota4n.org/database	PROTA	4,534	Sub-Saharan Africa	Online resource of useful plant information from Sub-Saharan Africa
Useful Plants of West Tropical Africa	Barkhill 1994	UPWTA	3,349	Western Africa	Database of useful plants from West tropical Africa
Survey of Economic Plants for Arid and Semi-Arid Lands	http://apps.kew.org/sepasalweb/sepasab	SEPA:SAL	3,479	Arid and semi-arid African lands	Database of 16,407 uses records for species from the African Arid and Semi-Arid areas
Plants for Malaria, plants for fever	Miliken 1997	MALARIA	748	Latin America	Literature review of plant species used to treat malaria and fever
Project MGU-Useful Plants Project	Utian <i>et al.</i> 2017	UPP	905	Sub-Saharan Africa and Mesoamerica	Database compiling plant uses by local communities and partners of the UPP project from Botswana, South Africa, Mali, Kenya and Mexico

Table S1. Data sources of the World Checklist of Useful Plant Species.

Variable	Response	Standardised regression coefficient + 95% Confidence Interval			Student's t-test		Correlation ⁽²⁾
		Estimate	Lower CI	Upper CI	t-value	p-value ⁽¹⁾	
Total plant species richness							
	All uses	0.865	0.782	0.947	20.703	1.86e-48***	exponential
	Animal Food	0.676	0.573	0.779	12.966	3.24e-27***	exponential
	Environmental Uses	0.874	0.765	0.982	15.890	1.66e-35***	exponential
	Fuels	0.494	0.392	0.597	9.518	1.89e-17***	exponential
	Gene Sources	0.747	0.631	0.863	12.705	2.38e-26***	exponential
	Human Food	0.692	0.589	0.796	13.184	1.04e-27***	exponential
	Invertebrate Food	0.684	0.567	0.801	11.554	3.95e-23***	exponential
	Materials	0.662	0.570	0.754	14.256	9.81e-31***	exponential
	Medicines	0.822	0.737	0.906	19.232	1.29e-44***	spherical
	Poisons	0.788	0.668	0.908	12.995	3.60e-27***	exponential
	Social Uses	0.522	0.431	0.613	11.305	1.99e-22***	exponential
Human cultural diversity							
	All uses	0.358	0.235	0.481	5.743	4.52e-08***	gaussian
	Animal Food	0.306	0.194	0.419	5.394	2.42e-07***	spherical
	Environmental Uses	0.349	0.217	0.481	5.217	5.53e-07***	spherical
	Fuels	0.377	0.275	0.479	7.322	1.10e-11***	spherical
	Gene Sources	0.314	0.183	0.445	4.737	4.74e-06***	exponential
	Human Food	0.355	0.239	0.471	6.043	1.01e-08***	exponential
	Invertebrate Food	0.392	0.268	0.517	6.219	4.14e-09***	exponential
	Materials	0.352	0.244	0.460	6.451	1.24e-09***	gaussian
	Medicines	0.328	0.204	0.451	5.242	4.93e-07***	gaussian
	Poisons	0.342	0.208	0.476	5.030	1.29e-06***	exponential
	Social Uses	0.323	0.223	0.424	6.359	2.01e-09***	exponential

⁽¹⁾Signif. codes: 0 '****' 0.001 '***' 0.01 '**' 0.05 '.' 0.1 ' ' 1

⁽²⁾Name of the correlation structure function selected for accommodating spatial autocorrelation

461
462 **Table S2. Generalised Least Squares (GLS) summary statistics for the relationships between utilised plant**
463 **species richness and total plant species richness, and human cultural diversity across 163 countries/territories.**
464 GLS models account for potential spatial (auto-)correlation in the data. Latitudinal and longitudinal coordinates of the
465 centroids of each country were used as spatial covariates, and three different correlation functions were tested:
466 exponential, gaussian and spherical. GLS models were fitted using the restricted maximum likelihood method (REML)
467 and the correlation structure with the lowest AIC was selected. Analyses were repeated for all utilised plant species
468 together and for each individual use category separately. Human cultural diversity is a composite index made of three
469 estimates: total number of languages, religions, and ethnic groups per country. Total vascular plant species richness
470 and utilised plant species richness were compiled from the World Checklist of Vascular Plants.

