

# Spatial and species-level predictions of road mortality risk using trait data

Article

Accepted Version

Gonzalez-Suarez, M. ORCID: https://orcid.org/0000-0001-5069-8900, Zanchetta Ferreira, F. and Grilo, C. (2018) Spatial and species-level predictions of road mortality risk using trait data. Global Ecology and Biogeography, 27 (9). pp. 1093-1105. ISSN 1466-8238 doi: 10.1111/geb.12769 Available at https://centaur.reading.ac.uk/77162/

It is advisable to refer to the publisher's version if you intend to cite from the work. See <u>Guidance on citing</u>.

To link to this article DOI: http://dx.doi.org/10.1111/geb.12769

Publisher: Wiley

All outputs in CentAUR are protected by Intellectual Property Rights law, including copyright law. Copyright and IPR is retained by the creators or other copyright holders. Terms and conditions for use of this material are defined in the <u>End User Agreement</u>.

www.reading.ac.uk/centaur

CentAUR



## Central Archive at the University of Reading

Reading's research outputs online

- 1 Spatial and species-level predictions of road mortality risk using trait data 2 Running title: Trait-based predictions of roadkill risk 3 Manuela González-Suárez<sup>1,3</sup>, Flávio Zanchetta Ferreira<sup>2</sup>, Clara Grilo<sup>2</sup> 4 <sup>1</sup> Ecology and Evolutionary Biology, School of Biological Sciences, University of Reading, 5 6 Reading, RG6 6AS, UK 7 <sup>2</sup> Departamento de Biologia, Setor de Ecologia, Universidade Federal de Lavras, 37200-000, 8 Lavras, MG, Brazil 9 <sup>3</sup>*email: manuela.gonzalez@reading.ac.uk* 10 11 Keywords: bird, Brazil, life-history, machine-learning models, mammal, random forest, road-12 associated mortality, species' traits. 13 14 Acknowledgments 15 This study was part of the project "Road Macroecology: analysis tools to assess impacts on
- 16 biodiversity and landscape structure" funded by CNPq (N° 401171/2014-0). F.Z.F. was
- 17 supported by a CAPES grant (N° 32004010017P3) and CG by CNPq (AJT N° 300021/2015-
- 18 1). We thank Michely Reis Coimbra for helping collecting trait data. Alex Bager, Anthony P.
- 19 Clevenger, Fernando Pinto, Andreas Kindel and two anonymous reviewers provided useful
- 20 comments on earlier versions of this manuscript.
- 21

#### 22 Biosketch

- 23 Manuela González-Suárez is a lecturer at the University of Reading interested in
- 24 understanding how species' traits influence vulnerability to extinction and anthropogenic
- 25 threats from a local to a global scale. She has recently become interested in road ecology
- 26 working in Africa and South America. Flavio Zanchetta completed his MSc in road ecology
- 27 working to understand the factors that affect wildlife-vehicle collision risk in Portugal and
- 28 Brazil. Clara Grilo is currently a researcher at Federal University of Lavras, Brazil. Her
- 29 primary interest is applied ecological research namely examining role of human activities on
- 30 landscape and wildlife. In recent years, her research has focused on road ecology, mainly the
- 31 effects of roads on the relative abundance, behaviour, mortality risk and implications on
- 32 genetic structure and population viability of birds and mammals in Portugal, Spain and Brazil.

#### 33 Abstract

- 34 Aim: Wildlife-vehicle collisions are recognized as one of the major causes of mortality for
- 35 many species. Empirical estimates of road mortality show that some species are more likely to
- 36 be killed than others but to what extend this variation can be explained and predicted using
- 37 intrinsic species characteristics remains poorly understood. This study aims to identify general
- 38 macroecological patterns associated to road mortality and generate spatial and species-level
- 39 predictions of risks.
- 40 Location: Brazil
- 41 *Time period*: 2001-2014
- 42 *Major taxa*: Birds and mammals

43 Methods: We fitted trait-based random forest regression models (controlling for survey

- 44 characteristics) to explain 783 empirical road mortality rates from Brazil, representing 170
- 45 bird and 73 mammalian species. Fitted models were then used to make spatial and species-
- 46 level prediction of road mortality risk in Brazil considering 1775 birds and 623 mammals
- 47 which occur within the country's continental boundaries.
- 48 Results: Survey frequency and geographic location were key predictors of observed rates, but
- 49 mortality was also explained by species' body size, reproductive speed and ecological
- 50 specialization. Spatial predictions revealed high potential standardized (per km road)
- 51 mortality risk in Amazonia for birds and mammals, and additionally high risk in Southern
- 52 Brazil for mammals. Given the existing road network, these predictions mean more than 8
- 53 million birds and 2 million mammals could be killed per year in Brazilian roads. Furthermore,
- 54 predicted rates for all Brazilian endotherm uncovered potential vulnerability to road mortality
- 55 of several understudied species which are currently listed as threatened by the IUCN.
- 56 Conclusion: With a fast-expanding global road network, there is an urgent need to develop
- 57 improved approaches to assess and predict road-related impacts. This study illustrates the
- 58 potential of trait-based models as assessment tools to better understand correlates of
- 59 vulnerability to road mortality across species, and as predictive tools for difficult to sample or
- 60 understudied species and areas.

#### 61 Introduction

62 Roads are increasingly prevalent features in global landscapes (Laurance & Balmford, 2013;

63 Ibisch, Hoffmann, Kreft, Pe'er, Kati, Biber-Freudenberger, ..., & Selva, 2016) leading to

64 growing concerns about their impacts on wildlife (Alamgir, Campbell, loan, Goosem,

65 Clements, Mahmoud, & Laurance, 2017). Wildlife-vehicle collisions are one of the most

66 visible road-related impacts (Coffin, 2007), which can reduce population abundance, limit

67 dispersal, decrease genetic diversity, and ultimately threaten population viability (e.g., Fahrig

68 & Rytwinski, 2009; Borda-de-Água, Grilo, & Pereira, 2014; Grilo C, Del Cerro, Centeno-

69 Cuadros, Ramiro, Román, Molina-Vacas, Fernández-Aguilar, ..., & Godoy, 2016). However,

70 not all species appear to be equally affected by road-associated risks (Fahrig & Rytwinski,

71 2009). Variation among species can reflect methodological issues: small species degrade

72 faster and are less conspicuous, which can reduce their detectability once collision has

73 occurred, leading to underestimated rates (Santos, Carvalho, & Mira, 2011; Teixeira, Coelho,

74 Esperandio, & Kindel, 2013). Differences may also occur due to true variability in collision

risk associated to local abundance, more common species being more likely to suffer high

76 mortality rates (Ford & Fahrig, 2007; Santos, Mira, Salgueiro, Costa, Medinas, & Beja,

77 2016). However, detectability and abundance do not appear to be the only sources of

variation. For example, the lowland tapir *Tapirus terrestris* and the crab-eating fox *Cerdocyon* 

*thous* have similar detectability and observed population densities in the Brazilian Pantanal

80 region (0.4 ind/km<sup>2</sup> - Desbiez, Bodmer, & Tomas, 2010), yet estimated mortality rates vary

81 considerably (fox: 0.24 ind/km/year; tapir: 0.01 ind/km/year. Souza, Cunha, & Markwith,

82 2014). An explanation is that variability in road mortality rates among species is also

83 explained by species' traits related to ecological habits, behaviour, and life-history traits.

