

The impacts of climate change on aircraft noise near European airports

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Article

The Impacts of Climate Change on Aircraft Noise near European Airports

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Abstract

The warmer air resulting from climate change reduces the lift force on a departing aircraft, potentially reducing its climb angle and causing more engine noise near the airport. Here, we study this phenomenon at a selection of 30 European airports in northern hemisphere summer (June–July–August). We first formulate and verify a low-complexity model of noise propagation around airports, although we emphasise that our high-level results do not explicitly depend on this agreement. The model includes anisotropic noise propagation, atmospheric absorption, and the ability to model the noise emissions from multiple engines. We study the Airbus A320, but the method could be straightforwardly generalised to other aircraft. We refer to the model as an emulator since (using Latin hypercube parameter sampling) it mimics a more comprehensive model against which it is verified. The model is used to calculate the area enclosed by the 50 dB SPL (sound pressure level) contour, A_{50} , which agrees well with a similar metric (using the day–evening–night sound level, L_{den}) from the verification target, A . Using temperature and pressure data from IPCC simulations of future climate, and using a straightforward relation between climb angle and air density, we assess how climate change could affect climb angles by mid-century (2035–2064). The value of A_{50} is obtained by efficiently covarying (1) the engine noise at 10 m from the engines and (2) the climb angle under ‘historical’ conditions (1985–2014). The median values (across 10 climate models) of climb angle reduction in the future warmer climate are around 1–3% (depending on the airport and climate model used), but individual days can show values as high as 7.5% for the most extreme warming scenarios. By considering the variation in the absorption coefficient of the air with frequency, we find that the number of people affected by noise pollution could increase by up to 4%—as much as 2500 people for the most highly populated areas—by mid-century and that these changes are maximised for the most damaging and psychologically ‘annoying’ (low) frequencies.



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Keywords: climate change; noise pollution; Europe; scenario

1. Introduction

In this work, we examine the effect of climate change on the area enclosed by the 50 dB SPL (decibel sound pressure level) contour around an ensemble of 30 European airports by considering ‘historical’ (1985–2014) and future projection (2035–2064) timescales.

This is a follow-up study to previous work which examined how take-off distance and maximum take-off mass are projected to change in the future [1]. This work uses the same sites as the basis for studying how climb angles may change (disregarding engineering advances) and how neighbouring populations could be affected.

The Williams et al. [1] study in fact is one of a growing number of studies regarding the perturbation of take-off performance as the world warms (see references therein). However, literature on the effect of climate change on noise around airports is only recently starting to appear.

In 2022 [2], in a model study of Chios airport using observed temperature and wind values, it was postulated that climate change could potentially have caused an increase in the area of the 60 dB L_{den} contour (day–evening–night; discussed below in Section 4) of almost 5% in the 25-year period up to 2022 [2]. This was attributed to worsened take-off performance due to climate change-induced temperature and headwind perturbations reducing climb angles. Climate change can also influence atmospheric boundary layer stability [3], which in turn can affect sound propagation [4]. This previous study used the IMPACT model (discussed below) throughout, whereas, in the present study, IMPACT is used as a first-order comparator of noise contour area.

The aim of the present study is to isolate the projected changes to noise pollution solely due to climate change. As discussed in Padhra et al. [2], engines are undoubtedly becoming quieter due to technological advances, but this will not form part of this study.

Commerical airports are a major source of noise pollution. For example, in 2023, the EASA—the European Union Aviation Safety Agency—estimated that almost 3.5 million people in Europe experienced ‘*highly annoying*’ aircraft-related sound exposure [5]. The EASA defines this as an L_{den} level of at least 55 dB, which is a weighted average of exposure at different times of the day (i.e., night flights are more disruptive).

Disruption is not limited to the immediately noticeable auditory impacts on humans, however. Indeed, increased airport noise has been identified as a contributor to cardiovascular disease [6] and increased blood pressure [7] in local populations. Wildlife can also experience negative impacts through induced environmental changes affecting birds, whales, and insects, amongst others (see, for example, Alquezar and Macedo [8] and Dominoni et al. [9]).

A good approximation is that, for a given point in the vicinity of an aircraft taking off, the noise experienced by a stationary observer is a function of the distance from the source; the source sound intensity, which is in turn a function of the aircraft’s thrust setting; and the air’s coefficient of absorption. However, in reality, there are many other factors which affect the overall level of sound transmission such as climb angle and surrounding vegetation and/or buildings.

Methods for reducing the exposure to noise pollution include the Noise Abatement Departure Procedures (NADPs) published by the International Civil Aviation Organisation, the ICAO [10]. For example, procedures exist for reducing exposure closer to and farther away from the airport, respectively, by changing thrust levels, acceleration profiles, and climb angles appropriately [11].

In this study, however, the thrust level and the historical climb angle are—post-optimisation—assumed to be constant, whilst the future climb angles change with air density. Work to understand the effects of climate change on more complex take-off

procedures, as well as the economic impacts of climate change on take-off fuel usage, is ongoing.

Previously documented comprehensive or ‘consolidated’ tools for conducting quantitative studies of airport noise—i.e., involving the explicit calculation of noise propagation and resulting contours—include those from the following:

1. The UK Civil Aviation Authority; Aircraft Noise Contour (‘ANCON’) Model [12].
2. The USA Federal Aviation Administration:
 - (a) The Integrated Noise Model, ‘INM’ [13].
 - (b) The Aviation Environmental Design Tool, AEDT, [14].
3. The European Organisation for the Safety of Air Navigation (EUROCONTROL)’s Integrated Aircraft Noise and Emissions Modelling Platform, or ‘IMPACT’ [15].

In addition to these consolidated tools, there is rich literature studying aircraft noise impacts. This ranges from comparisons of simple noise models and their associativity with environmental impact analyses [16], to studies examining passenger and flight number growth [17] and comparisons of various methods of obtaining thrust values—and hence noise spectra—from the rate of the rotation of the engine [18]. An expansive review is beyond the scope of this study, however.

The present study differs from these tools in that its purpose is to present initial analysis from a new, free, and open-source software tool—available at <https://github.com/jonnyhtw/noyse> (accessed on 28 August 2025)—and encourage future development and research work with it. It uses projected (future) temperature and pressure data from a previous take-off performance study [1] to quantify potential changes in climb angle and hence impacts on local populations. In this initial, idealised study, we do not consider changes to fleet, engine type, or thrust, therefore isolating the effects which may be expected from climate change. These assumptions could be lifted in follow-up studies, however.

Many of the metrics presented below do not explicitly rely on the ability of the model to reproduce the precise shape of the decibel contours around the airports in question. However, motivated by the previous take-off performance study, we have developed a spatially resolved methodology which not only can show plausible ‘emulated’ contours for the airports in question, but could also be straightforwardly extended to arbitrary sites.

Emulation techniques are widely used in modelling studies of climate-relevant parameters and tools. For example, the FAMOUS climate model [19] simulates (synonymous with ‘emulates’ in this context) the 3D climate response of the higher-resolution HadCM3 model [20], meaning that it can produce results faster. In essence, it largely uses another model as its ‘ground truth’ target.

Another example of emulation is the use of statistical techniques which use large ensembles of climate model results to explore parts of parameter space which are not explicitly resolved in the individual model results themselves. The interested reader is referred to the review article of Castruccio et al. [21] for further information.

We use the area enclosed by a particular decibel contour from IMPACT as our emulation target and use a 20-member Latin hypercube sampling technique—see, for example, Gregoire et al. [22]—to minimise the difference in the values obtained from the two models. We also ensure that the salient features of previously published noise contours are reproduced, although non-trivial differences remain due to the different levels of model complexity.

The choice of emulation target model does not materially affect any of the results or conclusions presented here. IMPACT was chosen due to its availability and to maintain traceability with its use in the previous study of Padhra et al. [2], which motivated this work.

After the establishment of reasonable model parameters for present-day conditions, we proceed to apply them to future conditions at 30 European airports using a simple relation connecting climb angle and air density. In these future conditions—see Williams et al. [1] for details—air density shows an overall decrease, largely due to increased temperatures. This then leads to smaller climb angles and an increased noise pollution footprint.

The effect of changing population density around the sites is also taken into account to provide estimates of changes to the number of people affected by noise pollution under different future climate and land-use change scenarios.