471

Variable	Response	Standardised regression coefficient + 95% Confidence Interval			Student's t-test		Correlation ⁽²⁾
		Estimate	Lower CI	Upper CI	t-value	p-value ⁽¹⁾	
Total plant species endemism							
	All uses	0.297	0.213	0.381	7.001	5.54e-11***	gaussian
	Animal Food	0.016	-0.118	0.149	0.231	8.18e-01	gaussian
	Environmental Uses	0.065	0.007	0.123	2.209	2.85e-02*	gaussian
	Fuels	-0.046	-0.174	0.082	-0.713	4.77e-01	gaussian
	Gene Sources	0.435	0.371	0.498	13.456	1.79e-28***	gaussian
	Human Food	0.146	0.030	0.263	2.477	1.42e-02*	spherical
	Invertebrate Food	-0.066	-0.196	0.064	-1.003	3.17e-01	gaussian
	Materials	0.184	0.111	0.256	5.009	1.38e-06***	gaussian
	Medicines	0.353	0.218	0.488	5.153	7.00e-07***	spherical
	Poisons	0.087	-0.048	0.222	1.279	2.03e-01	gaussian
	Social Uses	-0.009	-0.134	0.116	-0.146	8.84e-01	gaussian
Human cultural diversity							
	All uses	0.060	-0.037	0.157	1.218	2.25e-01	gaussian
	Animal Food	0.312	0.181	0.443	4.696	5.65e-06***	gaussian
	Environmental Uses	0.008	-0.050	0.066	0.269	7.88e-01	gaussian
	Fuels	0.340	0.217	0.462	5.453	1.83e-07***	gaussian
	Gene Sources	0.002	-0.104	0.108	0.038	9.70e-01	gaussian
	Human Food	-0.028	-0.171	0.116	-0.381	7.04e-01	spherical
	Invertebrate Food	0.336	0.209	0.463	5.223	5.37e-07***	gaussian
	Materials	0.013	-0.065	0.090	0.328	7.43e-01	gaussian
	Medicines	0.055	-0.108	0.218	0.670	5.04e-01	exponential
	Poisons	0.245	0.109	0.380	3.568	4.74e-04***	gaussian
	Social Uses	0.327	0.205	0.448	5.317	3.47e-07***	gaussian

⁽¹⁾Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

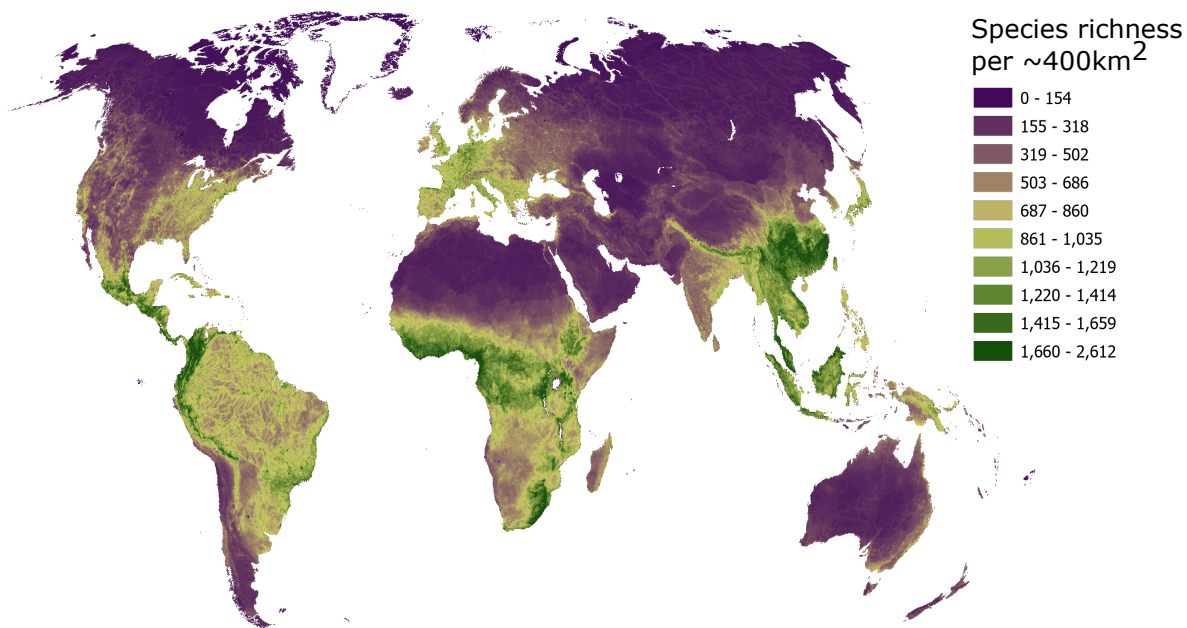
⁽²⁾Name of the correlation structure function selected for accommodating spatial autocorrelation