84 Species' traits can influence mortality risk via one or more of the processes that lead to

85 collision. First, the probability of encountering roads can be influenced by how the animal

86 moves on the landscape and acquires resources (Grilo, Molina-Vacas, Fernández-Aguilar,

87 Rodríguez, Ramiro, Porto-Peter, & Revilla, 2018). Previous studies have shown higher risks

88 for passerine birds that forage on foliage or bark and inhabit woodlands (Santos et al., 2016),

89 for herbivorous and omnivorous mammals (Barthelmess & Brookes, 2010; Cook &

90 Blumstein, 2013), and for more habitat generalist mammals (Núñez-Regueiro, Branch,

91 Fletcher Jr, Marás, Derlindati, & Tálamo, 2015). Second, the probability of crossing an

92 encountered road may be affected by how the road is perceived and the animal's mobility

(Jaeger, Bowman, Brennan, Fahrig, Bert, Bouchard, ..., Toschanowitz, 2005). For example, 93 94 nocturnal species appear to have higher risk because low traffic volume at night may prevent 95 roads from being perceived as a threat (Grilo, Sousa, Ascensão, Matos, Leitão, Pinheiro, ..., 96 & Revilla 2012). Conversely, species exposed to regular hunting or poaching may be more 97 aware of the human-associated risks often linked to roads, which lead them to avoid roads and 98 thus reduce the risk of collision (Laurance, Croes, Tchignoumba, Lahm, Alonso, Lee, 99 Campbell, & Ondzeano, 2006). Finally, the probability of being hit if crossing can be 100 influenced by the animals' agility, behaviour, and its visibility to drivers (Legagneux & 101 Ducatez, 2013). Collectively these previous results have shown that individual traits can be 102 useful to understand road mortality risk, but to gain a comprehensive understanding and 103 develop valuable predictive tools we need to simultaneously evaluate multiple ecological, 104 behavioural and life-history traits for a diverse group of species.

105 Trait-based models are powerful tools commonly used in macroecological studies to assess the mechanisms underlying the response of species to impacts and predict risks for 106 107 unstudied or difficult-to-detect organisms (González-Suárez, Gómez, & Revilla, 2013; Bland, 108 Collen, Orme, & Bielby, 2015). Here we used machine-learning trait-based models to assess 109 the role of a wide range of species' traits on estimated road mortality rates for bird and 110 mammalian species in Brazil. We focused on birds and mammals because these are well-111 studied groups that provide a suitable empirical dataset (e.g. Coelho, Kindel, & Coelho, 112 2008). Brazil is also an interesting case study because in the last two decades economic and 113 social growth has led to a 20% increase in the road network, increasing impacts on wildlife (DNIT, 2015) and this growth is likely to continue and expand into regions of exceptional 114 115 biological diversity and global ecological importance like the Cerrado and the Amazon 116 (Hoorn, Wesselingh, TerSteege, Bermudez, Mora, Sevink, ..., Antonelli, 2010; Lahsen, 117 Bustamante & Dalla-Nora, 2016). Therefore, there is an urgent need to better understand and 118 predict road risks within Brazil to guide infrastructure planning and mitigation measures, 119 including the protection of those species that are most susceptible to road impacts. To achieve this goal we fitted trait-based models to predict road mortality rates for all Brazilian birds and 120 121 mammals including the many species for which roadkill estimates are not currently available 122 (unstudied or undetected organisms). These predictions allowed us to identify unstudied species with high vulnerability to road-related mortality, and revealed areas where road 123 124 impact is likely high. Our study provides a first comprehensive evaluation of the key intrinsic

- 125 risk factors associated with road impacts in endotherms, and demonstrates the potential of
- 126 using macroecological approaches for road ecology to define predictive trait-based models
- 127 that can identify potentially vulnerable species and high risk areas.
- 128

#### 129 Methods

130 Data

Empirical road mortality rates for birds and mammals were collected from unpublished 131 132 databases (made available by individual researchers contacted via the Lattes platform 133 http://lattes.cnpq.br), grey literature sources (technical reports, proceedings of scientific 134 conferences, MSc and PhD theses), and scientific papers from peer-reviewed journals. 135 Published sources were located using the following keywords in English and their translations to Portuguese: ("roadkills" OR "road mortality") AND ("birds" OR "mammals" OR 136 137 "vertebrates"). We considered only rates from areas of Brazil in which systematic surveys had 138 been conducted at least once a week for a minimum period of three months to minimize the bias on roadkill rates among studies. From each study, road mortality rates per species were 139 140 calculated as the number of individual carcasses detected standardized per kilometre of 141 surveyed road per year (ind/km/year). Rates may be underestimated because not all carcasses 142 persist equally in the pavement. In addition, in some studies not all carcases were identified or reported at the species level. The calculated rates were modified using a correction for carcass 143 persistence estimated by Santos, Carvalho, & Mira (2011) that aims to reduce bias from 144 145 variable detection rates among species (see Appendix S1 in Supporting Information). Both corrected and uncorrected rates were tested to evaluate consistency of results. We present the 146 147 corrected rates in the main text, but results did not change qualitatively when using uncorrected rates (see Appendix S2). We characterized the study location using the 148 149 geographic coordinates of the surveyed road(s) midway point (Appendix S3, Fig. S3.1). 150 Taxonomic names used by each study were matched to the current IUCN taxonomy (IUCN 151 2017). One mammalian species with empirical road mortality data (*Guerlinguetus ingrami*) 152 was not recognized by the IUCN and was not considered in the analyses. Coendou spinosus was listed as two different species (Coendou villosus or Coendou spinosus) in separate studies 153 154 but treated as one species in our analyses. 155 We identified Brazilian birds and mammals based on overlap of distribution range maps

156 (IUCN 2017) with the continental Brazilian territory (based on the IUCN country boundary

157 map). We used distribution range polygons classified as presence "Extant" or "Possible

158 extant" and included all origins and seasonalities. These range polygons were also overlapped

159 with a 1° x 1° grid (equivalent to ~110 x 110 km near the equator) covering the continental

160 Brazilian territory.

We considered 12 species' traits as potentially important to predict the vulnerability of species to road mortality and used published trait databases to find information on all

163 Brazilian birds and mammals (see Table 1 for details).

164

165 Data analysis

166 We modelled empirical road mortality rates using random forest regression trees, a machine

167 learning technique that uses bootstrapped data samples to generate multiple regression trees

168 from which the importance of the predictors is defined (Breiman, 2001). Regression trees

169 have high predictive accuracy and the capacity to deal with complexity in relationships

170 including non-linearities and interactions (Cutler, Edwards, Beard, Cutler, Hess, Gibson, &

171 Lawler, 2007). Regression trees have also been show to offer comparable results to

172 phylogenetic methods that explicitly account for the fact that related species may share similar

traits due to shared evolutionary history (Bielby et al., 2009). While incorporating

174 phylogenetic information into regression trees is not possible, to capture the potential

175 importance of phylogeny we included taxonomic order (following the classification of the

176 IUCN, 2017) as a predictor in our models.

177 Rates from birds and mammals were analysed separately to reflect the intrinsic 178 differences between these groups. For each imputed version (15 per taxonomic group) we 179 fitted a random forest model with 2000 trees using the randomForest procedure from the 180 randomForest library (Liaw & Wiener, 2002) in R version 3.4.1 (R Core Team 2017). Model 181 predictors included traits (Table 1), taxonomic order, and the three survey related predictors 182 for each study: survey interval (time between surveys in days), and latitude and longitude of 183 the surveyed road(s) midway point.