Section 2 describes the outline model developed here, Section 3 details the relation between climb angle and air density, Section 4 describes the emulation procedure, Section 5 shows the results, Section 6 details recommendations for future model development with this tool, and Section 7 concludes.

2. Materials and Methods

2.1. Sound Propagation Model

The coordinate system used for noise intensity calculations at ground level is two-dimensional (x, y) and centred on the aircraft's take-off location and extends a set number of multiples of the runway length, L_{runway} , in $\pm x$ and $\pm y$. To simplify the calculations, the runway is assumed to run perfectly north–south (the y direction) and the true runway orientation is applied later by rotating the entire grid using third-order spline interpolation [23].

To account for the vertical, z , coordinate, after take-off, the aircraft is assumed to climb with a constant angle, ϕ . The Euclidean distance from the aircraft to each point in the grid is then calculated for every x, y, z in the trajectory. The resulting example noise level presented below for London Heathrow in Figure 1 is then simply the maximum value at each x, y in the aircraft's path.

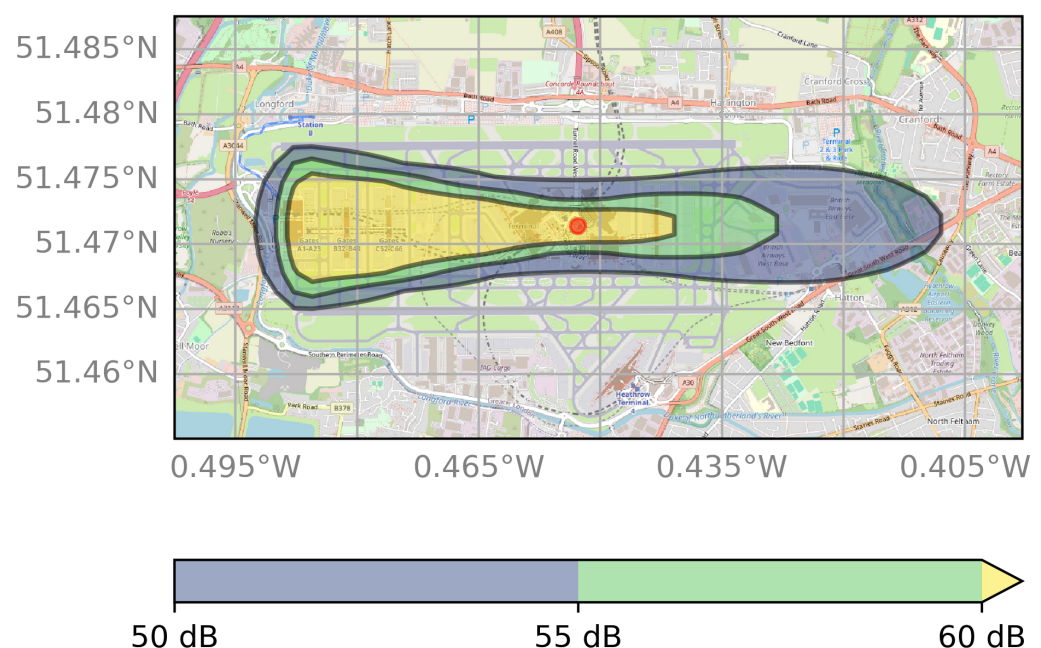


Figure 1. The extent of the 50, 55, and 60 dB contours around London Heathrow Airport calculated using the present emulator framework. The take-off direction is from west to east.

Assuming that sound radiation is isotropic, its intensity, I , is assumed to reduce with the distance, r , from the source with an inverse square law,

$$I = I_0 \left(\frac{r_0}{r} \right)^2, \quad (1)$$

where I_0 is the sound (power) intensity at a reference distance, r_0 , from the source, taken here to be 10 m. This value is discussed in detail in Section 4. In general, the exponent of r may differ from 2, especially very close to the source, but we do not consider this further here.

The power intensity at some arbitrary distance from the aircraft, I , has units of $\text{W} \cdot \text{m}^{-2}$, but usually, it is the intensity of the sound in decibels, I_{dB} , which is used, where

$$I_{dB} = 10 \log_{10} \left(\frac{I}{I_0} \right). \quad (2)$$

Combining the relations for I and I_{dB} gives

$$I_{dB} = I_{dB,0} - 20 \log_{10} r. \quad (3)$$

As well as the geometric factor, the sound intensity (in dB) is also reduced due to absorption processes, and this is included using a coefficient of absorption, α (units $\text{dB} \cdot \text{m}^{-1}$), which is a complex function of temperature and frequency (we assume 50% relative humidity and International Standard atmosphere pressure). The full expression is given in Bass et al. [24] and Figure 2 shows its variation between 0 and 50 °C for three frequencies: 50, 1000, and 1700 Hz. The highest of these is the ‘peak tonal frequency’ of A320 engines as reported in Figure 9 of Merino-Martinez et al. [25].

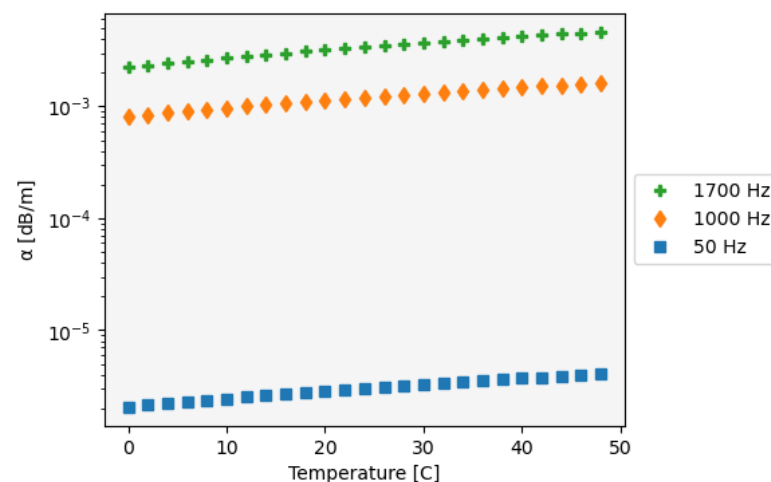


Figure 2. Absorption coefficient, α , as a function of temperature for three different frequencies. The full equation is described in Bass et al. [24].

We do not present results for values between 50 and ≈ 1000 Hz since the respective changes to the affected populations are uniform across this range. We do not consider higher frequencies in this study since it is the lower frequency range which—in general—causes more annoyance and deleterious health impacts [26].

This gives our final equation for sound fall-off with distance:

$$I_{dB} = I_{dB,0} - 20 \log_{10} r - \alpha r. \quad (4)$$

For the explicit consideration of multiple, n , engines (two here), the following expression is used:

$$I = 10 \log_{10} \left(\sum_n 10^{\frac{I_n}{10}} \right) \quad (5)$$

where each of the I_n is described by Equation (4).

In the earlier companion study [1], a force balance model was used to estimate the changes to take-off distance required and maximum take-off mass for aircraft at the same set of sites considered here. In this work, the acceleration (thrust) force is not a function of temperature. What this tacitly assumes is that any decreases in the density of the air—which, in reality, would cause a reduction in the resultant forward force—are precisely balanced out by changes in the engines' thrust settings. A full treatment of these scaling relations can be found in, for example, Kandula and Vu [27].

On the other hand, fan noise is an even more complex function of the air's ambient properties, and we make the further assertion that the fan noise decreases with increasing temperature due to changes in tip Mach number [28].

In reality, entirely separating the contributions of changing jet and fan noise is not possible due to the increase in fan rotation speed which results from a higher thrust setting. However it is pertinent to show a calculation of how these components' magnitudes change given an illustrative temperature range of 0–40 °C and put them in context of the wider uncertainties present.

The temperature increase (assuming constant pressure) causes a decrease in density and a concomitant decrease in thrust due to mass continuity. For a temperature difference of 40 °C and assuming ISA conditions, this is a decrease of $\frac{T+40}{T} \approx 1.15$, where T is zero Celsius measured in Kelvin.

Now, we require an estimate of the increase in sound power (in dB) for an increase in thrust of 15%.