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Table S3. Generalised Least Squares (GLS) summary statistics for the relationships between utilised plant species endemism and total plant species endemism, and human cultural diversity across 163 countries/territories. GLS models account for potential spatial (auto-)correlation in the data. Latitudinal and longitudinal coordinates of the centroids of each country were used as spatial covariates, and three different correlation functions were tested: exponential, gaussian and spherical. GLS models were fitted using the restricted maximum likelihood method (REML) and the correlation structure with the lowest AIC was selected. Analyses were repeated for all utilised plant species together and for each individual use category separately. Human cultural diversity is a composite index made of three estimates: total number of languages, religions, and ethnic groups per country. Total vascular plant species endemism and utilised plant species endemism were compiled from the World Checklist of Vascular Plants.

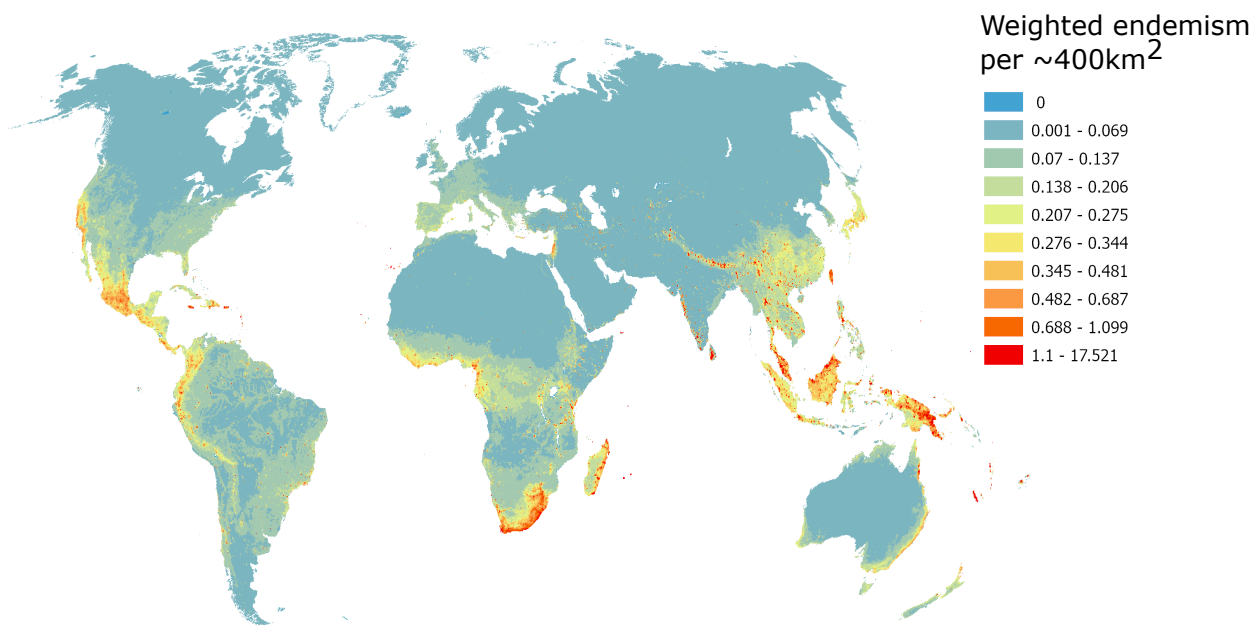
Global distribution of utilised plant species richness