Trait data were not available for all species (data limitations are common in
comparative studies, see González-Suárez, Lucas, & Revilla, 2012). To avoid excluding
observations, which can lead to biases, we estimated missing values for each taxonomic group
using nonparametric imputation based on random forest regression trees (Stekhoven &
Bühlmann, 2012). Since this imputation approach results in slightly different values each time

189 it is run, we imputed and analysed 15 datasets for each taxonomic group to capture

190 uncertainty in the imputation process and assess sensibility of results to that uncertainty. To

191 facilitate reproducibility and encourage open science the code (R script) and data used in this

192 study will be available at https://dx.doi.org/ 10.6084/m9.figshare.6237608 on 9 November

193 2018.

We assessed overall model performance using the total variance explained. We also calculated the importance of each variable by permuting all observed values within each variable across observations and evaluating the effect on model performance (changes in variance explained). The permutation of important variables decreases significantly the model performance whereas the permutation of less important variables should have little effect on the model performance.

200 Fitted models were subsequently used to generate predicted values for each Brazilian 201 species in each of its occupied 1° x 1° grid cells, representing a hypothetical systematic survey across all Brazil (cells treated as studies sites). Geographic coordinates were defined as those 202 203 of the grid cell centroid. The survey interval was set to the observed median value (3.5 days 204 between surveys for both birds and mammals). Different survey intervals did not qualitatively 205 affect results (Appendix 3 figure S3.2). We used the median prediction from the 15 models 206 based on the different imputed datasets for each cell and species combination. Predicted risk 207 for each species (species-level predictions) was then calculated as the median of the predicted 208 rates over all its occupied grid cells; thus, reflecting intrinsic vulnerabilities and spatial risks 209 within the species' distribution range. Spatial predictions were generated by aggregating 210 predicted values for each grid cell, thus, summarizing risk associated to geographic location 211 and to the intrinsic vulnerabilities of co-occurring species. We calculated a standardized risk based on ind/km/year, and a predicted total risk (ind/year) obtained by multiplying the 212 213 standardized risk by the total kilometres of paved road within each grid cell. Road network 214 data were obtained from the Openstreetmap available at https://www.openstreetmap.org. As 215 there is limited knowledge on the responses of the different species to roads, our approach 216 assumes no specific road avoidance.

217

#### 218 **Results**

We located 38 studies that reported road mortality rates in different areas of Brazil and met our criteria of minimum frequency and period of survey (Figure S3.1). From these studies we obtained 417 mortality rates for 170 bird species, and 366 road mortality rates for 73 mammal species. Based on current distribution range maps we identified 1831 birds and 623 mammals as present in Brazil. We made predictions for all mammals, but had to exclude 56 bird species classified in taxonomic orders not represented in our empirical dataset (predictions cannot be made for new levels in a categorical variable). All data are available on (to be deposited on

online repository upon acceptance, and made available as supplementary files for review).

Some species were reported by multiple studies, for example the smooth-billed ani *Crotophaga ani* was the most frequently detected bird (16 studies), while the crab-eating fox
was the most frequently reported mammal (32 studies). However, many species were only
observed in one study (90 bird species and 26 mammals). Observed (corrected) road mortality
rates ranged from 0.001 to 7.61 ind/km/year for birds (highest rate was observed for *Crotophaga ani*). For mammals rates ranged from 0.0007 to 18.52 ind/km/year (highest rate
was reported for the capybara *Hydrochoerus hydrochaeris*).

234 Fitted models explained 61.7% of the observed variance in road mortality rates for birds (median value, range across imputed datasets 61.1-62.4%) and 51.8% for mammals (range 235 51.2-52.4%). As expected survey predictors were important in both groups (Fig. 1). In 236 237 particular, survey interval was the most important predictor with intervals of 1-2 day 238 generally associated with higher estimates both for birds and mammals (Figs. 2 and 3). 239 Geographic location also explained observed road mortality rates (Fig. 1) with partial 240 dependence plots showing higher rates associated to western locations for both birds and 241 mammals, but contrasting patterns for latitude, with locations closer to the equator associated 242 with higher risk in birds but with lower risk in mammals (Figs. 2 and 3). Taxonomic order 243 was an important predictor for birds with higher rates found among cuckoos and anis 244 (Cuculiformes, eight species) and flightless birds (Cariamiformes, one species; 245 Struthioniformes, four species). Order was also important for mammals, with higher rates 246 among anteaters and sloths (Pilosa, 6 species) and armadillos (Cingulata, six species). 247 The analyses also revealed several species' traits as predictors of road mortality rates in both birds and mammals with high consistency among models based on the different imputed 248 249 datasets (Fig. 1). Partial dependence plots show higher road mortality rates in birds associated 250 to larger body mass (>2 kg), earlier maturity ages, shorter lifespans, ground foraging

behaviour, and habitat and diet generalism (Fig. 2). For mammals, higher rates were

associated with scavenging behaviour, early maturity, smaller home range sizes, intermediate
body masses (3-50 kg), and habitat generalism (Fig. 3).

Model predictions matched empirical data well for mammals, but showed a tendency to overestimate low values for birds (Fig. S3.3). Both observed and predicted rates showed considerable within-species variation. Overall, median observed and predicted road mortality rates per species were lower among species with more threatened conservation status as defined by the global IUCN Red List status (Fig. S3.4).

259 Median predicted rates for all Brazilian species ranged from 0.02 to 0.37 ind/km/year for birds and from 0.02 to 1.08 ind/km/year for mammals. Among species without empirical 260 261 data, the blue-black grassquit Volatinia jacarina and Reig's grass mouse Akodon reigi had the 262 highest predicted rates (0.37 and 1.08 respectively, data available at – online repository). High 263 rates were also predicted for several unstudied species of conservation concern, including the 264 white-lined antbird Myrmoborus lophotes and the rufous-fronted antthrush Formicarius *rufifrons*, both birds listed as Near Threatened by the IUCN Red List (predicted median rates 265 of 0.30 and 0.28 ind/km/year, respectively), and the greater Wilfred's mouse Wilfredomys 266 oenax (Endangered status, predicted median rates of 0.93 ind/km/year) and the Lami tuco-267 268 tuco Ctenomys lami (Vulnerable status, predicted median rates of 0.44 ind/km/year).

269 The map of standardized predicted spatial risk (the sum of all predicted mortalities per 270 kilometre of road per year for species co-occurring in a given grid cell) showed high expected 271 roadkill rates in the western Amazon region for both birds and mammals, and also high 272 mammalian rates in the Pampas region (southern Brazil), with the lowest values generally 273 found in eastern Brazil (Fig. 4a, b). These patterns were consistent when representing median 274 standardized predicted risk, the rates expected for an average species in each cell per 275 kilometre of road (Fig. S3.5). Areas with high predicted risk generally also showed high 276 variability among species (high standard deviation. Fig S3.5), reflecting differences in 277 intrinsic vulnerabilities of the local fauna. These spatial standardized risk patterns did not 278 solely reflect species richness. For example, the highest road mortality rates for mammals 279 were predicted in southern Brazil where mammalian richness is lowest (Fig. 4e, f).