Empirically, the sound power increases with the thrust raised to a power $n \in [6, 8]$ [29], and $10 \log_{10}(1.15^6) \approx 3.6$ and $10 \log_{10}(1.15^8) \approx 4.9$, which gives an estimate of the change in jet noise, N_J , of $\approx 4.2 \pm 0.6$ dB due to the increasing thrust setting of the engines to maintain constant ground acceleration.

For an estimate of the change to the fan noise, ΔN_F , over the same interval, we can examine Figure 4 of Malbéqui et al. [30], which shows that sound pressure from fan noise is a function of the tip Mach number, M . We convert the temperature change to a tip Mach number change estimate by assuming a representative tip speed of $340 \text{ m} \cdot \text{s}^{-1}$ (e.g., Sjögren et al. [31]) and calculate the speed of sound in air, v_s , as

$$v_s = \sqrt{\frac{c_p}{c_v} R_{\text{specific}} T} \quad (6)$$

where $c_{p,v}$ are the specific heat capacities at constant pressure and volume, and $\frac{c_p}{c_v} \approx 1.4$. This gives a difference in tip Mach number, ΔM , of

$$\Delta M = \left(\frac{340}{\sqrt{\frac{c_p}{c_v} R_{\text{specific}} (T + 40)}} \right) - \left(\frac{340}{\sqrt{\frac{c_p}{c_v} R_{\text{specific}} T}} \right) \approx -0.07 \pm 0.01. \quad (7)$$

and from Figure 4 in Malbéqui et al. [30], we obtain a value of $\Delta N_F \approx -3 \pm 0.5$ dB.

Bearing in mind the non-trivial uncertainties in the above calculations, as well as the exceptionally complex reality of the noise landscape around a turbofan engine, the fact that $N_J \approx -N_F$ justifies our assumption of the constancy of $I_{\text{dB},0}$, assuming negligible airframe noise at take-off [32] and $N_J \approx N_F$. It also, crucially, allows traceability with our previous

take-off performance study [1]. The assumption of $N_f \approx N_j$ will not be true under all operating conditions, and this should be borne in mind when considering the validity of the data and conclusions discussed below.

Specific data, from either models or test rigs, on how total (at source) jet noise varies as a *controlled* function of ambient air density is strikingly absent in the literature. This is partly due to the standardised nature of many noise assessment tools, imposing simple thermodynamic corrections to a ‘standard day’ of, for example, 25 °C (e.g., <https://www.legislation.gov.uk/ukxi/1990/1514/made>, accessed on 28 August 2025), as well as their almost-universal concentration on far-field effects, i.e., those of relevance to urban and peri-urban planners.

McAninch and Shepherd [33] even state—and show in their Figure 1—that, for the INM model discussed above, ‘*the aircraft noise measured for one set of atmospheric and aircraft operating conditions is assumed to be applicable to all other conditions as long as the corrected net thrust remains constant*’, which is indeed the case assumed here.

One of the only studies which explicitly considers observational campaign data with that from the semi-empirical models of, for example, Heidmann for N_F and Stone for N_j is the 2022 study of Vieira et al. [34]. This work compares the measured and modelled noise levels of various aircraft types (including the A320) at Amsterdam Schiphol Airport, and considers airframe, combustion, fan, and jet noise for a representative range of conditions in June, July, and September 2019. There are two key conclusions which are of specific interest to our work:

1. The overall sound pressure level (OPSL) range reported for the A320 is the smallest of any of the aircraft types considered, with an interquartile range (itself larger than the standard deviation of a normal distribution) of just ≈ 1 dB.
2. The only frequency range in which a good agreement between total (modelled) and measured noise occurs is above the frequency range considered by this study in the regime where fan noise dominates. For lower frequencies, the disparity can be as large as ≈ 20 dB, depending on the take-off speed, which is itself yet another uncertainty contributor.

In summary, we are aware that the assumed constancy of $I_{dB,0}$ represents a significant simplification of the real-world physics. However, even state-of-the-art noise-component models fail to reproduce observed values, and so we posit that our assumption is an unavoidable one given its use in previous published work using this input climate model dataset, as well as the lack of explicit data to the contrary from, for example, temperature–noise regression models.

Further work to examine the effect of varying $I_{dB,0}$ values in line with the Vieira et al. [34] study is ongoing.

In practice, the sound emission from a standard turbofan engine is anisotropic and highly complex—e.g., [35,36]—and is parameterised here by increasing the sound levels both behind and in front of the engine’s intake axis (the y direction, i.e., before taking the runway bearing into account) compared to the lateral, x , direction. This weighting, W , is shown diagrammatically in Figure 3 and described mathematically by

$$W = \exp\left(\frac{-\gamma^2}{2\omega^2}\right), \quad (8)$$

where γ is the angle over which multiple take-off paths are run in order to represent the enhanced power intensity in the direction parallel to the direction of travel and ω is an empirically determined scaling parameter. Each path is rotated around the centre of the

grid, which is defined as the starting position of travel (for the central estimate, $\gamma = 0$), where $x = 0$ and $y = -d_{\text{runway}}$.

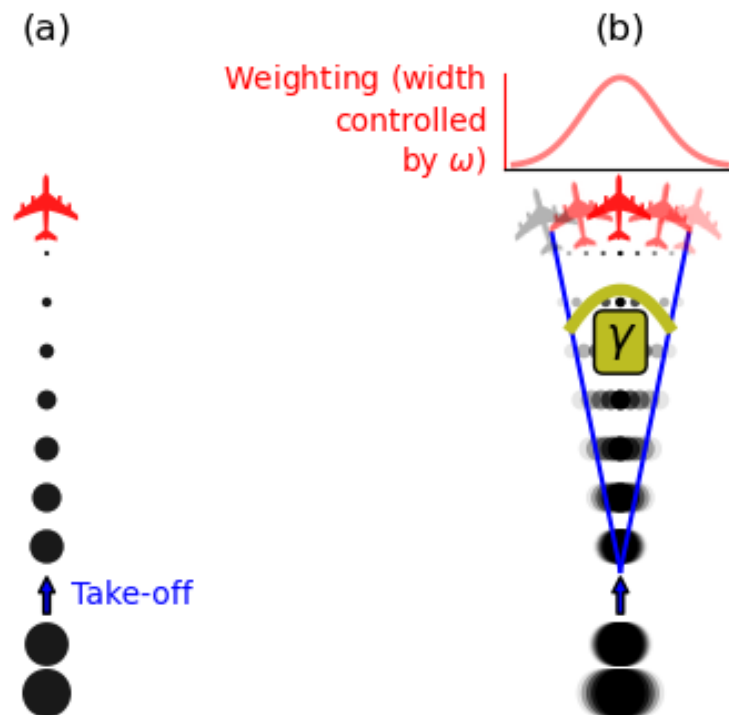


Figure 3. (a) Spherically isotropic sound propagation and (b) image modified to include a preferentially amplified sound intensity in the forward ($+y$) and reverse ($-y$) directions. Also shown in (b) are the angle γ over which the sound is amplified, as well as their respective weightings, parameterised by ω .

Equation (8) illustrates Gaussian directional weighting—see, for example, Lyu et al. [37]. Other parameterisations of this angular dependence (exponential; cosine) could be straightforwardly included in extensions to the model described here. However, for the A320's engines at take-off, a normal distribution is often used; see, for example, Clark et al. [38].

The γ parameter's value of 20° (see Table 1) is gleaned from Figure 2a from the previous study of Padhra et al. [2], and ω was obtained heuristically to fit with IMPACT's results. The value of $\omega = 31.8^\circ$ fits well within the expected range of $\approx 25\text{--}40^\circ$ (e.g., from Khavaran and Bridges [39]).

Table 1. Parameters used for the generation of Figure 1. The mx2t24 and P_s values are means of 30 years of historical data from the ACCESS-ESM1-5 climate model. The mx2t24 variable is the maximum temperature recorded at a site in the preceding 24 h and is the metric used for daily temperature variation, as in [1].