A

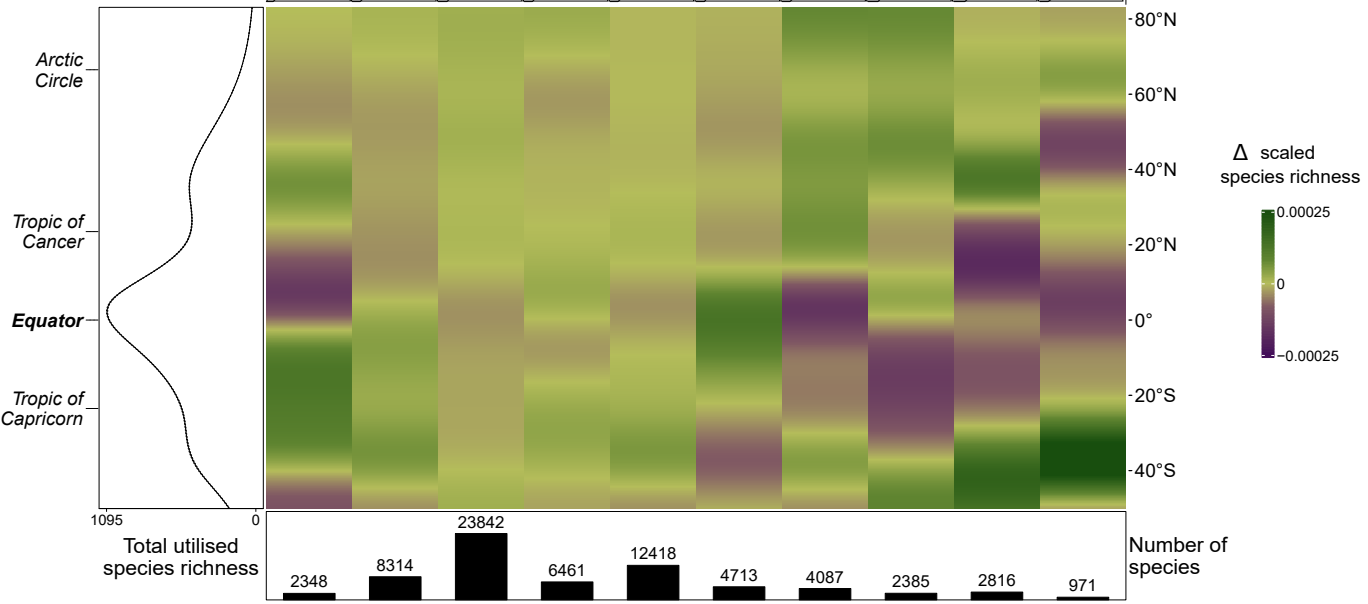
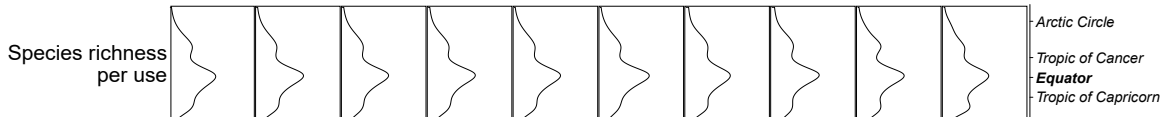
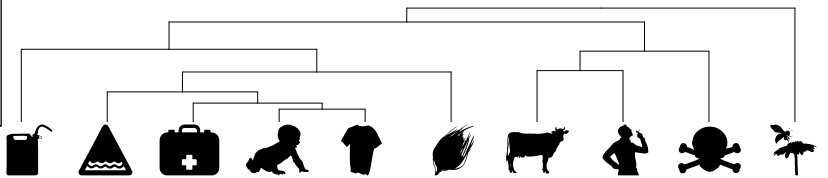
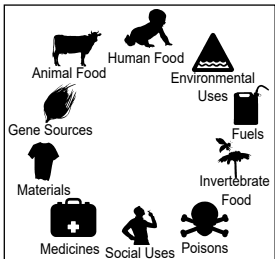


Global distribution of utilised plant species endemism

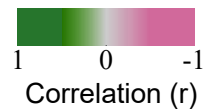
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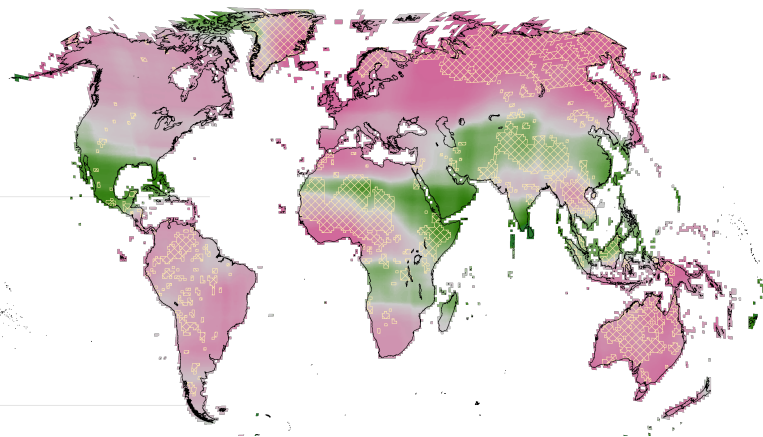
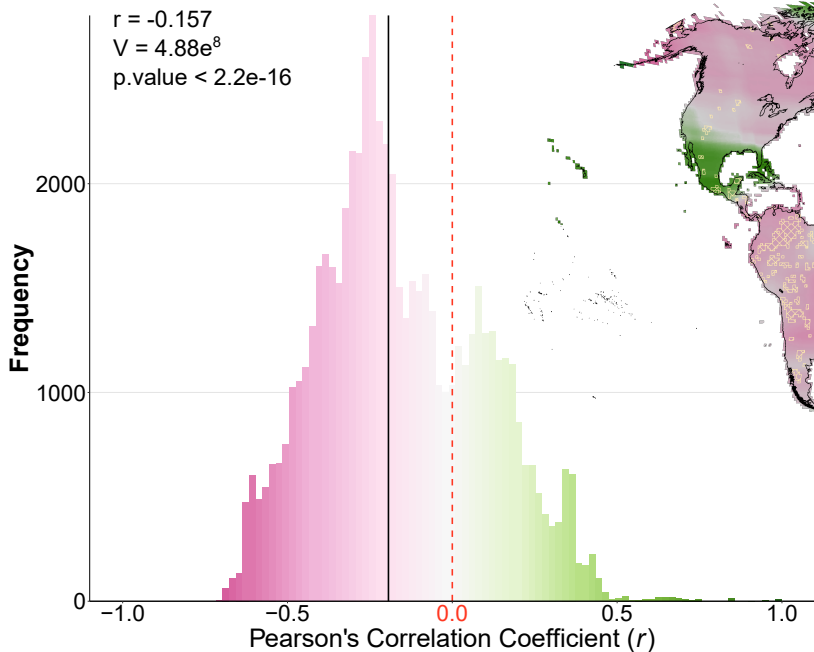
Plant uses