When considering the existing road network to estimate total mortality (number of mortalities per year in each grid cell, Fig. 4c, d), the patterns, as expected, changed, and revealed higher total predicted risk in coastal areas where most roads are found (Fig. S3.6). However, total mortality did not exclusively reflect existing roads. For example, risk was high

284 in Western areas (Mato Grosso and Rondonia) where there are relatively few roads, and 285 relatively low in the Northeast region where road density is high. Worryingly, in some areas 286 total risk was very high with expected rates of over 96,000 individual birds and over 53,000 287 individual mammals killed per year in some 1° x 1° areas. Adding all predicted rates across 288 the country our results suggest that as many as 8,351,120 birds and 2,225,101 mammals could 289 be killed per year in Brazilian roads. Considering potential impacts of additional planned 290 paved roads (Fig. 3.7) we find increased risk in the Amazon, which we identify as a sensitive 291 area (with high standardized and median risk rates).

Mapping predicted rates only for threatened species we found a slightly different spatial pattern with highest risk found in the eastern Amazon (Fig S3.8). The predicted total risk suggests as many as 179,601 threatened birds and 73,031 threatened mammals could be killed per year in Brazilian roads.

296

#### 297 Discussion

298 Our results provide evidence that road-associated mortality risk is not randomly distributed 299 among species and can be partly explained by ecological, behavioural and life-history traits in 300 avian and mammalian species. Previous studies explored the role of individual traits (e.g. 301 Ford & Fahrig, 2007; Cook & Blumstein, 2013), but our results reveal that road mortality is 302 associated to a combination of multiple predictors that reflect diverse characteristics of the 303 studied species. Although analyses were conducted separately, we found that several traits 304 explained vulnerability to traffic for both bird and mammal species suggesting generalized 305 intrinsic sources of risk for endotherms, for example associated with body size. Our study also 306 identified distinct sources of vulnerability, which highlight idiosyncrasies of the studied 307 species and/or distinct mechanisms associated with vulnerability for birds and mammals. In 308 addition, road mortality also varied among taxonomic groups, potentially reflecting untested 309 characteristics. For example, for mammals we found higher rates among relatively slow 310 moving species like armadillos and sloths, while for birds flightless groups exhibited higher 311 rates. As expected survey-related variables were also important predictors of risk. Although 312 we used a correction factor, survey interval remained a key predictor for both groups. Our 313 results suggest that 1-2 days intervals between surveys may be optimal to assess risks for 314 these species (those intervals were associated with the higher rates - see also Santos et al., 315 2011). Survey location was also important for both birds and mammals, with higher predicted

risk towards the west in both groups but contrasting latitudinal gradients that identifieddistinct risk regions within Brazil for birds and mammals.

318 We observed that species with weights above 2-3 kg had higher risk of being roadkilled, 319 although for mammals the risk decreased again for species above  $\sim 50$  kg. Generally, larger 320 species tend to be more mobile (Sutherland, Harestad, Price, & Lertzman, 2000), which 321 increases the probability of encountering and crossing roads. However, for the largest 322 mammals, collision risk may be reduced due to earlier detection by drivers, which provides 323 more time for response, and also due to more active avoidance responses by drivers seeking to 324 prevent potentially dangerous collisions. Lower risk among smaller species could be partly 325 explained by variation in detectability during surveys (methodological limitations). Smaller 326 species are more difficult to see, and degrade faster, and this could result in potentially 327 underestimated mortality rates. However, we actually found some increases in risk predicted 328 for some small species, suggesting detectability during surveys does not fully explain this 329 pattern. A previous study also reported a trend for higher mortality rates in smaller birds based on species <1.2 kg (Santos et al., 2016). 330

331 Our findings also suggest generalists (those with wider habitat and diet breadths) have 332 higher mortality rates than specialists (although for mammals diet specialization was not 333 clearly related to road mortality risk). However, contrary to previous studies, we did not find 334 an effect of trophic level (Ford & Fahrig, 2007; Cook & Blumstein, 2013). It is possible this 335 difference occurs because those previous studies did not consider diet specialization. By 336 exploring both variables we show that the degree of specialization may be more informative to predict roadkill likelihood than trophic level, at least for birds. In general, specialist species 337 338 may be less likely to approach and cross roads if these are perceived as unsuitable 339 environments. Previous studies have shown that species that are reluctant to cross open 340 grounds avoid crossing roads due to low availability of cover, and therefore have lower 341 likelihood of being roadkilled (Develey & Stouffer, 2001; Rytwinski & Fahrig, 2012). 342 Moreover, the high availability of resources and refuges in road verges can attract habitat and 343 diet generalist species to roads and increase their risk of being hit by passing vehicles (Ruiz-344 Capillas, Mata, & Malo, 2012; Barrientos & Bolonio, 2009). Among the resources that may 345 be available near roads are roadkill carcasses, which attract scavengers, potentially increasing 346 their collision risk. Although we found scavenging behaviour of mammals considerably 347 increased mortality rates, there was no clear association for birds. A previous study by Cook

349 contrasting results may be due to individual responses to traffic and risk perception (Jacobson, 350 Bliss-Ketchum, De Rivera, & Smith, 2016) and also be influenced by road-related features 351 including sinuosity and traffic intensity (Grilo, Ascensão, Santos-Reis, & Bissonette, 2011). 352 Population-level processes can also influence risk with locally abundant species being 353 more likely to have higher mortality rates. Although we did not have abundance estimates for the studied areas, higher local abundance is often associated with faster reproductive rates 354 355 (e.g., earlier maturity ages) which we found were associated with higher risk of collision in both birds and mammals. However, while consistent with predicted patterns, we note that 356 357 many estimates of maturity age were imputed, particularly for bird species, thus, support 358 should be interpreted with caution. Future studies would benefit from conducting both 359 roadkill and abundance surveys in the same areas to better understand road-associated risks. 360 Besides contributing to our understanding of the general drivers of road-associated 361 mortality, our analyses show trait-based models can be used as predictive tools for conservation assessment and management of road-related impacts. Our model predicted high 362 363 roadkill rates for several species of conservation concern for which empirical estimates are 364 not currently available. Empirical estimates may be unavailable for different reasons: 1) species only inhabit poorly sampled areas (e.g., Amazon); 2) species occur in sampled areas 365 but avoid modified habitats including roads; and 3) species are not detected (e.g., small size 366 367 bias) or are not correctly identified during surveys due to taxonomic uncertainty or cryptic 368 characteristics. A recommendation that emerges from our study is the need for targeted survey efforts for species identified as potentially susceptible here so their true risk can be quantified 369 370 and if needed adequate management and mitigation actions can be implemented.

& Blumstein (2013) reported no effect of scavenging behaviour for either group. These

348

371 Our results also generate spatial predictions of road mortality risks, which highlight 372 the apparent high vulnerability of Amazonian species (higher risk per kilometre of road). 373 Although mortality rates in this area are likely to be relatively small due to low road density, 374 this region has high biodiversity and our analyses suggest road infrastructure development 375 could have severe impacts on many species. Considering the existing road network, as expected, we found higher total risk in areas with higher road density. However, median risk 376 377 per species was not always high in areas, like Southern Brazil, with many threatened species and high road densities, perhaps reflecting former population depression due to road mortality 378 379 (Teixeira, Kindel, Hartz, Mitchell, & Fahrig, 2017). Worryingly, our results suggest that more

380 than 8 million individual birds (nearly 180,000 from species threatened by extinction) and 381 more than 2 million mammals (over 72,000 from threatened species) may be killed each year 382 in existing Brazilian roads. These high values are predicted despite the fact that Brazil has a 383 relatively high number of roadless areas (Ibisch, Hoffmann, Kreft, Pe'er, Kati, Biber-384 Freudenberger, ..., & Selva, 2016). Furthermore, our rates may be underestimated because we 385 do not consider unpaved roads. We also do not account for other indirect road impacts, such 386 as changes in spatial distribution due to road avoidance (Torres, Jaeger & Alonso, 2016), 387 which can reduce roadkill but may in turn impact population viability in other ways. Future 388 road development is also likely to result in increased risk, particularly in areas we identified 389 as sensitive (with intrinsically vulnerable species), like the Amazon. These results provide a 390 first overview of risk revealing potentially vulnerable species and areas. However, localized 391 and refined spatial predictions (accounting for unpaved roads and traffic levels) would be 392 useful to further describe areas in which road development is likely to have widespread 393 impacts on the local fauna, as well as areas in which mitigation measured would be most 394 valuable.