Parameter	Value	Unit
$I_{\text{dB},0}$	98.90	dB
γ	20	degrees
ω	31.8	degrees
ϕ	7.25	degrees
$\mathcal{A}$	4.82	km ²
A_{50}	4.71	km ²
mx2t24	20.06	K
P_s	1019.1	hPa
d_{engines} [40]	11.2	m
d_{runway} [1]	3902	m

2.2. Motivation and Utility of the Current Model

As mentioned above, there are multiple models available which are capable of modelling noise contours based on assumptions of atmospheric conditions and aircraft type, etc. However, common properties which reduce their utility and uptake include closed-source code bases, licencing restrictions, a lack of in-built version control, and the need to be run on remote servers. The software presented here is not subject to any of these limitations.

This is not just a matter of convenience and transparency; for many research funding bodies, openness and adherence to the FAIR principles of software development (Findable, Accessible, Interoperable, and Reusable) are *a priori* requirements.

For example, research outputs and software developed by projects funded by the European Union (https://commission.europa.eu/funding-tenders/find-funding/eu-funding-programmes/horizon-europe_en, accessed on 28 August 2025) are required to be open access, and the UK government is also currently consulting on its new Research Data Policy with the requirement that ‘publicly funded research data is made as open as possible’ [41].

Whilst these policies do not, yet, explicitly mandate the use of pre-existing open source software, there is a clear and ongoing trend in this direction.

Our software also has a built-in ‘batch’ mode. Because of this, even though the code runs serially (on one processor), arbitrary numbers of simulations can be run concurrently on a remote cluster of, crucially, the user’s choosing. In the current work, the Reading Academic Computing Cluster (RACC) was used to submit hundreds of simulations at the same time to the Slurm Workload Manager (<https://slurm.schedmd.com/>, accessed on 28 August 2025) and a standard run of 300 jobs (30 airports $\times$ 10 climate models) can take as little as ~ 10 min, or just 2 s per model integration on average.

3. Effect of Climate Change on Climb Angle

As greenhouse gases accumulate in the atmosphere and the air warms, the air density tends to decrease in line with the ideal gas law,

$$P = \rho R_{\text{specific}} T, \quad (9)$$

where P is the air pressure (which also varies in the models), ρ is the air density, R_{specific} is the specific gas constant for air ($287.05 \text{ J} \cdot \text{kg}^{-1} \cdot \text{K}^{-1}$), and T is the temperature.

Using 10 state-of-the-art climate models forced with three different scenarios of (summer only) 2035–2064 greenhouse gas forcing compared to 1985–2014, daily time series of temperature and air pressure were obtained, and from these values, the predicted air density was calculated using Equation (9). The future scenarios used (Shared Socioeconomic Pathways, SSPs) are SSP1-2.6, SSP3-7.0, and SSP5-8.5. The format SSPX-Y refers to the SSP ‘family’, X, and the net radiative forcing per m^2 at 2100, Y. These can be considered as low-, medium-, and high-magnitude climate change amounts, respectively [42].

The historical period does not extend beyond 2014 since this is when the relevant simulations in the 6th Coupled Model Intercomparison Project end [43]. This also removes the influence from the unprecedentedly warm years since 2014 [44], thereby increasing the signal-to-noise ratio in our conclusions.

The dependence of aircraft climb angle is a highly complex function of many variables such as atmospheric conditions, thrust level, aircraft variant, and altitude (see Takahashi and Sóbester [45] and references therein). However, in order to provide a straightforward and accessible presentation of this in software, we assume the following force balance at take-off:

$$\tau - D = W \sin \phi, \quad (10)$$

where τ , D , and W are the thrust, drag, and weight, respectively, and ϕ is the climb angle (see, for example, [46], Supplementary Material Part 2).

We assume that the variables in Equation (10) maintain their respective relationships with air density as the climate changes. There are many reasons why, in practical engineering terms, this assumption will not hold (e.g., fleet renewal) but it is an important assumption in order for us to isolate the effect of climate change on the parameters of interest in this study. This could of course change in future studies.

We further make the assumption that both τ and D are proportional to air density, which is a good approximation assuming incompressible flow in International Standard Atmosphere conditions [47]. This leads to the proportionality relation

$$\rho \propto \sin \phi, \quad (11)$$

or, equivalently,

$$\frac{\sin \phi_{\text{hist}}}{\rho_{\text{hist}}} = \frac{\sin \phi_{\text{future}}}{\rho_{\text{future}}}, \quad (12)$$

and

$$\phi_{\text{future}} = \arcsin\left(\sin \phi_{\text{hist}} \frac{\rho_{\text{future}}}{\rho_{\text{hist}}}\right). \quad (13)$$

Therefore, we now have an equation for describing how the climate change-affected climb angle can be calculated as a function of the historical and future air densities. The climb angle in the historical situations considered is assumed constant in this work, and its range of validity is discussed in Section 4.

4. Parameter Uncertainty and Latin Hypercube Parameter Sampling

In the model described above—indeed, as is the case for any model—there are many approximations and parameterisations. For example, the sound intensity near the source and the initial climb angle are both assumed constant, neither of which is strictly adhered to in real flight operations. Additionally, the inverse square relation and absorption coefficient described in Equation (4) deviate from these idealised mathematical formulations due to, for example, ground reflection and humidity dependence, respectively [36,48].

To account for these uncertainties, we have established plausible ranges of $I_{\text{dB},0}$ and ϕ and then used Latin hypercube sampling (LHS) to establish parameter values which enable a meaningful comparison with IMPACT, noting again that, in practice, the choice of comparator model for this initial model description would not affect our present conclusions.

LHS efficiently samples the parameter space defined by the input variables (ϕ , $I_{\text{dB},0}$) to avoid having to exhaustively sample the full two-dimensional landscape of possible combinations [22]. It is also straightforwardly extensible to much larger sets of parameters which may be explored in future studies with this software.

We used the area enclosed by the 50 dB L_{den} noise contour around London Heathrow airport from the IMPACT model (<https://www.eurocontrol.int/platform/integrated-aircraft-noise-and-emissions-modelling-platform>, accessed on 28 August 2025), $\mathcal{A}$, as our verification target. The 50 dB level is a commonly used baseline for aircraft-related noise disturbance [49,50] and is broadly representative of a residential, suburban environment [51]. IMPACT is formulated using guidelines from the Report on Standard Method of Computing Noise Contours around Civil Airports from ECAC [52]. This documents, in detail, the different types of noise which are generated by an aircraft.

The output for the model presented here represents one take-off procedure and is ‘unadjusted’; that is, we do not perform any weighting based on frequency, and this is compared against the ‘day–evening–night’ level output from IMPACT. This represents the average noise exposure over a 24 h period [53], includes adjustments for the sensitivity of human hearing, and is given by

$$L_{\text{den}} = 10 \cdot \log_{10} \left\{ \frac{1}{24} \left(12 \times 10^{\frac{L_{\text{day}}}{10}} + 4 \times 10^{\frac{L_{\text{evening}}+5}{10}} + 8 \times 10^{\frac{L_{\text{night}}+10}{10}} \right) \right\}, \quad (14)$$

which shows the 5 and 10 dB ‘penalties’ which are applied to evening and night flights, respectively. Using L_{den} alleviates some of the inherent issues in the comparison with IMPACT’s greater complexity and capability (see the Discussion on page 11).

To obtain initial estimates of the bounds for $I_{\text{dB},0}$ and ϕ , we first calculated the area enclosed by the 50 dB SPL contour (A_{50}) using $\phi = 7.7^\circ$, which is given as a typical value in Gratton et al. [46] and $I_{\text{dB},0} = 100$ dB for each engine at 1700 Hz.

The choice of 100 dB is the value at 10 m from an engine with a source intensity of 120 dB, assuming isotropic transmission from 1 m away (e.g., Huang and Zheng [54]). This 10 m ‘buffer’, as well as being of the order of the inter-engine distance, d_{engines} , also allows for some amelioration of the reflection effects referred to above. Therefore—and since our model is intended as an emulator rather than a simulator—the precise values used are not crucial as long as they emulate the assumed-equivalent L_{den} value from IMPACT, or indeed any other target model.

Using $I_{\text{dB},0} = 100$ dB and $\phi = 7.7^\circ$, we obtain $A_{50} = 5.32 \text{ km}^2$, compared to $\mathcal{A} = 4.77 \text{ km}^2$, an overestimate of $\approx 11\%$. Considering the assumptions and simplifications in the present model, this is an encouragingly small discrepancy. However, it is important to perturb the parameter assumptions to make sure that this is not entirely serendipitous.

