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A synthesis is emerging between biodiversity-ecosystem function and ecological resilience research - Reply to Mori

Tom H. Oliver,1,2* Matthew S. Heard,2 Nick J. B. Isaac,2 David B. Roy,2 Deborah Procter,3 Felix Eigenbrod,4 Rob Freckleton,5 Andy Hector,6 C. David L. Orme,7 Owen L. Petchey,8 Vânia Proença,9 David Raffaelli,10 K. Blake Suttle,11 Georgina M. Mace,12 Berta Martín-López,13,14 Ben A. Woodcock,2 and James M. Bullock2

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A recent paper by Mori [1] states the need for a unification of studies of ‘engineering’ and ‘ecological’ frameworks of resilience. Engineering resilience focuses on the capacity of a system to recover to equilibrium following some kind of perturbation, whilst ecological resilience explicitly recognizes multiples stable states and the capacity for systems to resist ‘regime shifts’ between alternate states. We find Mori’s argument somewhat surprising given the number of recent biodiversity-ecosystem functioning studies (B-EF) that incorporate aspects of both resistance and recovery [e.g. see references in 2, 3]. We would argue that a synthesis is well underway and that apparent discrepancies are more due to differences in the spatial, temporal and systems scale of focus, and ambiguities in defining this study context, rather than any fundamental incompatibilities in conceptual frameworks.

With regards to our recent review on the mechanisms which underpin the resilience of ecosystem functions [3], Mori states: “To avoid confusion, resilience in this case should be explicitly termed as recovery or defined as the analogy of engineering resilience”. We clearly consider both recovery and resistance mechanisms that promote the resilience of ecosystem functions. It is unclear what would be the benefit of narrowing the focus to recovery or engineering resilience.

Mori appears to feel that although there is some consideration of resistance in recent B-EF research (red text in his Box 1), it does not adequately embrace some of the concepts in the ‘ecological resilience’ definition, such as the potential for alternative stable states. We clearly define resilience at the level of an individual function, specifically as “the degree to which the ecosystem function can resist or recover rapidly from environmental perturbations, thereby maintaining function above a socially acceptable level” [3]. This definition does not preclude the existence of alternative stable states of the underlying
system, and, indeed, we include the potential to shift to alternate states that provide lower function delivery as one of several mechanisms underpinning the provision of resilient ecosystem functions. However, there are many other factors that operate at finer scales of biological organisation, such as the species-level (e.g. genetic variability, sensitivity to environmental change, adaptive phenotypic plasticity, Allee effects) and the community-level (e.g. correlation between response and effect traits, functional redundancy, network interaction structure). Most importantly, we feel that a focus on system state (relative to an assumed equilibrium) is not particularly helpful. The ecological resilience literature is somewhat vague with regards to what aspects of the system should be resistant in the face of an environmental perturbation. The relevant response is varying defined as the system ‘state’, the ‘persistence of relationships among state variables within the system’, or the ‘ways of functioning’ [4]. In our review, we promote a definition focusing on functions that are delivered by a system, because biological systems are clearly dynamic, not least because the environment is continually changing. So even a system close to equilibrium would show changes in state, not to mention that many systems of interest (e.g. agro-ecosystems) are far from any equilibrium, or that an equilibrium may not even exist [5]. Therefore, we feel it does not make sense to focus on inconstancy of system state variables, nor their inter-relationships; not least because changes in system state can actually ensure ecosystem functions are maintained (the example we give is that of species turnover in bee communities under climate change, which allow resilient pollination functions). Indeed, the ecological resilience (ER) literature itself highlights the importance of internal system re-organisations as a mechanism of maintaining resilience in the face of perturbations [‘adaptive capacity’; 4]. This clearly involves changes in a system state variables and their inter-relationships. Similarly, in the B-EF literature, as Mori states, the stabilizing effects of
biodiversity on ecosystem functioning are often realized through dynamic processes such as asynchrony and compensation amongst species [6]. So both camps, – the B-EF and ER research fields, seem to be in agreement here: it is not invariance in the system variables which is important, but the maintenance of the ecosystem functions that the system provides.

Although Mori calls for greater synthesis, we suggest that the two research fields of B-EF and ER have already started to converge. Traditionally, B-EF research has certainly adopted a more reductionist (and empirical) approach in contrast to holistic systems thinking of ER. As a consequence, original B-EF studies were conducted in small-scale experiments often focusing on a single function (e.g. plant productivity) and over limited time scales. However, recent research has considered a wider range of ecosystem functions and incorporated study of multiple functions simultaneously [e.g. 7]. Studies have moved from simply considering species richness of assemblages to functional diversity and interactions between species in wider food web networks [8]. Empirical studies have also been conducted over increasingly larger spatial scales [e.g. 9] and across scales [e.g. 7], moving B-EF increasingly in the direction of a broader research framework. Similarly, in the ER research field, key developments have been made from the original abstract theories of systems and simple analogies with real-world examples, to recent progress towards testing and implementation of these theories [e.g. through quantification of early warning systems; 10].