395 While our analyses offer useful insights, there are also limitations of the available data. 396 First, missing trait data is a common problem in comparative studies (González-Suárez et al., 397 2012). Data imputation methods may be helpful but some limitations (e.g. handling variable 398 correlation) need consideration (Penone, Davidson, Shoemaker, Marco, Rondinini, 399 Brooks, ..., Costa, 2014). Also imputation of large proportions of the dataset incorporate 400 uncertainty in results, this can be quantified as done here, but results should still be interpreted 401 with caution. In addition, to missing trait data, empirical estimates of road mortality were also 402 not available for all species or areas. Instead, these data reflect a non-random subset of species and locations, associated to biases in research effort and methodological issues. In fact, these 403 404 biases may explain the relatively abrupt geographic changes in our spatial predictions, 405 particularly for birds. Variation in detectability during surveys is a recurrent problem in road 406 surveys. Carcasses from small species may disappear from roads in less than 24 hours 407 (Teixeira et al., 2013). Carcasses in hot, humid areas, and those with high traffic levels may 408 also disappear faster. We applied corrections to the observed road mortality rates, but our 409 results were qualitatively the same when using uncorrected rates (Appendix 1) contrary to 410 what Santos et al. (2011) found in their study. The criteria we adopted to only include studies 411 with a 7-day minimum survey interval may have contributed to reduce differences; intervals

of one week or shorter have been found to significantly reduce bias for medium- and largesized birds and mammals (Bager & Rosa, 2011). Ultimately, data limitations and biases can
only be effectively addressed with additional research efforts. In the meantime, studies using
available data should consider uncertainty in results, explore correction methods, and interpret

416 findings with caution.

417 Overall, our results contribute to a better understanding of the biological drivers that 418 make species vulnerable to road traffic collisions. Previous studies have largely focused on 419 the importance of road characteristics (e.g. traffic, size, and design) and landscape features (e.g. vegetation type, and degree of fragmentation; Saeki & Macdonald, 2004; Grilo, 420 421 Bissonette, & Santos-Reis, 2009). However, our study shows that using available knowledge 422 on species traits and macroecological approaches can contribute to better evaluate risks and 423 offer insights into species and spatial level risks. Obtained predictions can guide future survey 424 efforts, pointing to poorly sampled areas with potentially susceptible species, and may also be 425 used to plan conservation strategies, road development, and mitigation measures. These predictions also offer insights into the magnitude of the threat imposed by roads, with 426 427 potentially millions of individual birds and mammals being killed each year within one 428 country.

429

#### 430 Literature Cited

- 431 Alamgir, M., Campbell, M. J., Sloan, S., Goosem, M., Clements, G., Mahmoud, M. I., &
- 432 Laurance, W.F. (2017). Economic, socio-political and environmental risks of road
- 433 development in the tropics. *Current Biology*, **27**, R1130-R1140.
- 434
- 435 Bager, A., & Rosa, C. A. (2011). Influence of sampling effort on the estimated richness of
- 436 roadkilled vertebrate wildlife. *Environmental Management*, **47**, 851-858.
- 437
- 438 Barrientos, R., & Bolonio, L. (2009). The presence of rabbits adjacent to roads increases
- 439 polecat road mortality. *Biodiversity and Conservation*, **18**, 405-418.
- 440
- 441 Barthelmess, E. L., & Brooks M. S. (2010). The influence of body-size and diet on roadkill
- trends in mammals. *Biodiversity and Conservation*, **19**,1611-1629.
- 443

- 444 Bielby, J., Cardillo, M., Cooper, N., & Purvis, A. (2010). Modelling extinction risk in
- 445 multispecies data sets: phylogenetically independent contrasts versus decision trees.
- 446 *Biodiversity and Conservation*, **19**, 113-127.
- 447
- 448 Bland, L. M., Collen, B., Orme, C. D. L. & Bielby, J. (2015), Predicting the conservation
- 449 status of data-deficient species. *Conservation Biology*, **29**, 250–259.
- 450
- 451 Borda-de-Água L., Grilo C., & Pereira, H. M. (2014). Modeling the impact of road mortality
- 452 on barn owl (*Tyto alba*) populations using age-structured models. *Ecological Modelling*, 276,
  453 29-37.
- 454
- 455 Breiman, L. (2001). Random Forests. *Machine Learning* **45**, 5-32.
- 456
- 457 Coelho, I. P., Kindel, A., & Coelho, A. V. P. (2008). Roadkills of vertebrate species on two
- 458 highways through the Atlantic Forest Biosphere Reserve, southern Brazil. *European Journal*459 of Wildlife Research, 54, 689-699.
- 460
- 461 Coffin, AW. (2007). From roadkill to road ecology: A review of the ecological effects of
- 462 roads. Journal of Transport Geography, 15, 396-406.
- 463
- 464 Cook, T. C., & Blumsteim, D. T. (2013). The omnivore's dilema: diet explains variation in
- 465 vulnerability to vehicle collision mortality. *Biological Conservation*, **167**, 310-315.
- 466
- 467 Cutler, D. R., Edwards, T. C., Beard, K. H., Cutler, A., Hess, K. T., Gibson, J. & Lawler, J. J.
- 468 (2007). Random forest for classification in ecology. *Ecology*, **88**: 2783–2792.
- 469
- 470 Desbiez, A. L. J., Bodmer, R. E., & Tomas, W. M. (2010). Mamalian densities in a
- 471 neotropical wetland subject to extreme climatic events. *Biotropica*, **42**, 373-378.
- 472
- 473 Develey, P. F., & Stouffer, P. C. (2001). Effects of roads on movements by understory birds
- 474 in mixed-species flocks in Central Amazonian Brazil. *Conservation Biology*, **15**, 1416-1422.

- 476 Dinerstein, E., Olson, D., Joshi, A., Vynne, C., Burgess, N. D., Wikramanayake, E., ...
- 477 Saleem, M. (2017). An ecoregion-based approach to protecting half the terrestrial
- 478 realm. *Bioscience*, **67**, 534–545.
- 479
- 480 DNIT (Departamento Nacional de Infraestrutura de Transportes) (2015). Sistema nacional de
  481 viação. Brasília, DF, Brasil.
- 482
- 483 Fahrig, L., & Rytwinski, T. (2009). Effects of roads on animal abundance: an empirical
- 484 review and synthesis. *Ecology and Society*, **14**, 21.
- 485
- 486 Ford, A. T., & Fahrig, L. (2007). Diet and body size of North American mammal road
- 487 mortalities. *Transportation Research Part D*, **12**, 498-503.
- 488
- 489 González-Suárez, M., Lucas, P. M., & Revilla, E. (2012). Biases in comparative analyses of
- 490 extinction risk: mind the gap. *Journal of Animal Ecology*, **81**, 1211-1222.
- 491
- 492 González-Suárez, M., Gómez, A., & Revilla, E. (2013). Which intrinsic traits predict
- vulnerability to extinction depends on the actual threatening processes. *Ecosphere*, 4, 76.
- 495 Grilo, C., Bissonette, J. A., & Santos-Reis, M. (2009). Spatial-temporal patterns in
- 496 Mediterranean carnivore road casualties: Consequences for mitigation. *Biological*
- 497 *Conservation*, **142**, 301-313.
- 498
- 499 Grilo, C., Ascensão, F., Santos-Reis, M., & Bissonette, J. A. (2011). Do well-connected
- landscapes promote road-related mortality? *European Journal of Wildlife Research*, 57, 707716.
- 502
- 503 Grilo, C., Sousa J., Ascensão, F., Matos, H., Leitão, I., Pinheiro, P., ... & Revilla, E. (2012).
- Individual spatial responses towards roads: implications for road mortality risk. *PLoS ONE* 7,9.