The climb angle used in the L_{den} calculation in IMPACT, when averaged over its first 10,000 ft, is approximately 7.2° . For $I_{\text{dB},0}$, from its maximum at ‘take-off roll’, the power setting is reduced by up to $\approx 40\%$ over the course of the first few nautical miles of its trajectory. A 40% decrease is equivalent to roughly 2 dB since, from Equation (2), $10 \log_{10}(0.6) \approx -2$, and so we use $98 < I_{\text{dB},0} < 100$ dB and $7.2^\circ < \phi < 7.7^\circ$. The results of this sampling are shown in Figure 4 for a sample size, N , of 20. In LHS studies, the recommended sample size is $N \geq 10$ times the number of free parameters ($I_{\text{dB},0}$, ϕ) [22].

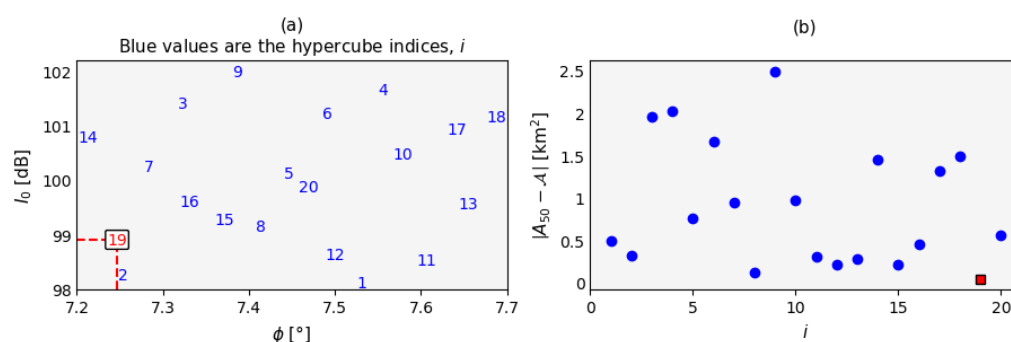


Figure 4. (a) Scatter plot of the values of $I_{\text{dB},0}$ and climb angle, ϕ ($N = 20$), directly from the Latin hypercube sampling. The respective optimum values of $I_{\text{dB},0}$ and ϕ are shown by the ordinate and abscissa values corresponding to the lowest shown value in (b), i.e., where the value of A_{50} is closest to $\mathcal{A}$. (b) Values of $|A_{50} - \mathcal{A}|$, blue circles, for each value, i , in the sampling distribution. The lowest value is marked with a red square.

Figure 4 shows that the 19th member of the LHS distribution gives the best agreement between A_{50} and $\mathcal{A}$: a difference of 0.11 km^2 . This corresponds to $I_{\text{dB},0} = 98.90 \text{ dB}$ and $\phi = 7.25^\circ$, and these values are used for the remainder of this study.

We justify our usage of single frequencies in our model output since the L_{den} metric from IMPACT is only used in the generation of the $I_{\text{dB},0}$ and ϕ parameters for the peak tonal frequency of 1700 Hz discussed above. Additionally, using the IMPACT model for the generation of a single take-off operation's noise level at a single frequency results in a noise distribution which is not well resolved by the current model's grid without significantly increasing its run time by enhancing grid resolution. Further work with this nascent framework to explicitly calculate L_{den} is encouraged.

5. Results

5.1. Example Noise Contours

Figure 1 shows the results of the verification procedure described above using London Heathrow airport for the parameters shown in Table 1.

Figure 1 shows the 50, 55, and 60 dB noise contours around London Heathrow airport and shows the effect of the anisotropic power intensity, amplified in the direction of travel (to the east in this case), and the tapering-off of the sound as the aircraft climbs.

In addition to the IMPACT comparison detailed above, the overall shape of the contours is in good empirical agreement with previous work by, for example, Tlałka and Rżucidło [55], and Sahai et al. [56] and with policy documentation from the US Federal Aviation Administration [57]. Indeed, considering the relative simplicity of our model, the agreement with previous results is encouraging. In any case, the precise shape does not change any of the conclusions drawn below since we are examining *changes* to the noise contour area (and affected populations) rather than absolute values.

5.2. Effect of Atmospheric Conditions on A_{50}

Figure 5 shows the projected increase in A_{50} around each site as a function of the climate change scenario used in the future modelling at 50 Hz. For higher frequencies, absorption is increased and reduces the magnitude of the A_{50} changes. This is discussed in Section 5.3 when population changes are also taken into consideration. The format of this figure uses box-and-whisker plots which show the distributions' median and interquartile range and $1.5\times$ this range either side of the first and third quartiles. Each sub-figure illustrates the effects of the varying air density changes across the sample sites as discussed in detail in Williams et al. [1]. It also takes into account the varying runway lengths by assuming that the take-off distance used is equal to either that at maximum take-off mass (2362 m [2]) or the respective runway length, whichever is shorter. This additional site-dependent assumption tacitly means that we are assuming that the aircraft in question are operating at the maximum allowed mass.

The effect on A_{50} shown in Figure 5 is broadly uniform across the locations considered. This, however, is important to emphasise given the significantly differing climates of the sites considered (e.g., Oslo, Madrid, and London) and the novelty of the method of the calculation of climb angle change. In addition, when we come to consider the effects of their differing (and temporally changing) population densities, this uniformity disappears.

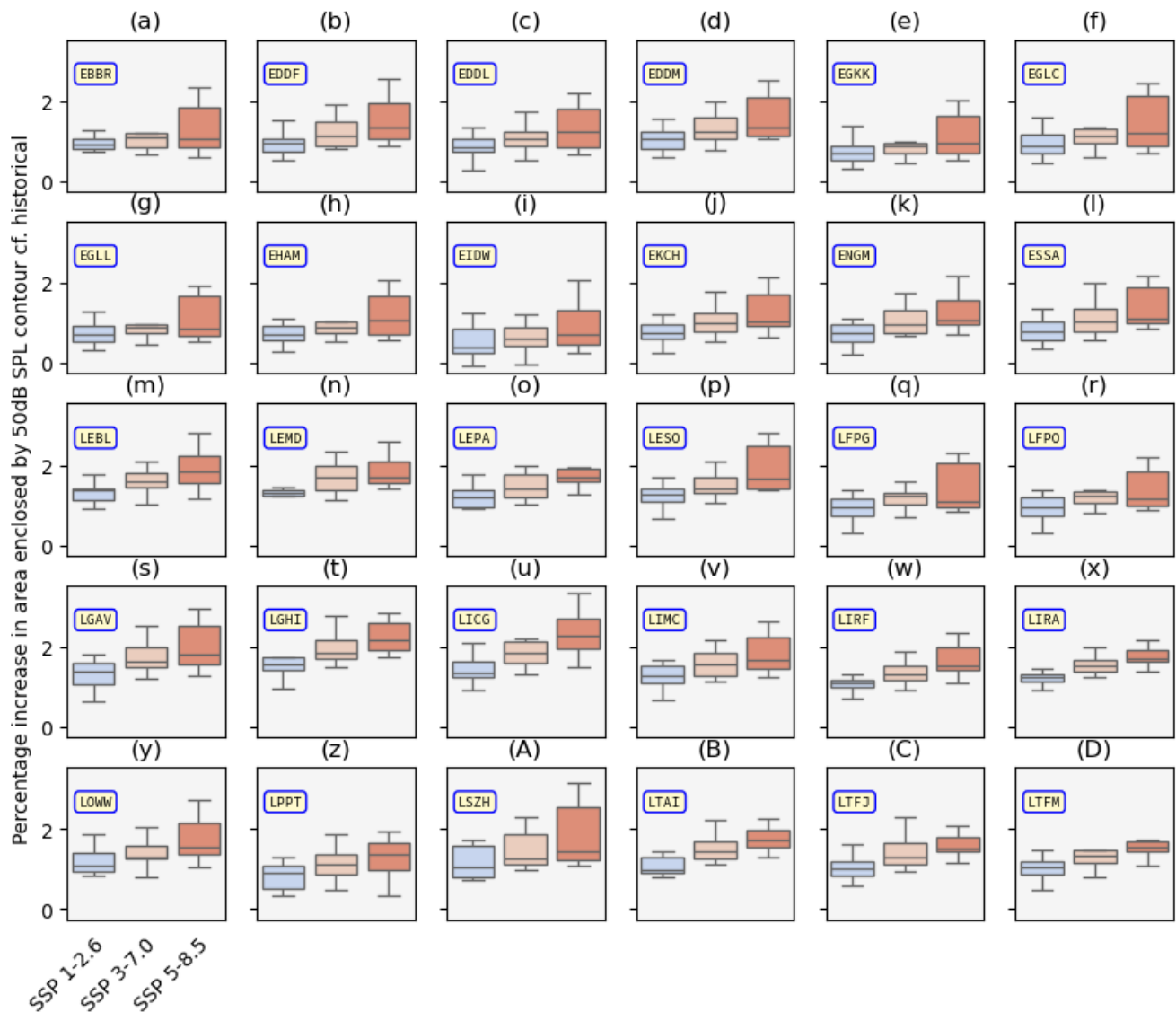


Figure 5. Percentage change to the area enclosed by the 50 dB SPL contour (A_{50}) at 50 Hz for each of the study sites and climate change scenarios studied (a–z, A–D).