To avoid further confusion, however, reducing ambiguity in the study system context is critical [11]. We propose that many of the apparent discrepancies between the B-EF and ER research fields are simply a result of researchers focusing at different temporal or spatial
scales and talking at cross purposes. We highlight some of these apparent discrepancies and their potential reconciliation in Table 1.

To conclude, both B-EF and ER approaches had initial weaknesses, such as the limited focus of empirical B-EF studies and the limited approach to quantification in more abstract, holistic ER theories. However, researchers in both fields have recognised this and, by increasing the scope of B-EF studies and adopting a more empirical perspective on ER theories, the two fields are now beginning to merge. It is hoped that this emerging synthesis will help in understanding, predicting and delivering solutions for the management of resilient ecosystem functions [12].
Table 1 - Perceived discrepancies in biodiversity-ecosystem function (B-EF) versus ecological resilience (ER) literature and potential reconciliation. To aid researchers a more extensively referenced version of this table is available online (see Online Supplementary Material Table S1).

<table>
<thead>
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<th>Perceived discrepancy</th>
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<td>B-EF literature has traditionally focused primarily on single ecosystem functions in isolation (e.g. plant productivity), whilst ER literature comprises a more holistic view of entire ecosystems (and even socio-ecological systems).</td>
<td>In recent years B-EF research has rapidly expanded beyond single ecosystem functions such as plant productivity to consider a varied range of functions in isolation as well as to consider multi-functionality [e.g. 7]. Similarly, attempts to test and apply the abstract concepts of ER literature have led to examination of specific systems and ecosystem functions.</td>
<td>The two fields of research appear to be converging. To facilitate this bridging, it remains essential for studies to be specific about the characteristics of a system they are measuring, the disturbance regime and the spatial and temporal scale of interest (see main text).</td>
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<td>B-EF literature focusses on stability and equilibrium and ignores the existence of alternate stable states. The existence of alternate stable states is a requisite for ER.</td>
<td>ER definitions concern the likelihood of a system crossing thresholds between alternate stable states (‘regime shifts’). A system need not have high constancy to be resilient- it may be dynamic around a semi-stable equilibrium (i.e. staying within a ‘domain of attraction’). Therefore ER authors have suggested that stability is not a relevant measure of resilience and may even lead to contradictory management outcomes (also see below).</td>
<td>The key point here is whether the focus is on system state variables or ecosystem functions provided by the system. If the focus is the latter, then studies do not rely on quantifying return to some equilibrium state; nor, indeed, do they need to posit the existence of alternate stable states as do ER studies (and some authors have questioned the extent to which these really exist [5]). With a focus on ecosystem functions, any system is suitable for study, even those that are managed far from any stable equilibrium (i.e. most managed ecosystems).</td>
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<td>This issue is often highlighted in the ER literature with a frequently cited example being the management of woodlands to</td>
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Framework) can be detrimental in the longer term. Prevent fires. If fires are regularly suppressed (i.e. to provide stable ecosystem functions from woodlands in the short term), this leads to the accumulation of deadwood, meaning that large fires eventually break out with detrimental effects. In contrast, an ER management perspective (adopting a wider spatial and temporal scale view) would allow frequent smaller fires in parts of the woodland system [4].

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how changes in the composition of communities promote the maintenance of functions provided by a system. Therefore, resilience does not mean the inconstancy of system state variables, and dynamic systems are needed to provide resilient ecosystem functions.
Acknowledgements

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References


Supplementary Material

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Table S1 - Perceived discrepancies in biodiversity-ecosystem function (B-EF) versus ecological resilience (ER) literature and potential reconciliation. This table is a more extensively referenced version of Table 1 in the main text.

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<td>The two fields of research appear to be converging. To facilitate this bridging, it remains essential for studies to be specific about the characteristics of a system they are measuring, the disturbance regime and the spatial and temporal scale of interest [5, 6, 7].</td>
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<td>Woodland system in the longer term is maintained by not continually suppressing fires locally. As highlighted in the main text, clarification on the system type and spatial and temporal scales of interest is critical to avoid researchers talking at cross purposes. Note also, that under a more recent suggestion the focus of management might not be for stability of ecosystem function <em>per se</em>, but just provision consistently above some socially acceptable threshold, although the two are likely to be correlated) [11, 12].</td>
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