- 507 Grilo, C., Del Cerro, I., Centeno-Cuadros, A., Ramiro, V., Román, J., Molina-Vacas,
- 508 G., ... Godoy, J. (2016). Heterogeneous road networks have no apparent effect on the genetic
- 509 structure of small mammal populations. *Science of the Total Environment*, **565**, 706-713.
- 510
- 511 Grilo, C., Molina-Vacas, G., Fernández-Aguilar, X., Rodríguez, J., Ramiro, V., Porto-Peter,
- 512 F.,... Revilla, E. (2018). Species-specific movement traits and specialization determine the
- 513 spatial responses of small mammals towards roads. *Landscape and Urban Planning*, **169**,
- 514 199-207.
- 515
- 516 Hoorn, C., Wesselingh, F. P., Ter Steege, H., Bermudez, M. A., Mora, A., Sevink, J., ...
- 517 Antonelli, A. (2010). Amazonia through time: Andean uplift, climate change, landscape
- 518 evolution and biodiversity. *Science*, **330**, 927-931.
- 519
- 520 Ibisch, P. L., Hoffmann, M. T., Kreft, S., Pe'er, G., Kati, V., Biber-Freudenberger, L., ... &
- Selva, N. (2016). A global map of roadless areas and their conservation status. *Science*, **354**,
  1423-1427.
- 523
- 524 IUCN (2017). The IUCN Red List of threatened species. <a href="http://www.iucnredlist.org">http://www.iucnredlist.org</a>>.
- 525 Downloaded on 11 October 2017.
- 526
- 527 Jacobson, S. L., Bliss-Ketchum, L. L., De Rivera, C. E., & Smith, W. P. (2016). A behavior-
- 528 based framework for assessing barrier effects to wildlife from vehicle traffic
- 529 volume. *Ecosphere*, **7**, 1-15.
- 530
- 531 Jaeger, J. A. G., Bowman, J., Brennan, J., Fahrig, L., Bert, D., Bouchard., J., ... Toschanowitz
- 532 K. T. (2005). Predicting when animal populations are at risk from roads: an interactive model
- 533 of road avoidance behaviour. *Ecological Modelling*, **185**, 2, 329-348.
- 534
- 535 Jones, K. E., Jones, K. E., Bielby, J., Cardillo, M., Fritz, S. A., O'Dell, J., ...Purvis, A. (2009).
- 536 PanTHERIA: a species-level database of life history, ecology, and geography of extant and
- 537 recently extinct mammals. *Ecology*, **90**, 2648-2648.
- 538

- 539 Lahsen, M., Bustamante, M. Mc., & Dalla-Nora, E. L. (2016). Undervaluing and
- 540 overexploiting the Brazilian Cerrado at our peril. *Environment: Science and Policy for*
- 541 Sustainable Development, **58**, 6.
- 542
- 543 Laurance, W. F., Croes, B. M., Tchignoumba, L., Lahm, S. A., Alonso, A., Lee, M. E.,
- 544 Campbell, P., & Ondzeano, C. (2006). Impacts of roads and hunting on central African
- 545 rainforest mammals. *Conservation Biology*, **20**, 1251-1261.
- 546
- Laurance, W. F., & Balmford A. (2013). Land use: a global map for road building. *Nature*,
  495, 308-309.
- 549
- Legagneux, P., & Ducatez, S. (2013). European birds adjust their flight initiation distance to
  road speed limits. *Biology Letters*, 9(5), 20130417.
- 552
- Liaw, A., & Wiener, M. (2012). Classification and regression by random forest. *R News*, 2,
  18-22.
- 555
- 556 Lindsted, S. L., Miller, B. J., & Buskirk, S. W. (1986). Home range, time and body size in
- 557 mammals. *Ecology*, **67**, 413-418.
- 558
- 559 Myhrvold, N. P., Baldridge, E., Chan, B., Sivam, D., Freeman, D. L., & Ernest, S. K. M.
- 560 (2015). An amniote life-history database to perform comparative analyses with birds,
- 561 mammals, and reptiles. *Ecological Archives*, **96**, 3109-3109.
- 562
- 563 Núñez-Regueiro, M. M., Branch, L., Fletcher, Jr R. J., Marás, G. A., Derlindati, E., &
- 564 Tálamo, A. (2015). Spatial patterns of mammal occurrence in forest strips surrounded by
- agricultural crops of the Chaco region, Argentina. *Biological Conservation*, **187**,19-26.
- 566 Paglia, A.P., Fonseca, G. A. B., Rylands, A. B., Herrmann, G., Aguiar, L. M. S., Chiarello, A.
- 567 G., ..., Patton, J. L. (2012). Annotated checklist of Brazilian mammals 2<sup>nd</sup> Edition. Occasional
- 568 Papers in Conservation Biology, 6, 1-76.
- 569

570	Penone, C., Davidson, A. D., Shoemaker, K. T., Marco, M. D., Rondinini, C., Brooks, T. M.,,
571	& Costa, G. C. (2014). Imputation of missing data in life-history traits dataset: which approach
572	performs the best? Methods in Ecology and Evolution, 5, 961-970.
573	
574	R Core Team (2017). R: A language and environment for statistical computing. R Foundation
575	for Statistical Computing, Vienna, Austria. URL https://www.R-project.org/.
576 577	Ruiz-Capillas, P., Mata, C., & Malo, J. E. (2012). Road verges are refuges for small mammal
578	populations in extensively managed Mediterranean landscapes. <i>Biological Conservation</i> , <b>158</b> ,
579	223-229.
580	223-229.
581	Rytwinski, T., & Fahrig, L. (2012). Do species life history traits explain population responses
582	to roads? A meta-analysis. Biological Conservation, 147, 87-98.
583	
584	Saeki, M., & Macdonald, D. W. (2004). The effects of traffic on the raccoon dog (Nyctereutes
585	procyonides viverrinus) and other mammals in Japan. Biological Conservation, 118, 559-571.
586	
587	Santos, S. M., Carvalho, F., & Mira, A. (2011). How long do the dead survive on the road?
588	Carcass persistence probability and implications for roadkill monitoring surveys. PLoS ONE,
589	<b>6</b> , 9.
590	
591	Santos, S. M., Mira, A., Salgueiro, P. A., Costa, P., Medinas, D., & Beja, P. (2016). Avian
592	trait-mediated vulnerability to road traffic collisions. Biological Conservation, 200, 122-130.
593	
594	Souza, J. C., Cunha, V. P., & Markwith, S. H. (2014). Spatiotemporal variation in human-
595	wildlife conflicts along highway BR-262 in the Brazilian Pantanal. Wetlands Ecology and
596	Management, 23, 227-239.
597	
598	Stekhoven, D. J., & Bühlmann, P. (2012). MissForest-non-parametric missing value
599	imputation for mixed-type data. Bioinformatics, 28, 112-118.
600	