There is significant inter-model spread within the individual scenarios' results in Figure 5, and the most extreme warming scenario considered, SSP5-8.5, shows the largest variability. This is the expected behaviour—as previously noted in Williams et al. [1]—because of the range of equilibrium climate sensitivity responses from the models in the ensemble; that is, the susceptibility of temperature change to changing greenhouse gas concentrations.

The values shown in Figure 5 show the response of A_{50} for each climate model to the *mean* change in air density in each of the scenarios compared to the historical period. This averaging masks the day-to-day variability in the summer temperatures and pressures. To investigate the full distribution of these values, Figure 6 shows the (daily maximum) temperature, daily mean pressure, and resulting ϕ values for UKESM1-0-LL and SSP5-8.5 at another of our study sites, San Sebastián in Spain.

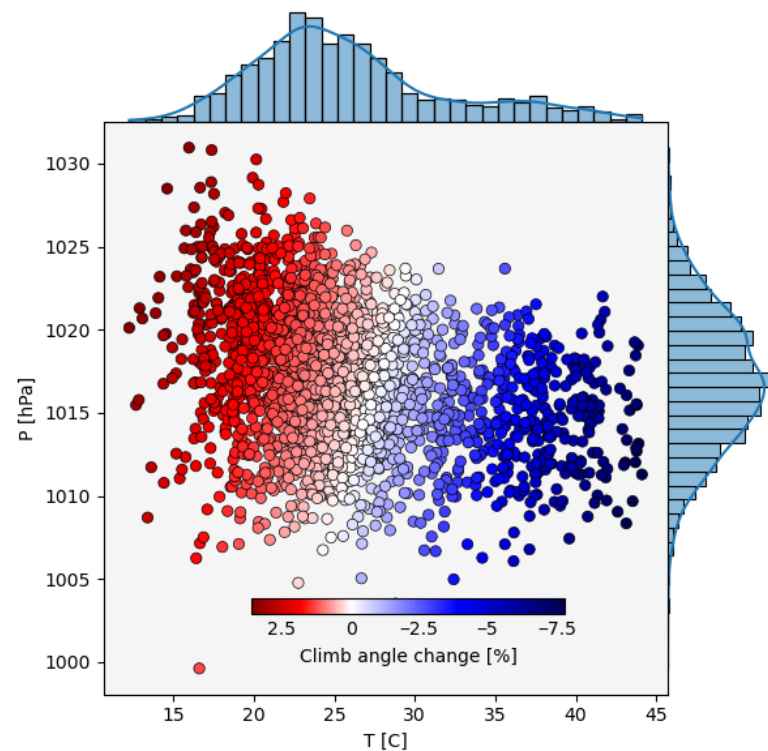


Figure 6. Scatter plot of climb angles (compared to the assumed historical constant value in Table 1) calculated for San Sebastián airport under SSP5-8.5 conditions using the UKESM1-0-LL climate model [1]. The marginal plots on the top and right show the relative frequency of occurrence of the daily maximum temperature and daily mean pressure values, respectively. Note that the distribution shown here is independent of frequency since this only reflects climb angles and not the impact of any resulting changes to A_{50} .

Figure 6 shows the individual ϕ values for each day in the period 2035–2064 for SSP5-8.5 and clearly shows the expected decrease in ϕ with increasing temperature as discussed above. It also shows decreases in ϕ as high as 7.5% (blue values), which is significantly higher than the values of $\Delta\phi \approx 1\text{--}3\%$ obtained in the calculation of the A_{50} data in Figure 5, which only use the mean ϕ value.

In the next section, we examine how these geometric variables can be translated into quantified impacts on populations.

5.3. Climate Change Impacts on Residential Populations near Airports

Although the changes to A_{50} shown in Figure 5 may appear small, we shall see below that, once the population density around airports is taken into account (which can exceed 10^4 people per square km), the number of additional people, compared to historical values, potentially affected by these changes can be significant, even without considering extreme values (i.e., Figure 6).

With this in mind, we use the Olén and Lehsten [58] dataset at a 30-arc-second (≈ 1 km) resolution, which gives population densities in a manner corresponding to the future climate model scenarios used here. Figure 7 shows the population densities around our 30 study sites for each SSP scenario.

The metric of primary interest in this study is the *change* to the number of people—relative to the historical baseline—affected by perturbations to the area enclosed by a given decibel contour.

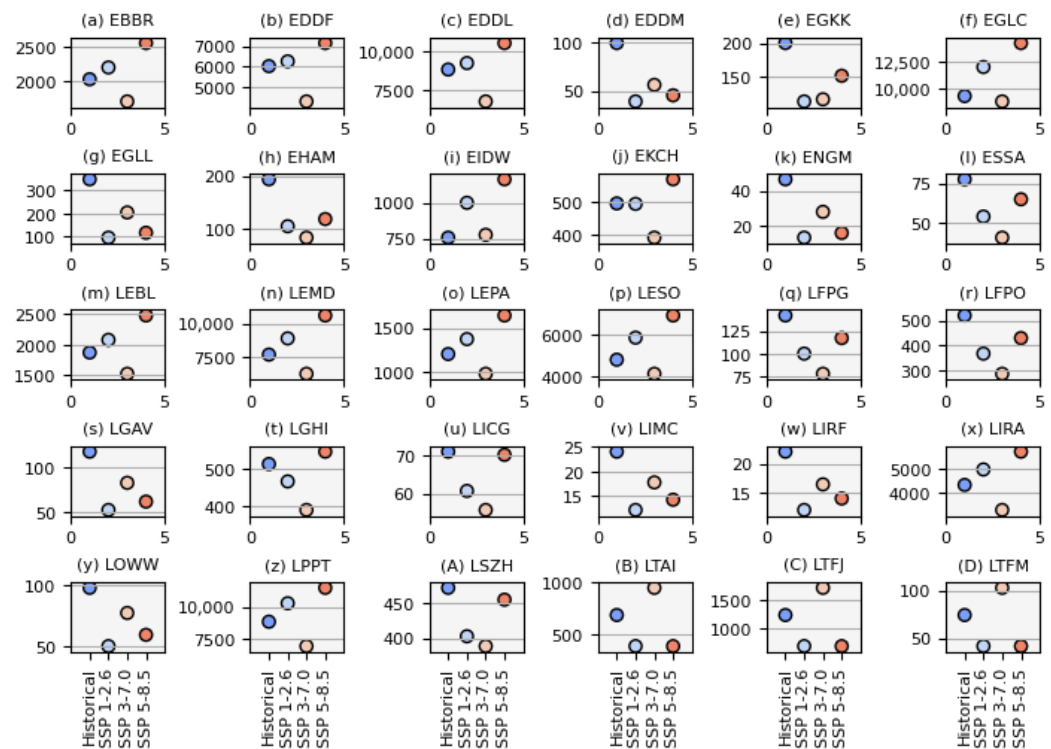


Figure 7. Population densities (people per square kilometre) at a 30-arc-second (≈ 1 km) resolution for each of the study sites calculated using data from Olén and Lehsten [58]. The titles of each sub-figure (a–z, A–D) show the ICAO airport code, and the population density is averaged over a 3×3 km area around each location.