Indigenous land areas



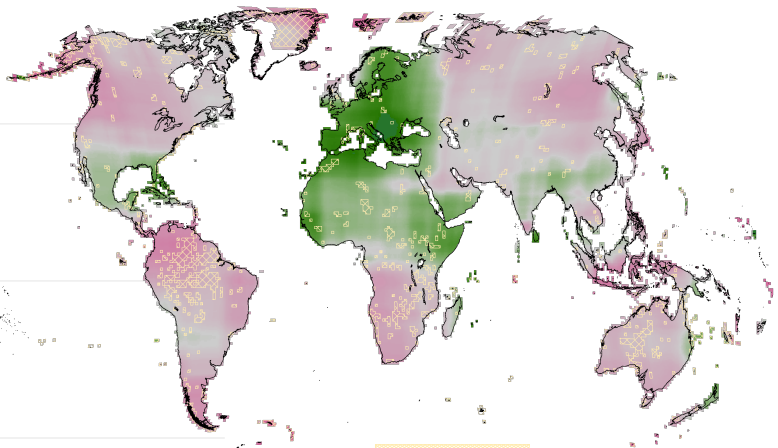
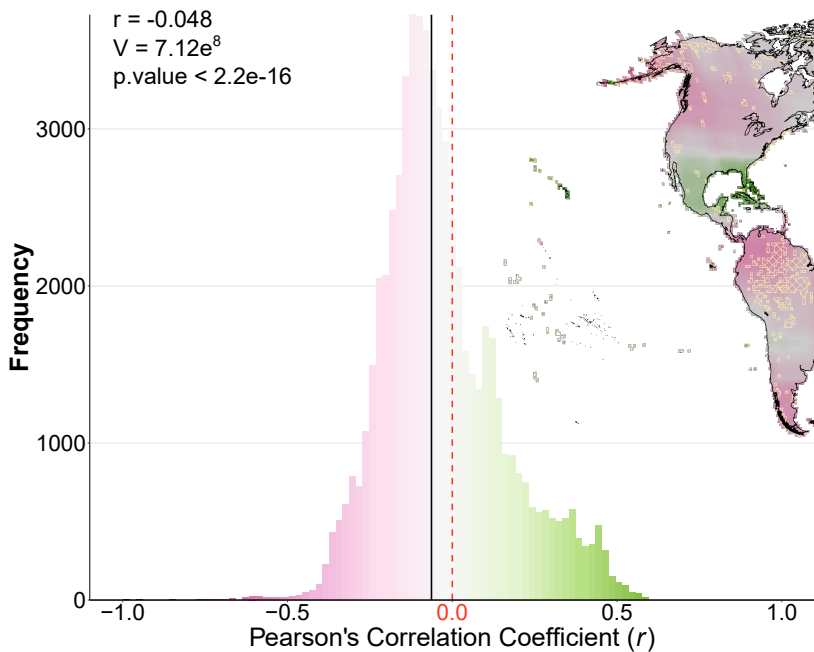
A



Indigenous land
area $\geq 50\%$

Terrestrial protected areas

B



Protected area
 $\geq 50\%$