- 601 Sutherland, G. D., Harestad A. S., Price K., & Lertzman K. P. (2000). Scaling of natal
- dispersal distances in terrestrial birds and mammals. *Conservation Ecology*, **4**, 16.
- 604 Teixeira, F. Z., Coelho, A. V. P., Esperandio, I. B., & Kindel, A. (2013). Vertebrate road
- 605 mortality estimates: Effects of sampling methods and carcass removal. *Biological*
- 606 *Conservation*, **157**, 317-323.
- 607
- Teixeira, F. Z., Kindel, A., Hartz, S. M., Mitchell, S., & Fahrig, L. (2017). When roadkill
- hotspots do not indicate the best sites for roadkill mitigation. *Journal of Applied Ecology*, 54,
  1544-1551.
- 611
- 612 Torres, A., Jaeger, J. A., & Alonso, J. C. (2016). Assessing large-scale wildlife responses to
- 613 human infrastructure development. Proceedings of the National Academy of Sciences of the
- 614 United States of America, **113**, 8472-8477.
- 615
- 616 Wilman, H., Belmaker, J., Simpson, J., de la Rosa, C., Rivadeneira, M. M., & Jetz, W. (2014).
- 617 EltonTraits 1.0: Species-level foraging attributes of the world's birds and mammals.
- 618 *Ecological Archives*, **95**, 2027-2027.
- 619

#### 620 Data accessibility

- 621 The R scripts and full database including available empirical road mortality rates, trait
- 622 descriptors, and predicted rates for all Brazilian birds and mammals will be available at
- 623 https://dx.doi.org/ 10.6084/m9.figshare.6237608 on 9 November 2018.

Table 1. Definitions, hypothesis, data sources and sample size availability for the species traits considered as predictors of road mortality in

625 Brazilian birds and mammals in this study. Total species considered for birds data N<sub>all</sub>=1831, birds with empirical roadkill rates N<sub>RK</sub>=170, total

626 mammalian species  $N_{all}$ =623, and mammals with empirical roadkill rates  $N_{RK}$ =73. Home range and sociality were only considered for

627 mammals.

Trait	Definition and hypothesis	Data source	Birds		Mammals	
			N <sub>all</sub>	N <sub>RK</sub>	N <sub>all</sub>	N <sub>RK</sub>
Diet breadth	Total number of 10 possible dietary categories consumed by each	Wilman et al., 2014	1527	153	454	70
	species.					
	Diet generalist are more likely to use resources on road verges and	1				
	thus, approach roads leading to higher roadkill rates.					
Scavenger	Binary descriptor to identify species with diet consisting of $\geq 10\%$	Wilman et al., 2014	1700	168	585	72
behaviour	carrion.					
	Scavengers can forage on roadkill, and thus will spend time on					
	and near roads leading to higher roadkill rates.					
Trophic level	Trophic level described as: Herbivore (>80% plant diet),	Wilman et al., 2014	1700	168	585	72
	Carnivore (>80% animal diet), or Omnivore (<80% animal or					
	plant diet). We used 80% thresholds as some diet data may					
	include accidentally ingested material.					

	Herbivores are more likely to utilize road verges, where					
	vegetation can benefit from run-off, leading to higher roadkill					
	rates.					
Habitat	Total number of ecoregions within Brazil overlapping the current	Dinerstein et al., 2017	1831	170	544	70
breadth	geographic range area of each species.					
	Habitat specialists are likely to avoid novel environments like					
	roads, leading to reduced roadkill rates.					
Ground	Prevalence of foraging on ground estimated as percentage of use	Wilman et al. 2014	1646	167	623	73
foraging	of that substrate for birds. For mammals a categorical variable					
	describing species classified as "ground foragers".					
	Species that forage in arboreal or aerial environments are less					
	likely to be on roads, leading to reduced roadkill rates.					
Activity cyc	e The main period a species is active defined as: nocturnal, diurnal	Jones et al., 2009;	1700	168	585	72
	or other.	Wilman et al. 2014				
	Nocturnal species are active at times with less traffic (which can					
	reduce their perception of risk) and also when visibility is limited					
	for drivers (affecting collision avoidance behaviours), both					
	mechanisms can lead to higher roadkill rates.					
	for drivers (affecting collision avoidance behaviours), both					

Exploitation	Binary descriptor to identify species classified as at risk from	IUCN Threats	1643	166	617	73
	direct exploitation based on the IUCN (species classified as	Classification Schemes,				
	affected by categories 5.1 and/or 5.4).	category 5.1 (Version				
	Hunted species are more likely to perceive roads as risks and	3.2).				
	avoid them, leading to reduced roadkill rates.					
Body mass	Average body mass in grams of an adult individual. Median	Jones et al., 2009, Paglia	1664	170	546	72
	values per species based on all available sources.	et al., 2012, Wilman et al.				
	Large species are generally more mobile which could increase	2014,				
	roadkill rates.					
Home range	Average home range size in km <sup>2</sup> .	Jones et al. 2009	NA	NA	85	30
	Species with wider home ranges are more likely to overlap with					
	roads, leading to increased roadkill rates.					
Lifespan	The maximum recorded age of an individual of the species in	Jones et al., 2009;	230	36	254	68
	days. Median values per species based on all available sources.	Myhrvold et al., 2015				
	Species with longer lifespans may be more likely to learn about					
	road risks, leading to reduced roadkill rates.					
Maturity age	Average age in years at which individuals reach sexual maturity.	Jones et al., 2009;	177	30	197	50
	Median values per species based on all available sources.	Myhrvold et al., 2015				

	Species with delayed maturation are more likely to prioritize survival, which can lead to greater perception of risk from roads					
	and reduced roadkill rates.					
Sociality	Binary descriptor to identify species in which individual spend	Jones et al., 2009	NA	NA	136	33
	most of their lives in a group. Defined as groups size=1 for					
	solitary.					
	Social species can benefit from collective vigilance and learning					
	experienced that can reduce roadkill rates.					

#### 629 Figures and their legends



630

631 Figure 1. Relative importance of predictors associated with empirical road mortality based on

632 random forest regression models for birds (a) and mammals (b). Boxplots show results for the

633 15 imputed datasets for each taxonomic group. Study predictors (representing study site

634 coordinates and survey sampling frequency) are in bold, taxonomic predictor (order) is in

635 italics, traits with <50% empirical data are marked with an asterisk.



637

638 Figure 2. Partial dependence plots for all tested predictors on the predicted road mortality

639 rates of Brazilian birds. Predictors are in descending variable importance order (see Fig. 1a).