Figure 8 shows the change in the number of people affected by noise pollution for each climate change scenario and for the three frequencies considered. This is found by taking the population density (Figure 7) and the change in the area of the 50 dB contour (Figure 5) into account.

Figure 8 shows that the fractional changes are the highest for low frequencies, and that, for frequencies around the peak tonal frequency of 1700 Hz, the expected changes are negligible. The values for 50 Hz lie between ≈ 1 and 4% and for, say, London City airport (airport code EGLC), this corresponds to values of up to 500 people per square kilometre, or 2500 when taking into account the area of the 50 dB contour of ≈ 5 km². These values are reduced slightly for 1000 Hz and are negligible for 1700 Hz. This can be attributed to the increased attenuation at higher frequencies as shown in Figure 2, which will also reduce the values shown above for 50 Hz in Figure 5. In general, exposure to low (compared to high) frequency noise causes more health disbenefits [26] and indeed is known to cause more ‘annoyance’, especially in the 20–60 Hz range [59]. Therefore, in summary, our results indicate that climate change has the potential to increase the number of people affected psychologically and physically by noise pollution in the future. In this preliminary analysis, we have not taken the different relative frequency weightings into account, but this should be performed in future studies with this model.

As the frequency is increased, some values of the changes in Figure 8c become negative. This is because the increase in sound attenuation—Figure 2—outweighs the increase in A_{50} due to the climb angle decrease. Figure A1 in Appendix B shows the same data as in Figure 8 but in absolute terms rather than fractional ones. It also shows outliers in the model ensemble which are more than 1.5 times the interquartile range away from the first or third quartiles, respectively. We do not consider this A_{50} reduction effect further here, but it should be borne in mind in future work.

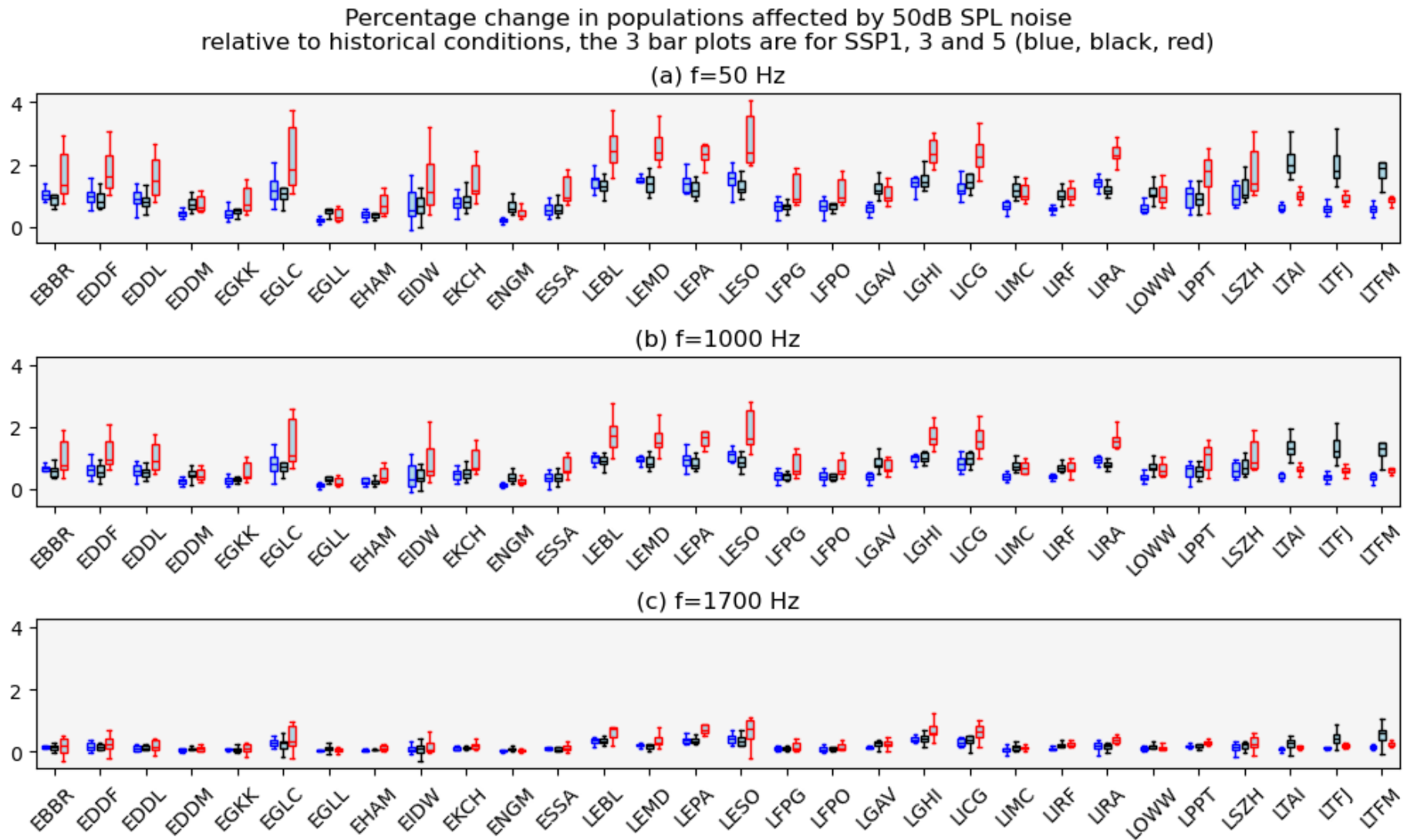


Figure 8. As for Figure 5 but showing the population changes taking into account the historical average area for each site and the population densities in Figure 7. Each panel shows the results of the simulations described above for each frequency given in the respective titles. Figure A1 in Appendix B shows the same data in absolute terms and without common y-axis limits.

6. Recommendations for Future Model Development

There are several key aspects which should be considered in future use of the model; either as presented here or as part of future development.

Firstly, the present model assumes a simple force balance leading to a linear, analytical relation between air density and climb angle. In practice, this will be complicated by factors such as variations in thrust level, headwind, and aircraft type. Because of these uncertainties, and the flexibility and relative simplicity of the model, future studies using different functional relationships are encouraged.

A related ‘upgrade’ to the current model could include more detailed verification metrics. The present study uses the area enclosed by the 50 dB SPL noise contour, and compares this to an assumed-equivalent L_{den} value from the IMPACT model. Future work could include using a range of decibel values, a larger number of evaluation target models, and an explicit calculation of L_{den} .

From a geographical perspective, this study—by design—is limited to European airports, and so a further extension to this work could include a larger, ideally global, domain. To enable this, the site specificity and raw climate model data bias correction steps—see Williams et al. [1] and references therein—could be embedded within a common software framework. This would have the distinct advantage of being able to address the intimately related issues of noise pollution and take-off performance restrictions in a unified manner.

The horizontal resolution of the raw climate model input data itself (~ 100 km) is an aspect which will lead unavoidably to increased similarity between the results obtained from closely neighbouring sites. For example, although the model data used in this study has undergone data processing and bias correction, the three airports studied in the London region, for example, are close enough for their overall climate response to be essentially the same. This of course has the distinct advantage of making the datasets small enough to be easily distributable and processable on standard PCs and data connections; however, it would be of great interest to consider higher-resolution (raw) input datasets, especially with the advent of new, ultra-high-resolution regional climate projections [60].

Finally, extensions to this modelling framework to include more realistic noise spectrum components, e.g. airframe, as well as the explicit inclusion of atmospheric property-dependence on jet and fan noise, and a non-constant $I_{dB,0}$ value are encouraged.

7. Conclusions

In this work, we have presented the results of a new, open-source model which has been used to project the climate change-induced changes to populations affected by noise pollution at 30 European airports.

The model was optimised using the Latin hypercube sampling of two key parameters—near-engine sound intensity, and climb angle—to emulate the area enclosed by the 50 dB day–evening–night averaged sound level (L_{den}) contour produced by the IMPACT model.

After this verification step, we used results from 10 state-of-the-art climate models to project how climb angle may change by mid-century compared to a historical baseline. This is based on three possible future scenarios of greenhouse gas emissions which cause varying degrees of perturbation to air temperature; pressure; and, ultimately, air density, which is the input parameter in the climb angle calculation method used.

By combining the projections of the changes to the 50 dB SPL contour area (A_{50}) and local population densities, changes to the number of noise-impacted residents have been estimated. This has been performed for all 30 airports and each climate change scenario with the differing warming responses of the climate models providing a robust uncertainty estimate.

Although the fractional changes to A_{50} values are broadly uniform across sites, ranging from approximately 1 to 2% at 50 Hz, the population density may change significantly in the future. By considering fractional changes to populations affected by noise pollution over three distinct frequencies, we have shown that low-frequency noise is expected to affect as many as 4% more people in the future. This increase in the number of people affected could be as many as 2500 in the most densely populated areas, such as central London. This does not take engineering advances into account, such as quieter engines, but it should be considered as part of the package of measures taken into account by stakeholders going forward. The consistent relative response uses values of input variables averaged across the respective 30-year periods. By considering the day-to-day distributions of temperature and pressure (with insights into how extreme values may change in the most pessimistic climate change scenario considered), values of $\Delta\phi$ as high as 7.5% could result, with concomitant increases to affected populations.

Although this study ultimately focuses on the atmospheric physics of climate change, the results related to the number of people affected by noise pollution are of clear socio-economic importance. The 2020 study of Mashhoodi and van Timmeren [61] examined 76 European airports, subsetting them into five categories based on their population density, road network connectivity, and other factors. Of these five, the most relevant to this study are those designated as ‘urban’ or ‘urban periphery’ sites. Relatedly, the European Commission estimates that, by 2050, over 83.7% of Europe’s population will be living in ‘urbanised’ areas, up from 76.5% in 2015 [62,63]. A more fine-grained analysis taking these data into account in an expanded future study could provide more general recommendations for planners and airport operations professionals. A pertinent study is that of Wolfe et al. [64], which estimates the populations impacted by the combined operations of London Gatwick and Heathrow airports, monetises the environmental damage caused, and discusses the effects of operational and technological improvements in the sector.

It is hoped that this study, and others in the same vein, will help to assist in alleviating uncertainties in—and improve the co-development and sustainable planning of—near-airport developments.

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Data Availability Statement: The raw data presented in the study are openly available in the Earth System grid Federation—ESGF—repository [43].

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Conflicts of Interest: Author Marco Venturini was employed by the company Amigo s.r.l. The remaining authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

Abbreviations

The following abbreviations are used in this manuscript:

AEDT	Aviation Environmental Design Tool
ANCON	Aircraft Noise Contour (model)
CMIP6	6th Coupled Model Intercomparison Project
EASA	European Union Aviation Safety Agency
FAMOUS	Fast Met Office/UK Universities Simulator
HadCM3	Hadley Centre Coupled Model version 3
ICAO	International Civil Aviation Organisation
IMPACT	Integrated Aircraft Noise and Emissions Modelling Platform
INM	Integrated Noise Model
IPCC	Intergovernmental Panel on Climate Change
ISA	International Standard Atmosphere
LHS	Latin hypercube Sampling
SSP	Shared Socioeconomic Pathway

Appendix A. Mathematical Symbols Used

Table A1. Mathematical symbols used.

Symbol	Definition
I	Sound intensity
r	Distance from sound source
ρ	Air density
ρ_{ISA}	Air density under International Standard Atmosphere conditions at sea level
P	Pressure
α	Coefficient of sound absorption
$c_{p,v}$	Heat capacity at constant pressure and density respectively.
γ	Angle over which noise anisotropy is calculated
ω	Empirical scaling parameter controlling the shape of the noise anisotropy
$R_{specific}$	Specific gas constant for air
ϕ	Climb angle
τ	Thrust
$\mathcal{T}$	Temperature in Kelvin at 0 °C
D	Drag
W	Weight
L_{den}	Day–evening–night noise level from IMPACT
A_{50}	Area enclosed by the 50 dB SPL contour in the emulator
$\mathcal{A}$	Area enclosed by the 50 dB L_{den} contour in the IMPACT model

Appendix B. Absolute Population Changes

Absolute change in populations affected by 50dB SPL noise
relative to historical conditions, the 3 bar plots are for SSP1, 3 and 5 (blue, black, red)

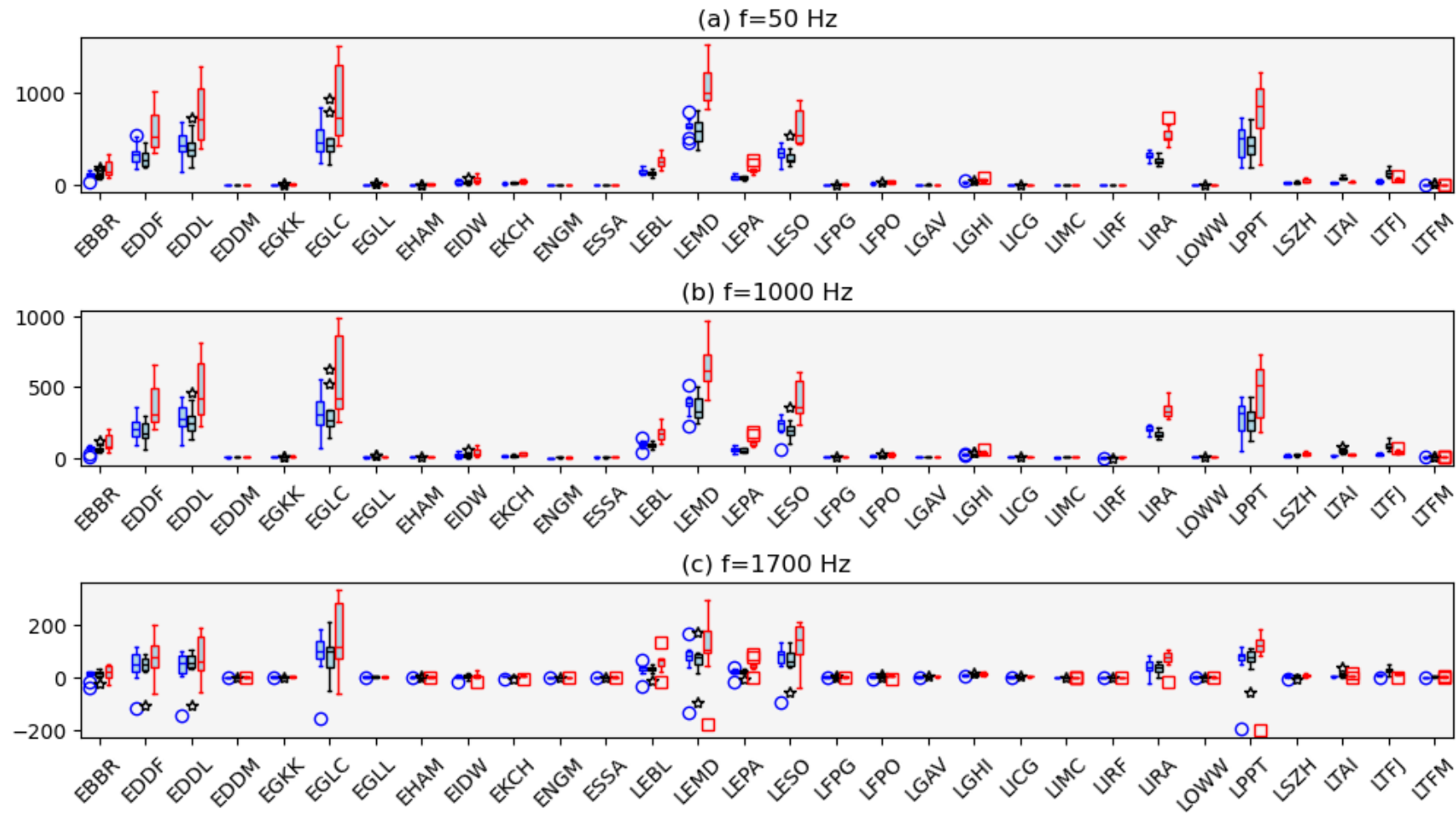


Figure A1. As for Figure 8 but showing absolute population changes and outlying values which are greater than 1.5 times the interquartile range from the first or third quartiles, respectively. SSP1-2.6 outliers are circles, SSP3-7.0 are stars, and SSP5-8.5 are squares.

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