640 Showing results for the 15 imputed datasets. Note that y-scales differ among plots.

641 Taxonomic orders are abbreviated as follows: Acc (Accipitriformes), Ans (Anseriformes), Ccl

642 (Cuculiformes), Chr (Charadriiformes), Clm (Columbiformes), Cpr (Caprimulgiformes), Crc

643 (Coraciiformes), Crm (Cariamiformes), Cth (Cathartiformes), Flc (Falconiformes), Gll

644 (Galliformes), Grf (Gruiformes), Pcf (Piciformes), Plc (Pelecaniformes), Pss (Passeriformes),

645 Pst (Psittaciformes), Slf (Suliformes), Strg (Strigiformes), Strt (Struthioniformes).



648

- Figure 3. Partial dependence plots for all tested predictors on the predicted road mortality 649
- 650 rates of Brazilian mammals. Predictors are in descending variable importance order (see Fig.
- 651 1b). Showing results for the 15 imputed datasets. Note that y-scales differ among plots.
- Taxonomic orders are abbreviated as follows: Crn (Carnivora), Ctr (Cetartiodactyla), Chr 652
- (Chiroptera), Cng (Cingulata), Ddl (Didelphimorphia), Lgm (Lagomorpha), Prs 653
- (Perissodactyla), Pls (Pilosa), Prm (Primates), Rdn (Rodentia). 654
- 655





657 Figure 4. Predicted standardized road mortality rates (in ind/km/year) for birds (a) and

658 mammals (b) in Brazil. Total road mortality rates based on existing paved roads (ind/year) for

659 birds (c) and mammals (d). Total bird (e) and mammalian (f) species richness.

- 661 Supplementary information for González-Suárez, M; Zanchetta Ferreira, F; Grilo, C.
- 662 Spatial and species-level predictions of road mortality risk using trait data. Global Ecology
- 663 and Biogeography
- 664 [Appendices S1-S3]
- 665

#### 666 Appendix S1. Correction of mortality rates

667

We used a correction factor derived from estimates of carcass persistence described in Santos, Carvalho, & Mira (2011). They estimated persistence probability (*S*), which we converted into a correction factor = 1+(1-S) adapting their estimates for survey intervals of 1, 2 and 7 days to our observed ranges, and combining some species groups to match our data (Table S1.1). Estimates from studies with survey intervals <1 day (ranging from twice a day to 16 times a day) were not corrected. Body mass was given priority when defining groups (e.g. rates for a bird of prey of 130 g were corrected based on the "Small birds" factor).

Groups	Survey intervals				
	1.0-1.4	1.5-3.0	3.0-7.0		
Small birds (4-200 g)	1.634	1.797	1.968		
Large birds (200-23000 g, excluding birds of prey)	1.283	1.391	1.717		
Birds of prey (175-1600 g)	1.255	1.327	1.555		
Small mammals (29-300 g)	1.611	1.759	1.970		
Large mammals (1100-170000 g)	1.196	1.294	1.457		
Bats (20-60 g)	1.854	1.963	2.000		

#### 678 Appendix S2. Results based on uncorrected road mortality rates





Figure S2.1. Relative importance of predictors associated with empirical uncorrected road 681 682 mortality based on random forest regression models for bird (a) and mammal (b). Boxplots 683 show results for the 15 imputed datasets for each taxonomic group. Study predictors (representing study site coordinates and survey sampling frequency) are in bold, taxonomic 684 685 predictor (order) is in italics, traits with <50% empirical data are marked with an asterisk. 686



688 Figure S2.2. Predicted standardized road mortality rates (in ind/km/year) for birds (a) and 689 mammals (b) in Brazil. Total road mortality rates based on existing paved roads (ind/year) for

690 birds (c) and mammals (d). Values based on models fitted for uncorrected road-kill rates.





693 Figure S2.3. Partial dependence plots for all tested predictors on the predicted uncorrected

694 road mortality rates of Brazilian birds. Rates are in ind/km/year. Predictors are in descending

695 variable importance order (Fig. S2.1). Showing results for the 15 imputed datasets. Note that

696 y-scales differ among plots. Taxonomic orders are abbreviated as follows: Acc

697 (Accipitriformes), Ans (Anseriformes), Ccl (Cuculiformes), Chr (Charadriiformes), Clm

698 (Columbiformes), Cpr (Caprimulgiformes), Crc (Coraciiformes), Crm (Cariamiformes), Cth

- 699 (Cathartiformes), Flc (Falconiformes), Gll (Galliformes), Grf (Gruiformes), Pcf (Piciformes),
- 700 Plc (Pelecaniformes), Pss (Passeriformes), Pst (Psittaciformes), Slf (Suliformes), Strg
- 701 (Strigiformes), Strt (Struthioniformes).
- 702



Figure S2.4. Partial dependence plots for all tested predictors on the predicted uncorrected

road mortality rates of Brazilian mammals. Predictors are in descending variable importance

706 order (Fig. S2.1). Showing results for the 15 imputed datasets. Note that y-scales differ among

707 plots. Taxonomic orders are abbreviated as follows: Crn (Carnivora), Ctr (Cetartiodactyla),

708 Chr (Chiroptera), Cng (Cingulata), Ddl (Didelphimorphia), Lgm (Lagomorpha), Prs

709 (Perissodactyla), Pls (Pilosa), Prm (Primates), Rdn (Rodentia).

710



713 Figure S2.5. Predicted and observed uncorrected road mortality rates for 170 species of birds

and 74 mammals. Symbol is the median value from all surveys for empirical data and across

715 all predicted locations –grid cells, for predicted rates. Error bars represent the minimum and

716 maximum range in observed and predicted values. Diagonal line indicates the 1:1

717 relationship.

#### 718 Appendix 3. Additional results with corrected road mortality rates

719



### 720

Figure S3.1. Location of the surveyed roads in Brazil represented in the 41 studies with road

mortality data for birds and mammals (note that some studies provided mortality rates for

both groups). The size of the symbols represent the median road mortality rates for each

taxonomic group. In birds the values range from 0.001 to 1.68 ind/km/year, while in

mammals values range from 0.003 to 18.5 ind/km/year.



Figure S3.2. Predicted standardized road mortality rates (in ind/km/year) in Brazil assuming
the minimum observed survey interval [(a) birds, (b) mammals, both 0.0417 days between

- surveys], mean observed interval [(c) birds 3.40 days between surveys, (d) mammals 3.64
- days between surveys], and maximum observed interval [(e) birds, (f) mammals, both 7 days
- 731 between surveys, which was the maximum interval we considered for our study).
- 732



Figure S3.3. Predicted and observed road mortality rates for 170 species of birds and 74 mammals. Symbol is the median value from all surveys for empirical data and across all predicted locations –grid cells, for predicted rates. Error bars represent the minimum and maximum range in observed and predicted values. Diagonal line indicates the 1:1

738 relationship.

733

739



740

Figure S3.4. Median predicted (grey boxplot) and observed (red symbols) road mortality rates

742 for birds (left panel) and mammals (right panel) classified in the different IUCN Red List

743 status categories.



744

Figure S3.5. Predicted standardized road mortality rates (in ind/km/year) for birds [(a) median

rates, (c) standard deviation among species within each cell] and mammals [(a) median rates,

747 (c) standard deviation among species within each cell] in Brazil.



- Figure 3.6. Road density map for Brazil reflecting current paved roads (panel a) and current
- and planned paved roads (b). Road data from <u>https://www.openstreetmap.org</u>.
- 751



- Figure S3.7. Predicted total road mortality rates (in ind/year) for birds (a) and mammals (b)
- 754 given the existing and planned paved road network in Brazil. Road data from
- 755 <u>https://www.openstreetmap.org</u>.
- 756
- 757
- 758





Figure S3.8. Predicted standardized road mortality rates (in ind/km/year) for threatened birds
(a) and threatened mammals (b) in Brazil. Total road mortality rates based on existing paved
roads (ind/year) for threatened birds (c) and threatened mammals (d). Total threatened bird (e)
and threatened mammalian (f) species richness.